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# Achieving carbon-neutral iron and steelmaking in Europe through the deployment of bioenergy with carbon capture and storage

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14	Abstract
15	The 30 integrated steel plants operating in the European Union (EU) are among
16	the largest single-point $CO_2$ emitters in the region. The deployment of bioenergy
17	with carbon capture and storage (bio-CCS) could significantly reduce their fossil-
18	based $CO_2$ emissions. In detail, the results demonstrate that $CO_2$ emission
19	reduction targets of up to 20% can be met entirely by biomass deployment. A
20	slow CCS technology introduction on top of biomass deployment is expected as
21	the requirement for emission reduction exceeds 20%. Bio-CCS could then be a
22	key technology, particularly in terms of meeting targets above 50%, of CO <sub>2</sub>
23	avoidance cost ranging between $\leq 60$ and $\leq 100 \text{ to}_{20}^{-1}$ at full-scale deployment. The
24	future of bio-CCS and its utilisation on a larger scale would therefore only be
25	viable if such CO <sub>2</sub> avoidance cost were to become economically appealing. Small

and medium plants in particular, would economically benefit from sharing CO<sub>2</sub>

pipeline networks. CO<sub>2</sub> transport, however, makes a relatively small contribution

to the total CO<sub>2</sub> avoidance cost. In the future, the role of bio-CCS in the European

iron and steelmaking industry will also be influenced by non-economic conditions, such as regulations, public acceptance, realistic  $CO_2$  storage capacity,

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#### 33 Keywords:

34 BECCS; bio-CCS; blast furnace; industry; charcoal; CCS

and the progress of other mitigation technologies.

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#### 38 Highlights:

- Bio-CCS can help iron and steel making become close to carbon neutral.
- 40 Average bio-CCS avoidance cost in the EU is about  $\textcircled{CO}_2^{-1}$ .
- Netherlands, France, and Belgium have the lowest bio-CCS deployment cost.

#### 42 **1 Introduction**

The European iron and steel industry annually generates over 200 million tons of carbon 43 dioxide (Mt<sub>CO<sub>2</sub></sub>) (Borkent and Beer, 2016), which amounts to 5% of all CO<sub>2</sub> emissions 44 45 produced across EU-28 countries in 2016 (Eurostat, 2016). The majority of these emissions 46 come from the 30 integrated steel plants that produce 60% of the European steel output 47 (World Steel Association, 2017). Their high emission intensity is due to the nature of the iron 48 and steel production process from iron ore, which in comparison to scrap recycling, is two and 49 half times more emission intensive (Beer et al., 2000). As the steel scrap recycling rate is not 50 sufficient to meet the increasing demand for steel, ore based steel production via a blast 51 furnace-basic oxygen furnace (BF-BOF) route is expected to remain dominant until at least 52 2050 (Pauliuk et al., 2013). Therefore, to achieve the EU emission reduction targets for 2020, 53 2030 and 2050 (European Commission, 2017), the 30 integrated plants will have to implement breakthrough technologies for CO<sub>2</sub> emission abatement (European Commission, 54 55 2013). A key technology that can contribute significantly to deep emission cuts is carbon 56 capture and storage (CCS) (European Commission, 2011a, 2011b; ZEP, 2013). A hybrid 57 approach that combines CCS with biomass (bio-CCS) could provide even further emission reductions in this industry (Arasto et al., 2014). The average 2017 price of European emission 58 allowances of  $\in 5.80 \text{ t}_{\text{CO}_2}^{-1}$  (Business Insider, 2018) and an absence of bio-CCS specific 59 60 incentives, make its application in Europe unrealistic for the moment (EUROFER, 2013). 61 However, the likely overshoot of the remaining CO<sub>2</sub> budget for limiting global warming to 62 below 2°C (UNEP, 2017), in combination with the hitherto slow transition to low-carbon iron 63 and steel making technologies, is increasing the need for the deployment of significant CO<sub>2</sub> 64 emission reduction measures like bio-CCS in Europe in the near future (Mintenig et al., 2017; 65 Scott and Geden, 2018).

Broadly speaking, the key role of negative emission technologies is to generate negative emissions that would compensate for  $CO_2$  emissions from sectors that may have a hard time reaching carbon-neutrality (such as agriculture, aviation or industry) (Erbach, 2015). Specifically, bio-CCS offers a way to generate energy that is carbon neutral/negative, which makes it suitable for co-application during energy conversion or with energy intensive

71 industrial processes. Scenarios for the decarbonisation of the iron and steel industry generally 72 involve CCS, either on its own (Pardo and Moya, 2013; Solano Rodriguez et al., 2017), or in 73 combination with a top gas recycling blast furnace process (EUROFER, 2013; Remus et al., 74 2013). Due to the technical role that fossil fuels play in the iron ore reduction process, only a 75 limited biomass substitution is feasible (Mousa et al., 2016; Suopajärvi et al., 2017). 76 Therefore, additional measures such as bio-CCS deployment would be needed to achieve high 77 levels of CO<sub>2</sub> reduction across an integrated steel plant. The introduction of bioenergy with CCS could theoretically achieve carbon-neutral steelmaking (considering that bioenergy can 78 79 substitute over 40% of fossil-based CO<sub>2</sub> emissions (Mandova et al., 2018) and that CCS can capture over 60% of the CO<sub>2</sub> emissions that occur on-site (IEAGHG, 2013)) without a 80 81 significant retrofit of a steel plant. However, this carbon-neutral iron and steelmaking 82 opportunity is currently being impeded by the challenges raised by any deployment of bio-83 CCS.

