

1 **Stocking strategies for coldwater fish populations under temperature stress**

2 HARALD FICKER^{1,3}, RUPERT MAZZUCCO³, HUBERT GASSNER², JOSEF WANZENBÖCK¹ AND ULF
3 DIECKMANN³

4 **Research Institute for Limnology of the University of Innsbruck, Mondseestraße 9, 5310 Mondsee, Austria*

5 *‡Institute for Freshwater Ecology, Fisheries Biology and Lake Research, Federal Agency for Water Manage-
6 ment, Scharfling 18, 5310 Mondsee, Austria*

7 *‡Evolution and Ecology Program, International Institute for Applied Systems Analysis, Schlossplatz 1, 2361
8 Laxenburg, Austria*

9

10 Corresponding Author: Harald Ficker
11 *University of Innsbruck*
12 *Research Institute for Limnology*
13 *Mondseestraße 9*
14 *A-5310 Mondsee, Austria*
15 *Phone: +43 650 5800 187*
16 *e-mail: harald.ficker@hotmail.com*

17

18 2nd Author: Rupert Mazzucco, e-mail: mazzucco@iiasa.ac.at

19 3rd Author: Hubert Gassner, e-mail: hubert.gassner@baw.at

20 4th Author: Josef Wanzenböck, e-mail: josef.wanzenboeck@uibk.ac.at

21 5th Author: Ulf Dieckmann, e-mail: dieckmann@iiasa.ac.at

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29 **Summary**

- 30 1. Coldwater fish stocks are increasingly affected by steadily increasing water temperatures.
31 The question arises whether stock management can be adapted to mitigate the conse-
32 quences of this climatic change.
- 33 2. Here, we estimate the effects of increasing water temperatures and different stocking strat-
34 egies on fisheries yield by recreational anglers. Using a process-based population model
35 based on an empirical long-term data set for the whitefish population (*Coregonus lavaretus*
36 (L.) species complex) of Lake Irrsee, Austria, we project density-dependent and tempera-
37 ture-dependent population growth and compare established stock enhancement strategies
38 to alternative stocking strategies under the aspect of increasing habitat temperatures and
39 cost neutrality. Additionally, we contrast the results obtained from the process-based model
40 to the results from simple regression models and argue that the latter show qualitative in-
41 adequacies in projecting catch with rising temperatures.
- 42 3. Our results indicate that increasing habitat temperatures reduce population biomass and
43 catch by the fishery through their effect on growth and survival.
- 44 4. Regarding stocking strategies, we find that stocking mostly small fish produces higher pop-
45 ulation biomass than stocking mostly larger fish, while catch remains almost constant.
- 46 5. *Synthesis and applications.* Stocking larger fish is more beneficial for the angling fishery
47 under the aspect of increasing temperatures. Adaption to climate change by changing stock-
48 ing strategies cannot, however, prevent an overall reduction in catch and population size of
49 coldwater fish.

50

51 *Keywords:* Alpine lake, angling, density dependence, growth probability, matrix model, natural
52 mortality, temperature dependence

53 **Introduction**

54 Compared to lakes in lowland areas, lakes in Alpine areas are typically characterized by great
55 depth and low water temperatures (Dokulil et al. 2010). Mean temperatures of surface and deep-
56 water layers in Alpine lakes of Central Europe have, however, increased between 0.5°C and
57 1°C over the last 40 years and further warming is expected because of ongoing climatic changes
58 (Dokulil et al. 2006; IPCC 2007; Dokulil et al. 2010; Dokulil 2014). This change in the thermal
59 regime is very likely to affect population dynamics of fish species that are living in Alpine lake
60 ecosystems, and consequently also the related fishery could be affected (Ficke, Myrick & Han-
61 sen 2007; Jeppesen et al. 2012).

62 Whitefish (Salmoniformes: *Coregonus* spp.) are typical coldwater fish that grow optimally
63 at low water temperatures (Casselman et al. 2002; Siikavuopio et al. 2013). They are very im-
64 portant for freshwater fisheries in northern temperate regions (Berka 1990; Petr 1999; Ebener
65 et al. 2008; Jeppesen et al. 2012). The planktivorous European whitefish (*Coregonus lavaretus*
66 (L. 1758) species complex) lives in the cold-water layers of Alpine lakes and was exploited
67 mainly by commercial fisheries before the 1970s. With improving angling techniques over the
68 last decades, whitefish has become also very important for recreational fisheries.

69 To compensate for harvesting by fisheries, managers of exploited whitefish populations com-
70 monly conduct stocking programs. In general, stocking strategies comprise introductions of
71 small (e.g., larvae) and large (e.g., one-summer-old) fish in various proportions. Stocking small
72 fish is common, although many authors argue that stocking larger fish is more profitable for
73 whitefish fisheries compared to stocking smaller fish (Salojärvi & Huusko 1990; Wanzenböck
74 & Jagsch 1998; Lasenby, Kerr & Hooper 2001; Gerdeaux 2004).