84 Deployment of bio-CCS has so far been stagnant, with only a few small demonstration-scale bio-CCS projects currently being operational (e.g., the Illinois Industrial CCS Project) (Global 85 86 CCS Institute, 2018). Any bio-CCS application within fully fossil fuel-based processes would 87 necessitate simultaneously overcoming barriers to both bioenergy and CCS implementation. 88 Issues related to the actual implementation and cost of CO<sub>2</sub> capture, transport and storage, 89 uncertainties in the long term response of the environment to CO<sub>2</sub> storage, and public 90 acceptance or ability to prolong reliance on fossil fuels, are the main arguments limiting CCS progress (Fuss et al., 2014). As of 2018, there are only 30  $Mt_{CO_2}$  stored annually worldwide 91 92 (Global CCS Institute, 2018). CCS deployment will therefore have a hard time reaching the 93 annual CO<sub>2</sub> storage volumes required by, for instance, the International Energy Agency (IEA) 94 2°C scenario of 400 Mt<sub>CO2</sub> by 2025 (IEA, 2014). Insufficient policy support to create a 95 business case for CCS, for example, in the EU Emission Trading System (ETS) (Purvis and 96 Vaghi, 2015), makes the required CCS expansion unrealistic over the next decade. On the 97 same note, sustainable biomass supply constraints, concerns associated with competition 98 between bioenergy and food production, the complexity of emission accounting, as well as 99 direct and indirect land use change, are major arguments against increased bioenergy use 100 (Sanchez and Kammen, 2016).

101 There is currently no commercialised application of bio-CCS in the iron and steel industry, 102 even though bioenergy and CCS indepently, are commercialised (e.g., charcoal utilisation in 103 Brazilian mini blast furnaces (Machado et al., 2010) and a CCS facility in Abu Dhabi with an

annual capture capacity of 0.8 Mt<sub>CO2</sub> (Global CCS Institute, 2018; IEA, 2014)). The 104 105 suitability of bio-CCS is highly dependent on geographic location, which diversifies 106 opportunities for large-scale bio-CCS application across steel plants. Factors such as 107 industrial plant structure, the availability of CO<sub>2</sub> storage and transport options, sufficient 108 sustainable biomass resources, supportive regulatory frameworks, etc. (Gough and Upham, 109 2011), differ for individual plants across different countries and regions. There is currently no 110 comparison of bio-CCS opportunities for individual integrated steel plants, or evaluations of 111 bio-CCS as a strategy for carbon-neutral iron and steelmaking available for the iron and steel 112 industry in Europe. A few studies previously focused on either bioenergy or CCS for iron and 113 steel production in Europe, but to our knowledge, no other studies have considered combining 114 the two technologies. Specifically, both Mandova et al. (2018) and Suopajärvi and Fabritius (2013) conclude that biomass deployment in European iron and steelmaking is limited by 115 116 economic feasibility rather than biomass availability. The CCS studies by Birat (2010) and 117 Remus et al. (2013) on the other hand, point out a lack of sufficient experience with this technology. All of these studies, however, show that neither bioenergy nor CCS would 118 119 achieve a 100% emission reduction in the iron and steel sector on their own. Therefore, 120 research on combining both technologies as bio-CCS is important in order to understand their 121 compatibility, particularly if iron and steel industry aims to achieve carbon neutrality. Such 122 research is also significant to understand the role of other low carbon steelmaking processes 123 that are currently under development, including the use of blast furnaces with top gas 124 recycling (van der Stel et al., 2013), the HIsarna process (Meijer et al., 2011) or hydrogen 125 based steel making (HYBRIT, 2017; Ranzani da Costa et al., 2013).

126 The objective of this work is to evaluate bio-CCS as a strategy for achieving carbon-neutrality 127 across European iron and steel plants that produce steel via the BF-BOF route. Using the 128 techno-economic BeWhere-EU model, the work (1) identifies the importance of bio-CCS 129 within the technology mix when meeting different emission reduction targets, (2) estimates 130 the  $CO_2$  avoidance cost of the bio-CCS deployment, and (3) discusses the potential reduction 131 in CO<sub>2</sub> transport costs by large scale integrated CO<sub>2</sub> pipeline networks. This study bridges the 132 gap in the literature on bio-CCS opportunities in the iron and steel industry and increases the 133 general knowledge on bio-CCS deployment costs in Europe. The outcomes also provide an 134 opportunity to identify potential CO<sub>2</sub> clusters across integrated steel plants, as well as 135 knowledge about possibly integrated CO<sub>2</sub> transport networks.

#### 136 **2 Methodology**

#### 137 2.1 Modelling approach

138 Studying the potential of bio-CCS within a large system requires a modelling approach that accounts for the biomass supply chain, the considered industry, and the CCS network. The 139 140 approach also has to be able to study the interaction between the three systems across the 141 studied time frame, and take into account the spatial distribution of elements as well as the 142 technical limitations that occur when they are applied within the same system. In our previous 143 work using the BeWhere-EU model (IIASA, 2015), we already linked biomass and iron and 144 steel plants in this way (Mandova et al., 2018). This work extends the BeWhere-EU iron & 145 steel model by adding a CCS framework for iron and steel, including CCS linkage to biomass, 146 which provides an opportunity to simultaneously study both the CCS and bio-CCS systems. 147 The section below gives a brief overview of the model, with further information provided in 148 the supplementary material.

The *BeWhere-EU iron and steel* model is written in the General Algebraic Modelling System (GAMS), using Mixed Integer Linear Programming (MILP) and CPLEX as solver. The concept of the model is to split the studied geographic region (EU-28) into equally sized gridcells, each covering an area of  $40 \text{ km} \times 40 \text{ km}$ . Each grid-cell then contains area-specific information that is important for modelling the system, including:

- types, amounts and costs of available feedstock;
- existing biomass demand;
- distance, mode of transport and biomass transport costs between different grid-cells;
- annual CO<sub>2</sub> emissions and energy demand of integrated steel plants;
- CO<sub>2</sub> storage potential, as well as CO<sub>2</sub> capture, transport and storage costs.