75 Stocking strategies are almost never systematically evaluated in small fisheries (Arlinghaus,
76 Mehner & Cowx 2002; Cowx & Gerdeaux 2004). Fisheries managers often do not pay enough
77 attention to the cost-effectiveness of the applied stocking program and to possible negative im-
78 pacts of stocking due to, e.g., density-dependent effects on growth and mortality (Salojärvi

79 1991; Arlinghaus, Mehner & Cowx 2002). Moreover, in the context of climate change, the
80 question arises how stocking strategies can be adapted to ensure sustainable fisheries manage-
81 ment of coldwater fish under increasing habitat temperatures.

82 In general, fish are poikilothermic animals and live in specific temperature ranges, preferring
83 water temperatures that promote optimal growth (Jobling 1981; Ohlberger et al. 2008; Mehner
84 et al. 2010). Growth in turn is related to natural mortality (Pauly 1980; Jensen 1996; Lorenzen
85 1996). Fishery yield depends on how well the fish grow and survive. Therefore, a correlation
86 between water temperatures and catches often exist (Sutcliffe Jr. Drinkwater & Muir 1977;
87 Scarnecchia 1984; Sakuramoto, Hasegawa & Suzuki 2005; Biswas et al. 2009).

88 Mathematical models are very helpful to estimate how increasing temperatures and various
89 stocking strategies will affect population dynamics and the related catch by the fishery. Simple
90 regression models, fitted to observed water temperatures and catches, can be used to extrapolate
91 catches under higher temperatures. This model approach, however, does not account for the
92 relevant life-history processes and the resulting population dynamics.

93 In contrast, a process-based model approach provides additional opportunities for analyzing
94 population dynamics and can readily be extended to account for relevant mechanisms, such as
95 fishing, stocking, and density dependence. Models based on life-history processes are differen-
96 tial equations, matrix models (MMs), and individual-based models (IBMs).

97 Differential equations can be analytically solved for unstructured populations, while only
98 numerical solutions are feasible (and effectively become matrix models) for structured popula-
99 tions. In contrast, IBMs provide great flexibility and detailed insights into population dynamics,
100 primarily because they explicitly account for individual variation (Grimm 1999; DeAngelis &
101 Mooij 2005). Although IBMs and MMs often produce similar results, particularly when the
102 MMs account for aspects of variation, IBMs require substantially higher computational effort
103 (Pfister & Stevens 2003; Sable & Rose 2008). Therefore, matrix models provide a good com-
104 promise and allow studying structured populations with reasonable computational effort.

105 Conventional matrix models used for studying fish populations, also known as Leslie matrix
106 models (Leslie 1945; Caswell 2001), consider only age classes. Although age is a natural de-
107 mographic property in whitefish life history, vital parameters and management interventions
108 often depend on body size (Lorenzen & Enberg 2002; DeRoos, Persson & Cauley 2003; Lewin,
109 Arlinghaus & Mehner 2006; Ficker et al. 2014). A length-based model may therefore be more
110 suitable for whitefish populations.

111 Here, we use a length-structured matrix model with temperature dependence and density de-
112 pendence in growth and mortality to evaluate the effects of increasing habitat temperatures on
113 the total biomass and catch by recreational anglers of a European whitefish population. A long-
114 term (10 years) dataset of experimental gillnet catches was used to derive model parameters for
115 the whitefish population of Lake Irrsee (Gassner, Hassan & Wanzenböck 2004; Gassner &
116 Wanzenböck 2007). We further compare our modeling results to projections by simple regres-
117 sion models describing the correlation between catch and habitat temperature. We additionally
118 assess the cost-effectiveness of the applied stocking strategy on the Lake Irrsee population and
119 compare it to various other strategies with consideration of the fraction of invested money on
120 small (i.e., 1 cm total length) and large (i.e., 10 cm total length) fish under constant and under
121 continuously increasing temperature scenarios. Finally, we offer policy recommendations for
122 stocking strategies of European whitefish under the aspect of climate change.

123

124 **Material and Methods**

125 We develop a process-based model to project the whitefish population of Lake Irrsee under
126 different stocking and temperature scenarios. The resulting length-structured matrix model aug-
127 mented with stochastic elements includes all relevant processes for population dynamics of
128 whitefish, which are: temperature-dependent and density-dependent growth, survival, and re-
129 production.

130 Stocking strategies and catch by anglers are incorporated into the model through vectors of
131 stocked and caught whitefish, respectively. Assuming different temperature scenarios, we pro-
132 ject annual biomass and catches over a period of 50 years with different stocking strategies.
133 Below, we briefly discuss selected points specifically. Details can be found in the supplemen-
134 tary material.

135

136 *Sampling data*

137 The pre-alpine Lake Irrsee, Austria (N 47° 53', E 13° 18') is classified as an oligo-mesotrophic
138 lake with a holomictic-dimictic mixing regime. Its maximum depth is 32 m and its surface area
139 stretches over 3.6 km². European whitefish is the dominant fish species in Lake Irrsee and im-
140 portant for the local recreational fishery.

141 Since the year 2000, the whitefish population of Lake Irrsee is studied by means of gillnetting
142 carried out annually in October (pre-spawning census; Gassner, Hassan & Wanzenböck 2004;
143 Gassner & Wanzenböck 2007). The overall catch amounted to 2,013 individual whitefish be-
144 tween years 2000 and 2009. Gillnet fleets with different randomized mesh sizes between
145 15 mm and 70 mm were assembled and set over night in part of the lake in 12 to 15m depth.