159 The cost of biomass upgrading, the types of fossil fuels used in an integrated steel plant, and 160 different CO<sub>2</sub> transport network possibilities are also included in the model. Figure 1 161 illustrates all aspects considered in this work. Based on this information, the model minimises 162 the total cost of the system on an annual basis. The total system cost includes the cost of the 163 biomass supply chain, fuel used in iron and steel plants, as well as all expenditure related to 164 the deployment of CCS. The opportunities for bio-CCS implementations across different 165 plants are then studied by introducing a range of CO<sub>2</sub> emission reduction targets as one of the 166 constraints.



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Figure 1: Aspects considered within the bio-CCS supply chain in this study.

169 As shown in Figure 2, the complexity of the modelled system requires the inclusion of a 170 variety of input data, constraints and internal data calculations. Specifically, the model is 171 composed of three modules, where the core module *BeWhere-EU iron & steel* is using the 172 outputs of the biomass module (labelled *BeWhere-EU*) and the CCS module (labelled  $CO_2$ ) 173 *TranStorage*). In particular, the biomass module is used to subtract the biomass requirement 174 of the existing industries from the total biomass potential. The CCS module has been 175 developed to obtain different CCS infrastructure configurations connecting the plants to potential CO<sub>2</sub> storage sites using a minimum spanning tree algorithm (Hillier, 2012). The core 176 177 - iron and steel - module connects the two modules and provides outputs specific to the iron 178 and steel industry study. A mathematical description of each module can be found in the 179 supplementary material. Table 1 presents a summary of input data values specifically for costs 180 and the following sections give further details on the calculations performed.

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183 Figure 2: Summary of inputs and outputs considered for this study. Values used for each input parameter is

184 provided in the supplementary material.

#### 185 2.2 Biomass supply chain

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The biomass supply chain considers feedstock supply, transport and upgrading. The total 186 theoretical biomass potential within the EU in 2020 is estimated to be 8.5 EJ year<sup>-1</sup>. This 187 188 potential includes stumps, stemwood and logging residues of coniferous and non-coniferous trees, with costs ranging from  $\in 0.20$  up to  $\in 8.30$  GJ<sup>-1</sup> (with price depending on the type of 189 wood and country of origin) (Dees et al., 2017). To incorporate biomass sustainability aspects 190 191 in the modelling, only 70% of the theoretical potential is considered. The model allows inter-192 European biomass trade, as well as biomass imports from non-EU countries to specific 193 harbour locations. The imported biomass from non-EU countries is assigned a cost 20% 194 higher than the average biomass cost in the country where a specific harbour is located, in 195 order to account for additional expenditure due to import taxes and long-distance transport. 196 Biomass harvested outside the EU is generally imported already pre-processed, for example, 197 in the form of pellets. However, as the current work assumes that biomass upgrading to the 198 final product is done on-site of the iron and steel plant, the modelling approach required raw 199 biomass import from outside of the EU. The cost of biomass imports from outside the EU 200 ranges from  $\in 3.56$  to  $\in 6.01$  GJ<sup>-1</sup> (exact values are available in the supplementary material). 201 Transport of biomass from supply points to demand points is considered by truck, train and 202 ship, with the specific cost of each biomass type approximated on energy basis. Form of 203 transport and the corresponding distances are obtained from spatial data using the network 204 analysis tool in the ArcGIS software. The studied biomass demand includes the pulp and

paper industry (total of 1.4 EJ year<sup>-1</sup>) (CEPI, 2017), sawmills (1.6 EJ year<sup>-1</sup>) (FAO, 2016) and heat and power plants (1.0 EJ year<sup>-1</sup>) (Platts, 2017). In total, 2.0 EJ year<sup>-1</sup> of available biomass potentially suitable for iron and steel production is identified from the biomass module (*BeWhere-EU*) after meeting the existing demand. The distribution of the available biomass in relation to the 30 integrated steel plants is shown in Figure 3.

- 210 Table 1: Summary of cost input values considered for this study. Further details are given in the supplementary
- 211 material.

	Input value	Citation	Note		
Biomass feedstock					
Domestic coniferous trees	€0.0 – €6.9 GJ	(Dees et al., 2017)	Spatially explicit prices		
Domestic non-coniferous trees	€0.1 – €8.3 GJ	(Dees et al., 2017)	Spatially explicit prices		
Non-EU feedstock	€3.6 – €6.0 G <b>J</b>		Value 20% higher than average biomass cost in the country of the importing harbour.		
Biomass transport					
Lorry	~€0.00255 GJ <sup>1</sup> km <sup>-1</sup>	$\sim$	Average values dependent on the distance		
Train	~€0.00299 GJ <sup>1</sup> km <sup>-1</sup>		travelled, as defined in a work by Börjesson and Gustavsson (1996), and fuel cost in the country.		
Freight	~€0.00210 GJ <sup>1</sup> km <sup>-1</sup>		Further details are provided in the supplementary material.		
Biomass upgrading					
Pelletisation	€1.03 – €2.98 GJ	(Uslu et al., 2008)			
Torrefaction	€1.28 – €3.72 GJ	(Uslu et al., 2008)	Country specific values defined using purchasing		
Slow pyrolysis	€1.15 – €3.34 GJ	(Norgate et al., 2012)	power parties (European Commission, 2010).		
Fossil fuel cost		$\rangle'$			
Coking coal	€3.98 GJ <sup>1</sup>	(IEAGHG, 2013)			
Coke	€5.35 GJ <sup>1</sup>	(IEAGHG, 2013)	2017 values obtained using a 2010-2017 inflation		
PCI	€3.17 GJ <sup>1</sup>	(IEAGHG, 2013)	rate.		
Coke breeze	€5.35 GJ <sup>1</sup>	(IEAGHG, 2013)			
CO <sub>2</sub> capture cost					
CASE 1:	€54.4 - €93.4 do <sub>2</sub> <sup>-1</sup>	(IEAGHG, 2013)	2017 values obtained using a 2010-2017 inflation rate. Country specific values obtained based on		
CASE 2:	€53.1 – €96.5¢02 <sup>-1</sup>	(IEAGHG, 2013)	(Eurostat, 2017). Further details on calculations performed are given in the supplementary material.		
CO <sub>2</sub> transport cost:					
Individual network	€0.523 - €36.7¢02 <sup>-1</sup>	(IEAGHG, 2005)	2017 values obtained using a 2005-2017 inflation factor. Further details are provided in the supplementary material.		
Collaborative network	€0.191 – €63.3¢0 <sub>2</sub> <sup>-1</sup>	(IEAGHG, 2005)			
CO <sub>2</sub> storage					
Saline aquifers	€15.8 to <sub>2</sub> <sup>-1</sup>	(ZEP, 2011)	2017 values obtained using a 2010-2017 inflation		
Depleted oil and gas fields	€10.8 to <sub>2</sub> <sup>-1</sup>	(ZEP, 2011)	rate.		