146 Individual length (± 0.5 cm), weight (± 5 g), age, sex and ripeness of gonads were deter-
147 mined for all caught whitefish. Age identification was achieved by scale reading according to
148 the method used by DeVries & Frie (1996) and Gassner, Hassan & Wanzenböck (2004).

149 The examination of sex and ripeness stages according to Nikolsky (in: Ricker 1970) was
150 done after dissection by classifying individuals into male, female, or juvenile and as spawners
151 or non-spawners. Fresh eggs of mature female individuals were counted per unit weight in the
152 year 2010 according to the gravimetric sub-sampling method described by Bagenal (1978).

153 Total fish biomass in Lake Irrsee was estimated through simultaneously performed hydro-
154 acoustic surveys in the open water area with two split-beam echo sounders in the year 2000
155 (Wanzenböck et al. 2003). The population biomass of European whitefish was assumed to ac-
156 count for 60% of the total observed biomass.

157 Temperatures and oxygen concentration were available from water samples collected in
158 0, 2, 5, 8, 10, 12, 15, 20, 25, and 30m depth at the deepest site of the lake on a monthly basis.
159 Temperatures were measured in the field with a mercury thermometer and oxygen concentra-
160 tions were determined in the laboratory according to the Winkler procedure (Winkler 1889).
161 Annual mean growth temperatures for European whitefish during the growth period from May
162 to October were derived from temperature measurements in the suitable oxythermal habitat for
163 coldwater fish (i.e., $O_2 > 3\text{mg l}^{-1}$ and $T < 21.2\text{ }^\circ\text{C}$; Stefan et al. 1995)

164

165 *Spawning, eggs, and larvae*

166 European whitefish reproduce in early winter and spawned eggs develop over the winter months
167 till larvae hatch in spring (Fuller, Scott & Fraser 1976; Wahl & Löffler 2009). We calculated
168 the biomass of female spawners using the observed sex ratio, a sigmoid maturity function
169 (Ficker et al. 2014), and an allometric length–weight relationship. The average fecundity, that
170 is, the average number of eggs per unit weight female fish, is estimated from our data and
171 modeled as a stochastic variable. Finally, the number of hatching larvae, and thus the success
172 of natural reproduction, is obtained from the effective fecundity, which is defined as the number
173 of produced offspring that survives till hatching from the egg.

174 Survival is usually much lower for early development stages compared to larger fish, like in
175 eggs and freshly hatched larvae (Salojärvi 1982; Fuiman & Werner 2002). We assume egg
176 mortality over the developmental period and larval mortality over the first four weeks of life to
177 be much higher compared to mortality rates of larger whitefish (see supplementary material).

178

179 *Density-dependent and temperature-dependent growth*

180 Growth of a fish is depends primarily on size and is also affected by population density and
181 environmental temperature. Small fish grow almost linearly and large fish grow according to a
182 von Bertalanffy model toward an asymptotic length (Quince et al. 2008). The asymptotic length
183 depends on total biomass and therefore on population density via a Maynard Smith–Slatkin-
184 type functional response (Smith & Slatkin 1973; Beverton & Holt 1993; Lorenzen & Enberg
185 2002; Ylikarjula et al. 2002), while the von Bertalanffy growth coefficient depends on environ-
186 mental temperature (Ricker 1979; Fontoura & Agostinho 1996; Jensen 1996; see supplementary
187 material for details). Asymptotic length and growth coefficient are related (Pauly 1980; Jensen
188 1996), which makes the asymptotic length also indirectly dependent on temperature. We as-
189 sume a lognormal distribution of monthly growth increments and allow growth to vary among
190 individuals of the same length.

191

192 *Natural and fishing mortality*

193 Natural mortality of a fish is related to growth and environmental temperature (Pauly 1980;
194 Quinn & Deriso 1999; Kenchington 2013) and therefore indirectly depends on population den-
195 sity. We estimated natural mortality through two different methods (Pauly 1980; Jensen 1996;
196 see supplementary material) from density-dependent and temperature-dependent growth pa-
197 rameters. Additionally, we consider fishing mortality. Fisheries impose certain size limits
198 which leads to selective removal of fish of certain lengths. We model this size-selective removal
199 as a stochastic process. We assume a constant angling effort per unit time, which implies that

200 the total catch is limited, and that total catch drops faster than linearly as abundance in the
201 catchable size range decreases towards 0. We used catch statistics of the local angler association
202 for parameterization of stochastic fish removal by anglers.

203

204 *Stocking strategies*

205 Currently, fisheries stock small whitefish (around 630,000 individuals of ~1 cm length with
206 an individual price of € 0.014) in March and larger whitefish (around 6,000 individuals of
207 ~10 cm length with an individual price of € 0.30) in September. This means that about 83%
208 of the money invested into stocking is used for stocking small fish and the remainder for stock-
209 ing large fish. To compare the cost-effectiveness, we investigate stocking strategies that allocate
210 the same total amount of money in different ratio (thus, a stocking ration of 0.1 means 10% of
211 the money is invested into stocking small fish etc.).

212

213 *Temperature scenarios*

214 We consider three different temperature scenarios (i.e., constant temperature, +1°C, and +2°C
215 over 50 years) The two scenarios with increasing temperatures are based on the observed tem-
216 perature increase in surface waters of Lake Irrsee over the last decades (i.e., annual average
217 with +0.9°C and average of spring and summer temperatures with +1.9°C; Dokulil et al. 2010)
218 and we also consider deep water warming and projected future temperature development of
219 Austrian lakes described in Dokulil et al. (2006) and Dokulil (2014).

220

221

222 **Results**

223 We projected population biomass and anglers catch under changing annual habitat tempera-
224 tures, investigating three basic temperature scenarios. We compared the predictions from sim-
225 ple regression models to our process-based model; we investigated the effects of increasing
226 temperatures on biomass and catch; we analyzed the mechanism underlying the temperature
227 effect; and finally assessed stocking strategies comprising introductions of small and large
228 whitefish in different ratios.

229

230 *Process-based model vs. regression models*

231 Projections with the process-based model are shown for two different estimates of natural mor-
232 tality (Pauly 1980; Jensen 1996), both resulting in qualitatively very similar predictions. We
233 project annual catches (with a three year delay) as a function of growth temperature with our
234 process-based model and extrapolate catches with simple regression models fitted to observa-
235 tions. The quadratic regression model agrees with the process-based model in that both project
236 saturating catch at low growth temperatures. The exponential regression model agrees with the
237 process-based model in that both project decreasing catches with increasing growth tempera-
238 tures showing a non-linear pattern (although projected catches differ substantially). Quadratic
239 and linear regression models project a complete collapse in catches for a relatively modest in-
240 crease in growth temperatures similar to the collapse projected by the process-based model. In
241 contrast, the linear and the exponential regression model also project high catch without satu-
242 ration for low growth temperatures. No regression model shows qualitative agreement with the
243 process-based model over the whole range of growth temperatures considered (Fig. 1).

244

245 *Temperature effects*

246 Using our process-based model we project changes in population biomass and catch by anglers
247 over a period of 50 years under three temperature scenarios (Fig. 2.a). We find that population

248 biomass and catch by anglers decrease with increasing temperatures. The effect is stronger
249 when the temperature increase is larger. Our projections with Jensen's estimate of natural mo-
250 rality show that increasing habitat temperature reduce biomass by about 2.6% (i.e.,
251 -0.9 kg ha^{-1}) and by about 4.4% (i.e., -1.6 kg ha^{-1}), respectively (Fig. 2.b), while catch de-
252 creases by about 24% (i.e., -1.2 kg ha^{-1}) and 45% (i.e., -2.3 kg ha^{-1}), respectively (Fig.
253 2.c). Our projections with Pauly's estimate show that increasing habitat temperatures reduce
254 biomass by about 4.3% (i.e., -1.7 kg ha^{-1}) and by about 7.9% (i.e., -3.1 kg ha^{-1}), respec-
255 tively, and that catch decreases by about 26% (i.e., -1.4 kg ha^{-1}) and 48% (i.e.,
256 -2.6 kg ha^{-1}), respectively (not shown).

257

258 *Underlying mechanism*

259 Temperature has direct and indirect effects in our process-based model. The growth coefficient
260 depends directly on temperature (Fig. 3.a) via a simple relation (see material and methods sec-
261 tion and supplementary material). Since population dynamics in the model depends on growth,
262 also the density-dependent parameters asymptotic length and survival probability are indirectly
263 dependent on temperature. Increasing temperature increases the growth coefficient (Fig. 3.a)
264 and decreases asymptotic length (Fig. 3.b) and annual survival (Fig. 3.c). Our projections show
265 that increasing habitat temperature increase the growth coefficient by about 6.7% (i.e.,
266 $+0.02 \text{ y}^{-1}$) and 12.4% (i.e., $+0.02 \text{ y}^{-1}$), respectively, while asymptotic length decreases by
267 about 2.9% (i.e., -1.3 cm) and 5.2% (i.e., -2.3 cm), respectively, and natural annual survival
268 decreases by about 3.7% (i.e., -0.02%) and 6.7% (i.e., -0.04%), respectively. Our projections
269 using Pauly's estimate show that increasing habitat temperature increase the growth coefficient
270 by about 6.7% (i.e., $+0.02 \text{ y}^{-1}$) and 12.4% (i.e., $+0.05 \text{ y}^{-1}$), respectively, while asymptotic
271 length decreases by about 2.7% (i.e., -1.2 cm) and 4.8% (i.e., -2.1 cm), respectively, and
272 natural annual survival decreases by about 4.6% (i.e., -0.03%) and 8.6% (i.e., -0.05%), re-
273 spectively.