Figure 3: Location-specific biomass availability (locally sourced) after the demand from existing bio-based
 industries has been met. Seven trade points for biomass supply from outside of the EU-28 countries were
 considered.

217 Upgrading of any biomass to bio-products: wood pellets, torrefied fuel and charcoal, is assumed to take place on-site at iron and steel plants, at production costs of  $\notin 2.15 \text{ GJ}^1$  for 218 wood pellets (Uslu et al., 2008),  $\in 2.68 \text{ GJ}^1$  for torrefied fuel (Uslu et al., 2008) and  $\in 2.41 \text{ GJ}$ 219 <sup>1</sup> for charcoal (Norgate et al., 2012). The production costs (both converted and original values 220 221 as presented in the supplementary material) have been scaled up or down using purchasing 222 power parity (European Commission, 2016). CO<sub>2</sub> emissions related to biomass harvesting, 223 upgrading and transport are not included, as the study considers only direct emissions based 224 on steel production.

#### 225 2.3 Technologies for CO<sub>2</sub> emission reduction in integrated steel plants

In total, 30 integrated steel plants – the full number of currently operating plants using BF-BOF across EU-28 countries – are considered. In order to maintain transparency under limited data availability and confidentiality, this work assumes that each plant has the same technology and structure as a typical West European plant, as described in the IEA Greenhouse Gas (GHG) report (IEAGHG, 2013). The energy demand of each plant is estimated from the plants' annual hot rolled coil (HRC) production. This is obtained from each plant's data on hot metal production in 2016 (VDEh data exchange, 2017), which is then

further calibrated so that country specific crude steel production corresponds to data published by the World Steel Association for the same year (World Steel Association, 2017). In addition, it is assumed 1 t of hot metal produces 1.113 t of crude steel and 1.027 of hot rolled coil, as presented in the IEAGHG report (IEAGHG, 2013).





*Figure 4: Possibilities for bioenergy integration and post-combustion CO*<sub>2</sub> *capture in an integrated steel plant.* 

239 Substitution of fossil fuels by biomass is considered on an energy basis. Figure 4 240 demonstrates the bioenergy integration possibilities in a typical integrated steel plant for 241 different coal-based fuels. It is important to note, that due to differences between fossil fuels 242 and bio-products in terms of mechanical strength, reactivity, chemical composition, heating 243 value, etc., only partial substitution opportunities are provided (Fick et al., 2014). Table 7 in 244 the supplementary material provides further details on the maximum substitution possibilities 245 of each coal-based fuel by the specific bio-product considered in this work. In the BeWhere-246 EU iron & steel module then, bioenergy is first integrated into the iron and steel plants based 247 on the supply cost in comparison to that of conventional fossil fuels. Generally, the bioproducts are not economically competitive with fossil fuel prices (ranging from €3.52 to €5.94 248 GJ<sup>-1</sup> (IEAGHG, 2013)) and so, no fossil fuel substitution is experienced in the model. 249 250 Therefore, the bio-products are also introduced based on the amount of emissions they could 251 potentially offset, in order to meet the imposed emission reduction targets, while keeping a 252 record of the additional costs incurred by each individual integrated steel plant. These aspects 253 are at the core of the BeWhere-EU iron & steel module and follow the model development 254 process presented in our previous work (Mandova et al., 2018).

The integration of CCS in iron and steel plants is considered in terms of the deployment of post-combustion capture, which can eliminate emissions from existing plants without

significant retrofit. The shorter shut-down time and lower capital investment in comparison to other  $CO_2$  capturing technologies (e.g., pre-combustion capture, oxy-fuel combustion capture or capture from industrial process streams (IPCC, 2005)) makes it a more likely near-term capture option. This work uses the specifications of the  $CO_2$  post-combustion capture technology that incorporates standard monoethanolamine (MEA) solvent for iron and steel plants, as described in the IEAGHG report (IEAGHG, 2013). As per the report, two cases of  $CO_2$  capture possibilities are considered:

- Case 1: CO<sub>2</sub> is captured only from flue gases from the hot stoves and steam generation plant. The net emission intensity of the final steel product (set to 2.09  $t_{CO_2} t_{HRC}^{-1}$ ) can be reduced by a maximum of 50% (to 1.04  $t_{CO_2} t_{HRC}^{-1}$ ) (IEAGHG, 2013).
- Case 2: On top of capturing all CO<sub>2</sub> from the units listed in Case 1, additional CO<sub>2</sub> is captured from flue gases coming from the coke ovens and lime kilns. The maximum CO<sub>2</sub> avoidance potential would increase to 60% (resulting in an emission intensity of 0.828 t<sub>CO2</sub> t<sub>HRC</sub><sup>-1</sup>) (IEAGHG, 2013).
- 271 Because of multiple CO<sub>2</sub> sources across the plant, CO<sub>2</sub> capture across an integrated steel plant 272 is more challenging than, for example, from a power plant. Therefore, despite assuming a 273 90% capture rate for all of the  $CO_2$  absorbers, the other – uncaptured – sources of  $CO_2$ 274 emissions across the integrated steel plant and the increased CO<sub>2</sub> emissions attributed to the 275 extra energy demand from the CO<sub>2</sub> capture installation, results in a net emission reduction of maximum 60%. The estimated CO2 capture cost for each plant in 2017 includes the 276 277 expenditure related to retrofitting the plant and extra energy use. The cost varies across the 278 plants based on national electricity prices for the industry (Eurostat, 2017). In general, the average CO<sub>2</sub> capture costs applied are  $\leq 64.50 \text{ to}_2^{-1}$  and  $\leq 70.40 \text{ to}_2^{-1}$  for the first and second 279 capture case, respectively. The calculations performed can be found in the supplementary 280 281 material. Integration of the different options for post-combustion CO<sub>2</sub> capture within 282 integrated steel plants is illustrated in Figure 4. As CCS avoids the release of CO<sub>2</sub> into the 283 atmosphere, this work assumes zero emission intensity of captured fossil-based CO<sub>2</sub>, and a 284 negative emission value for captured bio-based CO<sub>2</sub>.