274

275 *Stocking strategies*

276 Stocking strategies, in our case, are expressed by the ratio of money invested into stocking small
277 fish to the total amount of money invested for stocking. This includes the extreme cases where
278 the money is invested either only into stocking small fish (corresponding to a stocking ration
279 of 1) or only into stocking large fish (corresponding to a stocking ratio of 0). To assess the cost-
280 effectiveness of stocking strategies for constant temperatures, we project population biomass
281 and catch by anglers for different stocking ratios with a fixed investment budget. Different
282 stocking ratios result in very different numbers of introduced fish, because large fish are sub-
283 stantially more expensive than small fish (e.g., in Lake Irrsee 10 cm fish cost 21.4 times more
284 than 1 cm fish). Our projections reveal that increasing the current stocking ratio of 0.83 in-
285 creases population biomass after 10 years, and decreasing the current stocking ratio decreases
286 biomass, while the catch remains nearly the same with a very inconspicuous peak at a stocking
287 ratio of about 0.6 (Fig. 4).

288

289 *Mitigation of climate change*

290 To evaluate how stocking strategies can be adapted to mitigate the effects of climate change,
291 we project population biomass and catch by anglers over a period of 10 and 25 years for in-
292 creasing habitat temperatures (+2°C over 50 years; Scenario 3 in Fig. 2 and 3) and different
293 stocking ratios. Compared to the projection with constant temperature (Fig. 4), population bio-
294 mass and catch by anglers is generally lower. The catch, however, is now clearly maximized at
295 lower stocking ratios of about 0.3 (Fig. 5).

296 **Discussion**

297 Whitefish stocks in cold Alpine lake ecosystems are affected through increasing temperatures
298 due to climatic changes. Fisheries management of coldwater fishes commonly uses stocking to
299 maintain available catches for recreational and commercial fisheries. To evaluate the often un-
300 known effects of stocking on population dynamics as well on the fishery itself, we have devel-
301 oped a process-based model of density-dependent and temperature-dependent population
302 growth. Density dependence has been introduced in the growth parameter asymptotic length:
303 higher population densities reduce asymptotic length (Jensen 1997). Additionally, the effect of
304 temperature has been integrated into the growth coefficient: higher temperatures lead to higher
305 growth coefficients (depending on the temperature optimum for coldwater fish; Jobling 1981;
306 Stefan et al. 1995; Casselman 2002).

307 Natural mortality of whitefish has been derived from growth parameters and temperatures
308 through two different methods (Pauly 1980; Jensen 1996). Both are considered to produce use-
309 ful estimates when the growth coefficient can be derived accurately from population data and
310 when adult life span is not exceptionally long (Kenchington 2013). We found that the simpler
311 method proposed by Jensen (Jensen 1996), generally leads to higher estimates of natural mor-
312 tality than the regression based model of Pauly (1980). Still, both methods produce qualitatively
313 and quantitatively similar results in our model projections.

314 The parameterization of the process-based model is based on an empirical long-term data set
315 of Lake Irrsee collected by annual gillnet samples and catch statistics. We have estimated initial
316 biomass, growth parameters, fecundity, maturity and sex ratio directly from the data. Because
317 of the importance of predation mortality in early life stages, we have modeled early life-stage
318 mortality separately as a density-independent process. Nevertheless, reproduction is tempera-
319 ture- and density-dependent because of the relationship between adult size and reproduction
320 efficiency (i.e., size-dependent maturation and size-dependent egg production). The optimal
321 temperature range for whitefish growth, as well as egg and larval mortality, which were not

322 available from field sampling, have been taken from literature. The sensitivity of our model to
323 egg and larval mortality is high, which is in accordance to theoretical expectations that early
324 life stages have a strong influence on population growth and consequently on recruitment to the
325 fishery (Ricker 1975; Chambers & Trippel 1997; Fuiman & Werner 2002).

326 The assumed optimal growth temperature range (i.e., $T_{\min} = 2^{\circ}\text{C}$, $T_{\max} = 22^{\circ}\text{C}$) had also a
327 great effect on the quantity of projected catches, whereas the decreasing trend with increasing
328 temperature was robust. The minimal temperature for growth that we used in our model was
329 very precisely evaluated by Siikavuopio et al. (2010) who showed that whitefish grows at 3°C
330 but not at 1°C water temperature. In contrast, the maximum temperature for growth is charac-
331 terized only vaguely in literature and ranges from 13.5°C to 22°C (Jobling 1981; EIFAC 1994;
332 Casselman 2002; Siikavuopio, Knudsen & Amundsen 2010; Szczepkowski, Szczepkowska &
333 Krzywosz 2006) and it is also very likely that the temperature-dependence in growth is species-
334 specific as proposed by Ohlberger et al (2012). Consequently, the temperature at which a col-
335 lapse of an actual fishery occurs may be different from the 13°C at which it was observed in
336 our model projections. To refine the prediction, the maximum temperature for growth needs to
337 be assessed more accurately.

338 The strength of our model is the consideration of important life-history processes with respect
339 to body size. Although simple statistical models showed similar trends of catches under a
340 changing climate, the underlying mechanisms in population dynamics remain unclear, and con-
341 sequently a process-based model is advantageous.

342 Our results clearly demonstrate that lower catches must be expected in cold-water fisheries
343 with continuously increasing temperatures in the future. Additionally, the process-based model
344 reveals that lower catches are mainly due to accelerated growth of juveniles resulting in smaller
345 sizes of adults and consequently lower recruitment into the established size-limit of the recrea-
346 tional fishery. We further found that population biomass decreases as a consequence of higher
347 natural mortality. Modeling results for different stocking strategies indicate that this trend could

348 be partly mitigated through stocking higher ratios of small fish. While changing stocking strat-
349 egies cannot prevent a reduction in catch with increasing temperatures, stocking larger white-
350 fish nevertheless seem to be more advantageous for the recreational angling fishery, insofar as
351 it maximizes catch under the circumstances and thus angler satisfaction.

352

353 **Acknowledgements**

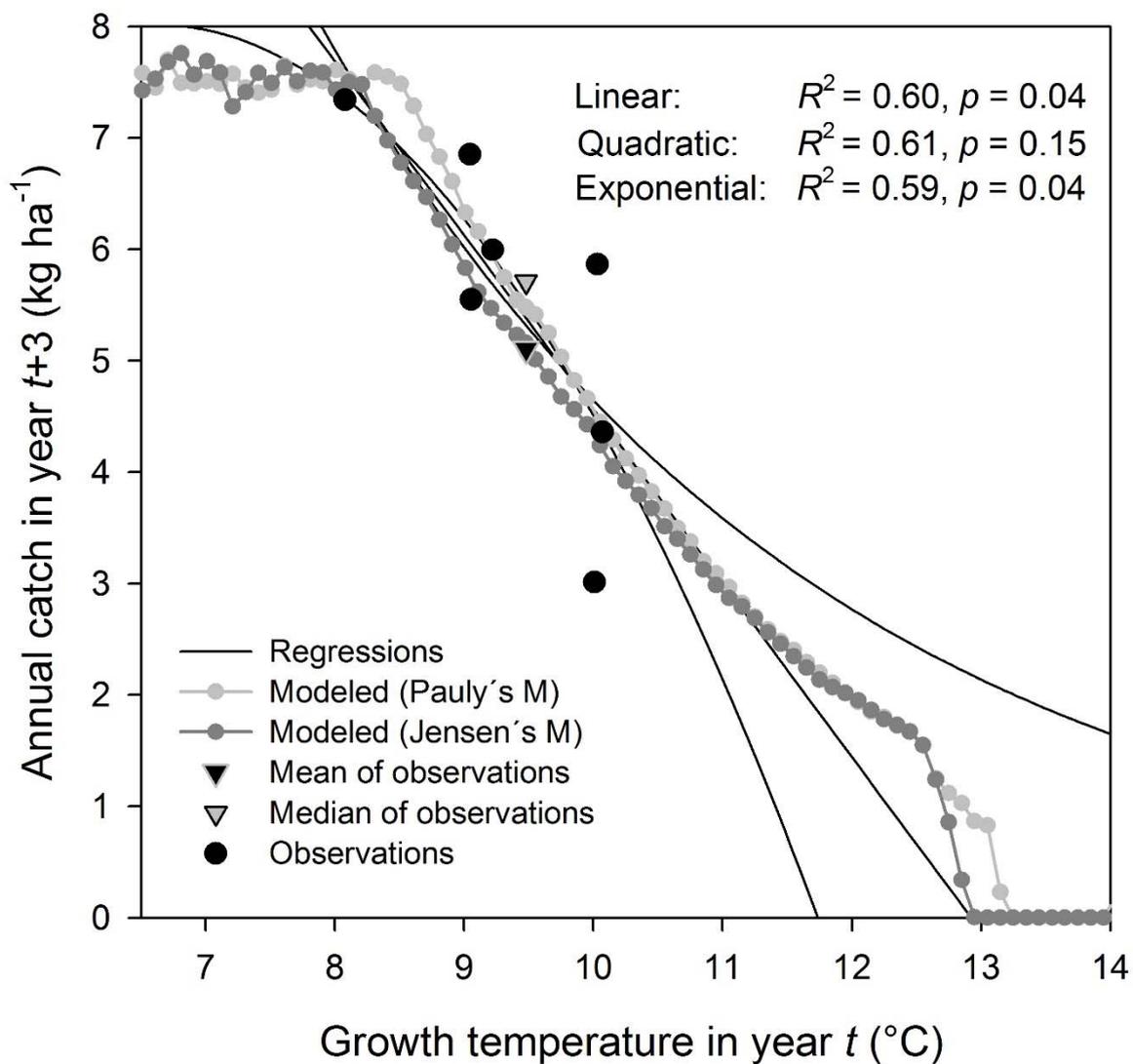
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360