#### 285 2.4 CO<sub>2</sub> transport and storage

In terms of considering the transportation of large amounts of  $CO_2$  and probable public opposition to onshore  $CO_2$  storage (Margriet Kuijper, 2011), this work focuses only on  $CO_2$ transport using pipelines for  $CO_2$  deposition in offshore storage locations. In the CCS module

 $(CO_2 \ TranStorage)$  the shortest pipeline network that connects all CO<sub>2</sub> sources with storage locations, is defined. The connections are established by adapting an existing minimum spanning tree algorithm (GAMS, n.d.), the idea of which is to connect all vertices without any cycle, while minimising the total weight of all its edges (Hillier, 2012). To account for obstacles related to the pipeline routing, an extra 10% and 20% are added to the distance (measured as a straight line in ArcGIS) for offshore and onshore pipelines, respectively.

295 The cost of building the pipelines and the final CO<sub>2</sub> transport cost for each plant are 296 calculated using the IEAGHG CO<sub>2</sub> transport cost curves (IEAGHG, 2005), scaled by the 2005 297 to 2017 inflation factor of 1.2 (Official Data Foundation, 2018). A concurrent development of 298 the proposed  $CO_2$  pipeline network is assumed, which is why the extra expenditure resulting 299 from gradual CO<sub>2</sub> network development that would likely evolve in practice, is not 300 considered. In addition, the network focuses only on connecting the 30 integrated steel plants, 301 excluding possibilities for network connection with other plants (such as power, heat, cement, 302 chemicals, etc.) and the corresponding possibilities for further cost reductions due to 303 economies of scale.

The key factors influencing the cost are the pipeline length and the specific  $CO_2$  flow. The CO<sub>2</sub> transport cost estimates also include the cost of compression up to supercritical pressure (above 73.8 bar), investment, operational and maintenance costs, as well as whether it is an onshore or offshore pipeline (IEAGHG, 2005). In addition, the calculation also takes into account the extra CO<sub>2</sub> flow as a result of increasing the amount of CO<sub>2</sub> produced at a plant due to the installation of CCS technology. A further description of the CO<sub>2</sub> pipeline cost calculations can be found in the supplementary material.

311 As mentioned above, only offshore CO<sub>2</sub> storage in saline aquifers or depleted oil and gas 312 fields is considered, with locations around Europe shown in Figure 5. The storage/injection 313 capacities are obtained from the Chalmers CO<sub>2</sub> storage database (Kjärstad and Johnsson, 314 2007). The storage and injection capacities, particularly in aquifers, are highly uncertain. The 315 values listed in the Chalmers CO<sub>2</sub> storage database should therefore be considered as rough preliminary estimates. The cost of CO<sub>2</sub> storage is set to  $\leq 10.80 \text{ t}_{\text{CO}_2}^{-1}$  for depleted oil and gas 316 fields and  $\leq 15.60 \text{ t}_{\text{CO}_2}^{-1}$  for saline aquifers (ZEP, 2011) (scaled by an inflation factor of 1.09 317 318 for 2010 to 2017 (Official Data Foundation, 2018)).





Figure 5: Locations of CO<sub>2</sub> sources and offshore storage locations relative to the location of integrated steel
 plants. Data on storage locations taken from Chalmers CO<sub>2</sub> storage database (Kjärstad and Johnsson, 2007).

#### 322 **2.5 Scenario setting**

323 To help answer our questions, we explore a range of scenarios that vary across two 324 dimensions: (1) the  $CO_2$  emission reduction goal to be achieved, and (2) the configuration of 325 the physical  $CO_2$  infrastructure.

326 To study the increasing importance of bio-CCS in the technology mix, we impose European 327 emission reduction targets ranging from 0 up to 100%, with a 5% step level. The analysis 328 focuses only on the CO<sub>2</sub> emissions occurring on-site for the integrated steel plants, in other 329 words, it does not consider the produced emissions during fuel transportation, upgrading or 330 production as such a study would require a detailed Life Cycle Analysis (LCA). The follow 331 up discussion takes place on both plant and country level, in order to evaluate whether any 332 country has an outstanding opportunity for bio-CCS deployment that would be able to 333 significantly reduce CO<sub>2</sub> emissions on its own.

To account for the possibility of several plants sharing a  $CO_2$  pipeline system, two  $CO_2$ networks, classified as individual or collaborative, are considered (Figure 6). In both cases, the costs are calculated for a "plateau flow" of  $CO_2$  (a  $CO_2$  pipeline network where all plants start delivering their maximum  $CO_2$  volumes from day one). It is important to note that

- achieving the proposed collaborative network would be difficult in practice since it is unlikelythat all plants will deploy CCS/bio-CCS at the same time.
- A number of non-economic barriers that can potentially influence  $CO_2$  pipeline construction can be identified. This includes, for example, the 1996 London Protocol prohibiting the export of  $CO_2$  for storage (International Maritime Organization, 2006), expected local opposition (Margriet Kuijper, 2011) or previous studies disclosing certain pipeline networks.



Figure 6: Notional a) individual vs. b) collaborative CO<sub>2</sub> pipeline network based on minimum distance criteria
 and capacities of the CO<sub>2</sub> storage reservoirs.

#### 347 **3 Results**

#### 348 **3.1** The importance of bio-CCS for various CO<sub>2</sub> reduction targets

349 The optimal technology mix to meet different CO<sub>2</sub> emission reduction targets is shown in 350 Figure 7. After considering the three technologies – biomass, CCS, and bio-CCS – it emerged 351 that the application of bio-CCS is required across all plants to achieve a 100% CO<sub>2</sub> reduction (of 189  $Mt_{CO_2}$  year<sup>-1</sup>) within the European iron and steelmaking industry. However, the 352 deployment of bio-CCS is not the most favourable technology for all plants in terms of 353 meeting low EU emission reduction targets. As Figure 7 demonstrates, the deployment of 354 biomass on its own is a key strategy to reduce up to 20% (38  $Mt_{CO_2}$  year<sup>-1</sup>) of the total CO<sub>2</sub> 355 356 emissions coming from integrated European steel plants. In addition, all countries provide a 357 similar share of CO<sub>2</sub> emission reduction in relation to their total emissions for the lower 358 targets. This demonstrates that no individual country would present an outstanding 359 opportunity for the quick introduction of low-cost biomass that would in turn help to 360 significantly reduce the total iron and steelmaking related emissions in the EU. Rather, the

361 results show that a collaborative effort from all plants is necessary. For targets above a 20% 362 reduction, a new technology (CCS) is introduced on top of the old one (from here on referred 363 to as bio-CCS), particularly for plants in the Netherlands, France, Sweden and Belgium. At a 364 50% emission reduction target, the bulk of the reduction is met by installations of bio-CCS, 365 which becomes the key technology for meeting any targets beyond the 50% mark. Germany 366 and the United Kingdom (UK) are the last countries seen to introduce a shift from biomass to 367 bio-CCS. The figure also shows that no country introduces CCS without also including biomass at any target. These results demonstrate that for European integrated steel plants, 368 369 biomass or bio-CCS is preferable over the deployment of CCS alone.



Figure 7: Changes in the technology mix based on different targets imposed on total  $CO_2$  emissions from the European iron and steel plants. Pure CCS technology is not represented as it was never selected.

373 Overall, the resulting maximum achievable emission reduction for the steel plants is 191 374  $Mt_{CO_2}$  year<sup>-1</sup>, which would lead to a negative emission potential of 2  $Mt_{CO_2}$  year<sup>-1</sup>. This result, 375 however, cannot be seen as significant due to the estimated error range of the obtained results, 376 and so no negative emission opportunities across the European iron and steel industry are 377 presented.

#### 378 **3.2** CO<sub>2</sub> avoidance cost of bio-CCS

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Figure 8 shows that the  $CO_2$  avoidance cost of emissions due to the deployment of biomass and of CCS within a bio-CCS system are comparable on plant level, particularly when comparing high levels of biomass substitution with the lowest costs of CCS deployment.

Complete CO<sub>2</sub> emission reduction across European iron and steel plants using bio-CCS will cost on average  $\notin 80 \ \text{to}_2^{-1}$  avoided, ranging from  $\notin 59 \ \text{to}_2^{-1}$  for a plant in France to  $\notin 97 \ \text{to}_2^{-1}$ for a plant in the UK.

385 The range of the CO<sub>2</sub> avoidance costs of bio-CCS is due to different economics behind the 386 deployment of biomass and CCS in each plant. For example, avoiding CO<sub>2</sub> emissions using biomass costs on average  $\leq 61 \text{ t}_{\text{CO}_2}^{-1}$  at the maximum technically-feasible substitution. For the 387 plant in Romania however, the CO<sub>2</sub> is avoided using biomass at costs as low as  $\notin 40 \text{ to}_{2}^{-1}$ . 388 389 The lower estimate of the CO<sub>2</sub> avoidance cost using biomass for certain plants can be 390 explained by a combination of factors, including the availability of cheap feedstock in the 391 plant vicinity, short transport distances between the feedstock supply locations and the plant, 392 or competitive prices for feedstock upgrading to the final bio-products in the countries where 393 the plants are located.

394 The economics of CCS on the other hand, are influenced by the distance of the plants to the 395 storage locations, the amount of  $CO_2$  transported annually, the type of  $CO_2$  storage reservoir, as well as country-specific electricity prices. The resulting average CO2 emission reduction 396 cost using CCS technology is estimated at  $\in 92 \text{ t}_{\text{CO}_2}^{-1}$  avoided. This cost includes the 397 398 technology investment, as well as the operational cost related to CO<sub>2</sub> capture, transport and its 399 injection into the reservoirs. In general, CCS deployment is the most expensive for plants in 400 Germany and the UK, as the biggest expense related to CCS deployment is the CO<sub>2</sub> capture 401 cost (around 76% of the overall CO<sub>2</sub> avoidance cost), which is heavily influenced by the cost 402 of electricity in the country.

403 Initial biomass substitution is cheaper than the deployment of CCS, as the CO<sub>2</sub> avoidance cost 404 for CCS technology exceeds the CO<sub>2</sub> avoidance cost for initial biomass substitution, as presented in Figure 8. However, plants in the Netherlands and Belgium have CO<sub>2</sub> avoidance 405 costs by bio-CCS that exceed the costs of CCS on its own ( $\leq 67 \text{ to}_2^{-1}$  and  $\leq 64 \text{ to}_2^{-1}$  for the 406 Netherlands, and  $\in 81 \text{ t}_{\text{O_2}}^{-1}$  and  $\notin 71 \text{ t}_{\text{O_2}}^{-1}$  for Belgium, for bio-CSS and CCS, respectively). 407 408 In these cases, biomass is economically preferable to CCS for only very low emission 409 reduction levels, and the introduction of CCS on top of biomass is expected even at lower 410 emission targets, before the maximum technically feasible substitution by biomass is 411 achieved. It is important to note that zero emissions across European integrated steel plants 412 can only be reached at maximum biomass substitution in combination with full CCS 413 deployment.