361 **Figures**

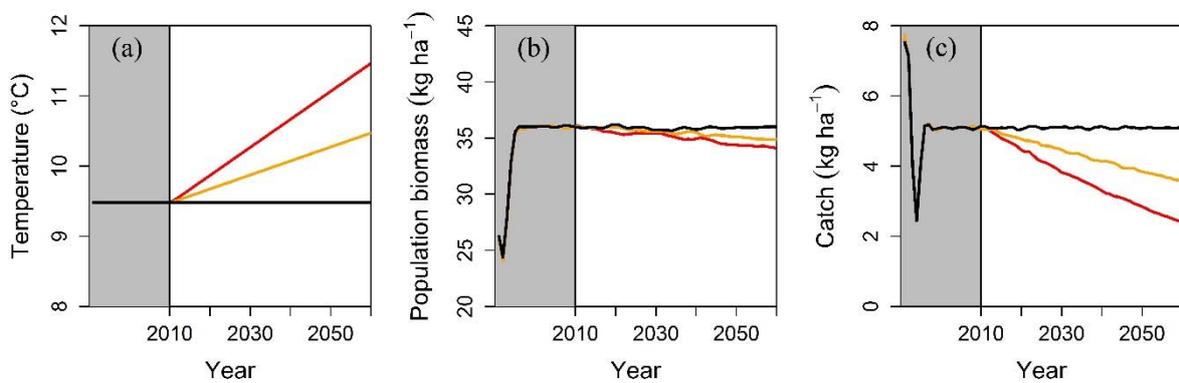
362 Figure 1:

363 Catch predictions of our process-based model compared to simple regression models. Black
364 solid lines show predictions of three regression models (linear, quadratic, and exponential) fit-
365 ted to observational data of growth temperature and anglers catch, with a time lag of three years
366 (black points; see text). Grey points and interpolation lines show predictions of our process-
367 based models using two different mortality estimation procedures. All models capture the de-
368 crease of anglers catch with increasing temperatures. They differ in whether they allow a satu-
369 ration of the catch towards low temperatures, and in whether they allow a collapse towards high
370 temperatures and in how this collapse is approached.



372 Figure 2:

373 Increasing growth temperatures decrease population biomass and catch. Projections for three
374 different temperature scenarios (a): constant temperature (black line), +1°C increase over 50
375 years (orange line) and +2°C increase over 50 years (red line). Population biomass of whitefish
376 decreases only slightly with increasing temperature (b), while catch by recreational angling
377 decreases substantially with increasing temperature (c). Grey shading indicates the initial sta-
378 bilization period (see text).



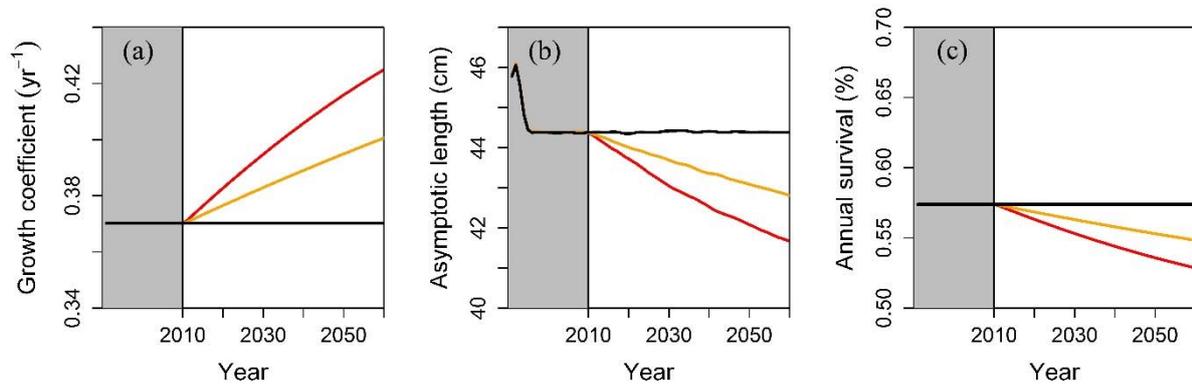
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380 Figure 3:

381 Higher temperatures affect growth and survival. Increasing temperatures (a) increase growth

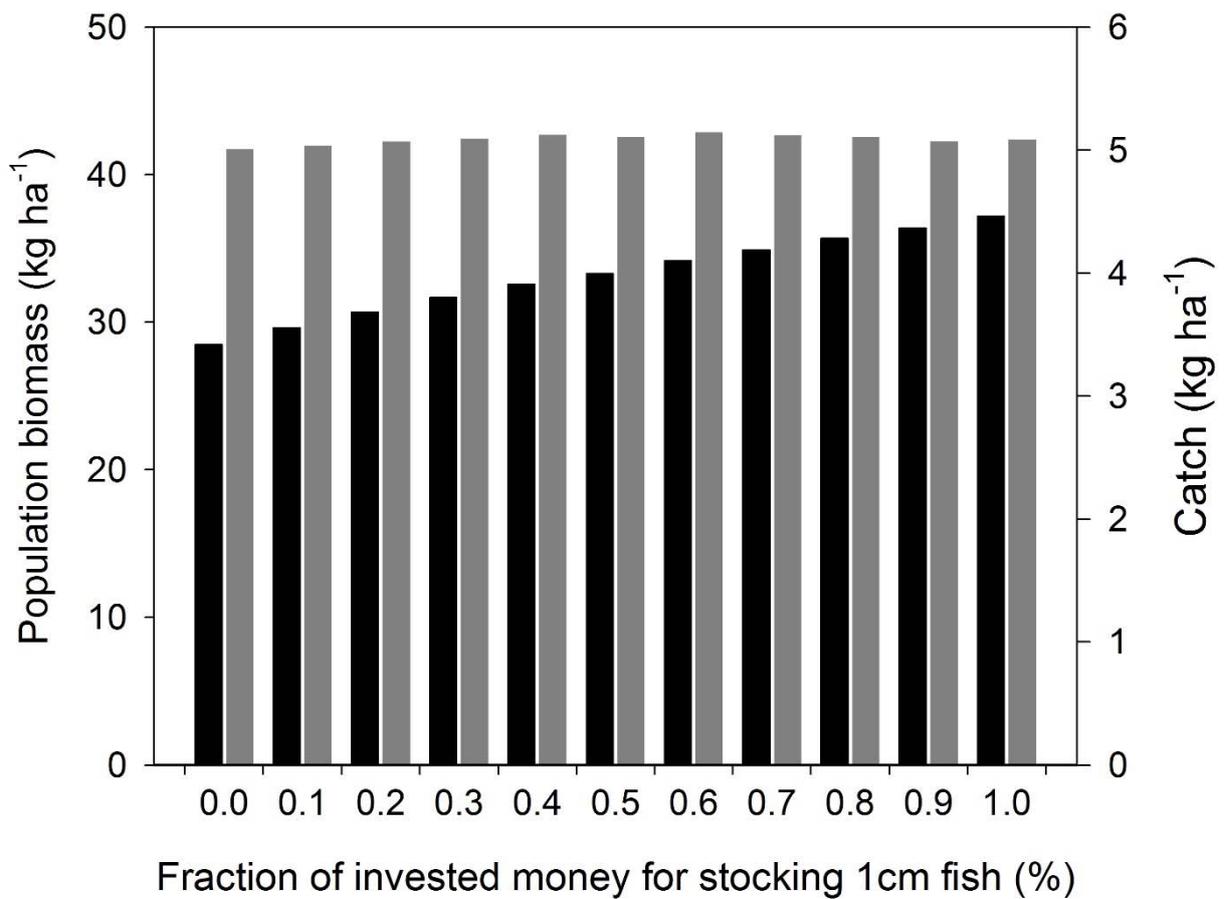
382 coefficients, (b) decrease asymptotic lengths and (c) consequently also reduce annual survival.

383 Colors as in Fig.2.



384

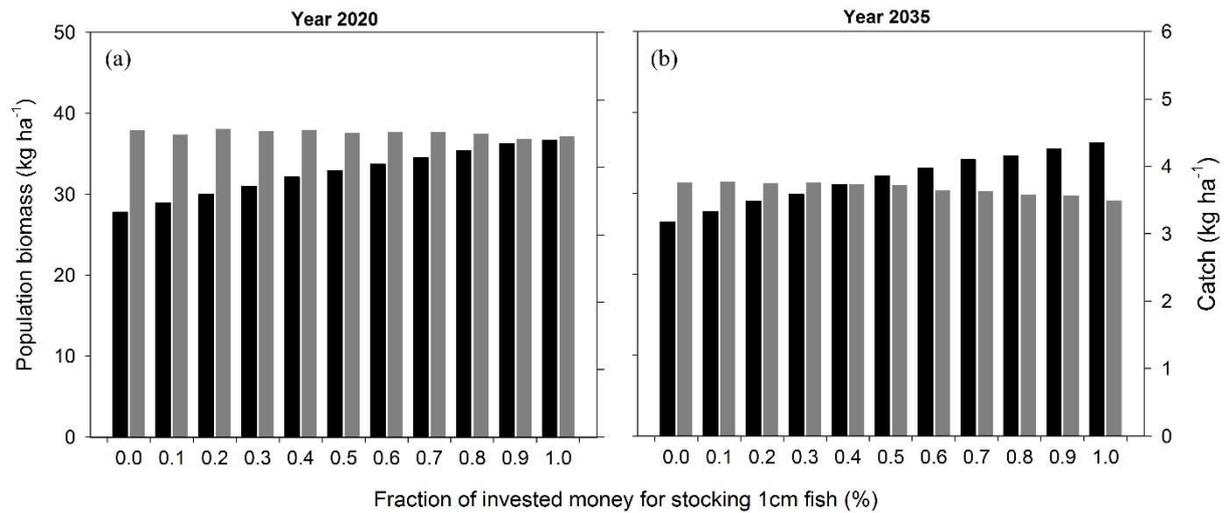
385 Figure 4:
386 Stocking ratio affects population biomass more strongly than catch. For constant temperatures,
387 solid bars show projected population biomass (black) and catch by anglers (grey) ten years after
388 changing the stocking ratio (i.e., fraction of money invested in small fish) from the current
389 stocking ratio in Lake Irrsee of 0.83.



390

391 Figure 5:

392 With increasing temperatures catch is maximized at lower stocking ratios. For increasing tem-
393 peratures (+2°C over 50 years; scenario 3 in figure 2 and 3), panels show projections of popu-
394 lation biomass and catch by anglers after (a) 10 years and (b) 25 years after changing the stock-
395 ing ratio from the current stocking ratio (see Fig.4).



396

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