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igure 8:  $CO_2$  avoidance cost of bio-CCS application for each plant achieved when meeting different  $CO_2$  reduction targets across the whole European iron and steel industry.

418 **3.3** 

#### The role of CO<sub>2</sub> transport and possibilities for cost reduction

CO<sub>2</sub> transport cost constitutes only a relatively small part of the CO<sub>2</sub> avoidance cost using 419 420 bio-CCS, (on average 6% of the total cost). The potential reduction of the CO<sub>2</sub> transport cost 421 when applying a collaborative  $CO_2$  pipeline network instead of an individual one is studied in 422 Figure 9. The figure demonstrates both plants for which collaborative networks will not 423 provide any significant CO<sub>2</sub> transport cost benefits (plants located close to the central line), 424 and plants for which cluster networks will result in significant reductions of the CO<sub>2</sub> transport 425 costs (plants in the coloured area). As can be observed, the biggest iron and steel plants (located in the zoomed-in box of transport costs of  $\notin 7 \text{ t}_{\text{CO}_2}^{-1}$  or less) do not significantly divert 426 427 from the central slope line. Hence, it can be seen that the big iron and steel plants would not 428 gain a significant economic advantage from collaborative CO<sub>2</sub> pipeline networks, due to the 429 large volumes that will be transported from these plants already. On the other hand, 430 collaborative CO<sub>2</sub> networks would significantly benefit smaller iron and steel plants. Cost 431 reductions exceeding 60% could be expected for the small plants in Austria, Hungary and 432 Poland, while for the smallest plants in Germany and Italy, the results show possible cost 433 reductions of over 90%. Medium plants in Slovakia, Czech Republic, Finland, etc. could also 434 benefit from collaborative pipeline networks, with transport cost reductions between 10 and 435 20%. The Swedish plant in Oxelösund (SWE2) is the only plant for which a collaborative

436 pipeline network would be unprofitable, due to a significant increase in the total  $CO_2$  transport 437 distance from this plant. Potential storage sites have been identified in the Swedish part of the 438 Baltic Sea, just 250 km southeast of the Oxelösund plant but storage and injection capacity in 439 these reservoirs are still highly uncertain due to a lack of data (Rokke et al., 2016). Moreover, 440 both potential storage sites identified in the Swedish part of the Baltic Sea are classified as 441 Natura 2000 areas which possibly could have effect on activities related to transport and 442 injection of  $CO_2$ (Natur Vards Verket, 2018).

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Figure 9: Impact of collaborative  $CO_2$  pipeline network on  $CO_2$  transport cost, compared to individual networks. Plants located close to the bottom right corner would experience the greatest cost reduction from the collaborative pipeline network. The closer a plant gets to the central line the less cost reduction per  $t_{CO_2}$  transported can be expected from joining the collaborative pipeline.

## 449 4 Discussion: Perspective for bio-CCS deployment across European 450 integrated steel plants – from modelling to reality

451 The modelling results demonstrate that bio-CCS can achieve a 100% CO<sub>2</sub> emission reduction 452 across European integrated steel plants. However, these results are related to the emissions 453 occurring only on-site, and rely heavily on the assumption of carbon neutrality of biomass. As 454 emissions of the bio-CCS system are also produced off-site due to land use change, biomass 455 harvesting, transport and upgrading, as well as due to CO<sub>2</sub> capture, transport and storage, iron 456 and steelmaking in Europe would not be carbon-neutral from the whole system perspective. 457 For example, work by Fajardy and Mac Dowell (2017) calculated (for a specific case of US 458 switchgrass and BECCS application) that technically, only 45% of the geologically stored 459 biological-based CO<sub>2</sub> emissions could be considered as negative emissions. Therefore, the deployment of biomass or bio-CCS in the iron and steel industry could still result in a 460 461 significant amount of emissions contributing to the total European carbon budget. A detailed 462 LCA specific to each plant would be required to estimate the real environmental benefits of 463 those technologies.

With increasing biomass demand from other sectors also looking to reduce their CO<sub>2</sub> 464 465 emissions (e.g., as feedstock for transportation fuel production or for the chemical industry), 466 the biomass market can be expected to undergo significant transformations, which may in turn 467 lead to price increases. Olofsson (2018) analysed the impact on regional biomass markets of 468 introducing biomass to an integrated steel plant in Sweden (SWE1, in this study). He found 469 that while the total welfare effect in the region would be relatively small, certain market 470 segments, in particular regarding secondary biomass, could potentially be heavily affected, 471 leading to significant price effects for both the steel plant and other biomass users in the 472 region.

473 The introduction of bio-CCS can present a valuable opportunity for CO<sub>2</sub> emission reduction 474 and the defossilisation of the European iron and steel industry, which could also be 475 deployable on a relatively short term. The creation of an economic environment within the EU 476 and characterised by policy certainty (for example, giving extra credits under the EU-ETS 477 system for bio-CCS) that would make the investments in CCS/bio-CCS a strategic decision for the industry (ZEP, 2018), is key for this transition. The average CO<sub>2</sub> avoidance cost of €80 478  $t_{CO_2}$ -1 identified in this work would translate to a noticeable increase in steel production cost. 479 Even though Rootzén and Johnsson (2016) argued that a carbon price of  $\notin 100 \text{ to}_2^{-1}$  would 480 481 increase the price of the final steel product (e.g., a car) by only a tiny fraction, the economic

disadvantage of European steel against cheap imports from particularly China, might be further enhanced. This could in turn lead to plant shutdowns, which would also create a significant impact further down the line of the value chain by, for instance, losing a high number of steel-related jobs in Europe. Therefore, bio-CCS, especially in the European iron and steel industry, will not be deployed without a valid economic case and a stable policy regime.

488 Apart from economic barriers, the application of bio-CCS might not be possible due a variety 489 of social, technical and legislative issues, mostly related to CO<sub>2</sub> transport and storage. While 490 the inclusion of these aspects in the modelling was outside the scope of this work, it is, 491 however, still important to highlight them. The integrated steel plants would have to 492 overcome issues such as negative public perception, uncertainties in CO<sub>2</sub> storage capacities 493 around Europe, issues related to the 1996 London Protocol, and temporary bans on onshore 494 CO<sub>2</sub> storage in some countries, even though these issues are occurring outside of their 495 borders. However, as has been shown in this work, the costs of CO<sub>2</sub> transport and storage 496 constitute minor contributions towards the total cost of CCS/bio-CCS deployment, and non-497 economic barriers related to those parts might be of decisive importance.

498 If bio-CCS is excluded as a technology option, the maximum emission reductions are limited 499 to 20% by exclusively using the best presently available technologies. The deployment of 500 innovative technologies that are currently in development or pilot scales would thus be 501 necessary to meet the targets for the iron and steel industry (Pardo and Moya, 2013). Of the 502 emerging technologies, top gas recycling, which requires the retrofitting of the existing blast 503 furnace fleet, is closest to application (Moya and Pardo, 2013). HIsarna or direct reduction 504 processes such as ULCORED, Midrex, HYL or ULCOWIN are also being discussed, even 505 though their deployment is currently facing either technology readiness issues (expected by 2030 or even 2040) or economic barriers (CO<sub>2</sub> avoidance costs of over  $\notin 100 \text{ to}_{2}^{-1}$ ) (Pardo 506 507 and Moya, 2013). Opportunities for iron ore reduction using hydrogen, such as the HYBRIT 508 (HYBRIT, 2017) and H2FUTURE ("H2FUTURE Green Hydrogen," n.d.) projects in Sweden 509 and Austria, respectively, are now also becoming available. By 2035, the industry hopes to 510 have a process in place (Vattenfall AB, 2018) that could play a leading role in European iron 511 and steel making from 2050 onwards (Sgobbi et al., 2016). It is not possible to predict which 512 technologies and/or combinations of technologies are likely to emerge, but emission 513 reductions beyond 40% will still mean their co-application with CCS (EUROFER, 2013). 514 Therefore, overcoming CCS barriers should be a priority if CCS were to become the key 515 technology for emission reduction in this industry in the near future (ZEP, 2018). The

516 introduction of bio-CCS could achieve high emission savings in a relatively short time, since 517 bio-CCS requires comparatively small retrofits to plants, while the more innovative 518 technologies still face considerable research and development before they will be ready to be 519 deployed.

#### 520 **5 Conclusion**

521 This work explores the CO<sub>2</sub> emission reduction potential of bio-CCS in integrated steel plants 522 across the EU and compares opportunities for its deployment across the 30 operating plants. 523 Our findings show that bio-CCS can play a role in achieving carbon-neutrality across these 524 plants when considering only emissions produced on-site. However, bio-CCS would not be an 525 economically favourable option when aiming to reach specific CO<sub>2</sub> emission reduction targets 526 below 20% for which an autonomous deployment of biomass over full bio-CCS is more favourable. Therefore, biomass can be considered a strategic solution for an initial 527 decarbonisation, of which the CO<sub>2</sub> emission reduction potential could be enhanced through 528 529 the additional deployment of CCS (resulting in bio-CCS), if required.

In this study, an average CO<sub>2</sub> avoidance cost using bio-CCS in European iron and steel plants 530 is calculated to  $\notin 80 \notin o_2^{-1}$ . This is indeed a large additional expenditure that would 531 significantly increase the steel production cost of the plants, even for the most suitable ones. 532 The work shows that an initial biomass substitution is cheaper than CCS deployment, but then 533 534 costs related to the high level of biomass utilisation are similar to the deployment cost of 535 CCS. Despite CO<sub>2</sub> capture accounting for the biggest share of CO<sub>2</sub> avoidance cost by CCS, 536 the opportunities in cost reduction actually emerge in CO<sub>2</sub> transport as plants start sharing CO<sub>2</sub> pipeline networks. Especially for small integrated steel plants, the CO<sub>2</sub> transport cost 537 538 could be reduced by up to 90%. Opportunities for the reduction of CO<sub>2</sub> capture costs could 539 also occur in the future. Cost of a first-of-a-kind capture plant is usually significantly greater 540 than the cost of a mature nth-of-a-kind (Rubin et al., 2015). This has been demonstrated at, for 541 example, the Shand power plant, based on lessons learnt from the Boundary Dam, or 542 discussed in a work by van den Broek et al. (2009). Hence, there is a high likelihood that the CO<sub>2</sub> avoidance cost of using bio-CCS could be even lower than  $\in 80 \text{ t}_{\text{CO}_2}^{-1}$  in the future. 543 544 However, in the present, a significant cost reduction of bio-CCS is difficult, and the EU has to 545 propose stronger economic incentives that would ensure a competitive iron and steel industry 546 in the EU, if carbon-neutrality using bio-CCS is defined as the way to go.

547 From specifically a geographical viewpoint, no country presents an outstanding opportunity 548 for bio-CCS. In general, the technology is most likely to be developed in France, the 549 Netherlands, Belgium and in one of the plants in Sweden, since these plants achieve the 550 lowest bio-CCS deployment costs. On the other hand, the least favourable countries are 551 Germany and the UK due to the comparably high costs of CO<sub>2</sub> capture.

It is important to mention that if we want bio-CCS to be developed at a large scale in Europe, non-economic barriers of a regulatory-social-environmental nature must also be resolved, or at least accounted for in the policy agenda. Further study is necessary to identify the most essential problems that the EU or specific countries and regions are facing. It is recommended that a sensitivity analysis of the impact of overcoming barriers on the  $CO_2$  avoidance cost for each plant shown in this work be included in such a study.

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