# Global urban carbon networks: linking inventory to modelling

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#### 1 Abstract

2 Cities utilize and manipulate an immense amount of global carbon flows through their 3 economic and technical activities. Here, we establish the carbon networks of eight global cities 4 by tracking the carbon exchanges between various natural and economic components. The 5 metabolic properties of these carbon networks are compared by combining flow-based and 6 interpretative network metrics. We further assess the relations of these carbon metabolic 7 properties of cities with their socioeconomic attributes that are deemed important in urban 8 development and planning. We find that though there is a large difference in city-level carbon 9 balance and flow pattern, a similarity in inter-component relationships and metabolic 10 characteristics can be found. Cities with lower per capita carbon emissions tend to have 11 healthier metabolic systems with better cooperation amongst various industries, which indicates there may be synergy between urban decarbonization and metabolic system optimization. 12 13 Combination of indicators from flow balance and network models is a promising scheme for 14 linking carbon inventories to metabolic modelling efforts. With this done, we may be able to 15 fill the knowledge gap in current practices of carbon mitigation priorities as to how various 16 carbon flows in cities can be concertedly managed according to urban economic and 17 demographic changes.

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## 24 **1. Introduction**

25 Cities are a major contributor to greenhouse gas emissions and will probably remain so in the foreseeable future. With an occupation of less than 3% of global land surface,<sup>1,2</sup> urban areas 26 27 account for approximately 70% of global carbon emissions owing to the concentration of production and consumption activities.<sup>3</sup> In the coming decades, major growth of carbon 28 29 emissions will take place in many cities around the world as long as these cities continue their 30 carbon-intensive economic growth and land-use expansion, which is particularly the case for cities in less developed countries and regions.<sup>4, 5</sup> This poses a great challenge with regard to 31 achieving the 1.5° C global climate target<sup>6</sup> and the United Nations Sustainable Development 32 Goals (SDGs) for climate action and sustainable cities.<sup>7</sup> Cities can also contribute to the 33 34 decrease of the global carbon footprint owing to the consolidating urban population.<sup>8</sup> However, 35 their role in decarbonization is partially obfuscated by high diversity in socioeconomic status 36 and biogeochemical cycles.

37 Motivated by anti-global warming action, scientists struggle to establish carbon mitigation 38 approaches that can be applied to cities in different stages of development and with different economic structures, demographics, and climatic conditions.<sup>9-11</sup> One increasingly important 39 40 approach is to track both in-boundary and trans-boundary carbon emission associated with urban metabolic activities.<sup>12-16</sup> However, this requires that the data needed for trade models 41 42 (such as input-output tables) are accessible for cities. An alternative method is to consider 43 carbon emissions embodied in products by fusing material flow analysis (MFA) and life-cycle analysis (LCA).<sup>17,18</sup> In these methods, carbon emissions are quantified based on the energy and 44 materials consumed by urban economic sectors. An important feature of this integrated 45 approach is that it can be directly linked to carbon cycle models<sup>19-21</sup> by placing emissions in a 46 47 broader urban carbon metabolism. It is essential to track all physical carbon flows in cities 48 because from a systemic perspective, all the activities in economic sectors including the 49 consumption of carbon products (fossil fuels or non-fossil fuels products) will have an impact on carbon waste and emission via natural and economic transactions.<sup>21,22</sup> In addition to existing 50

carbon inventories, network-oriented model<sup>23,24</sup> can provide a supplementary approach for
 identifying carbon metabolic patterns in cities.<sup>25</sup>

53 Ecological network analysis (ENA) has been noted for its usefulness in uncovering flow 54 structures and patterns in biological systems<sup>26-28</sup> and more recently, its adaptability in humandominated systems.<sup>23,24</sup> ENA offers a set of powerful modelling approaches and metrics that 55 have already been used to support decision making in sustainable resource management.<sup>29-31</sup> 56 57 There have been studies establishing ENA models to track carbon metabolic pathways 58 associated ecological and economic activities in cities<sup>25,32,33</sup> and applied them to show the possible pathways for more efficient spatial urban planning and carbon mitigation.<sup>34,35</sup> Since 59 60 ENA metrics do not directly represent the dynamics of economy, the effectiveness and 61 implications of network metrics for the decarbonization of cities with highly diverse 62 geographical and economic traits may be better understood on a comparative basis.<sup>36</sup> In ecosystems, there frequently exists a common rule or pattern governing the carbon balance.<sup>37,38</sup> 63 64 It is natural to ask whether there are some common properties of carbon metabolic system 65 comprising of natural and economic components and how they are linked to urban development 66 and carbon emission mitigation. Currently, the interplay between metabolic properties and 67 socioeconomic properties has only been assessed in one single city that are focused on carbon exchanges among economic sectors rather than all relevant urban components.<sup>39</sup> 68

69 This study establishes the carbon networks of eight global cities based on a city-level energy 70 and material dataset, which captures carbon flows between urban economic sectors and natural 71 components. Combing flow-based metrics and interpretative network metrics, we identify and 72 compare the structure, patterns, and processes of urban carbon metabolism of high geographical 73 and economic diversity. We further assess the relations of the system properties of carbon 74 metabolic networks with socioeconomic attributes that are deemed important in urban 75 development and planning. The two categories of metrics developed can be a promising scheme 76 for linking carbon inventories to metabolic modelling efforts. Equipped with this network-77 oriented approach, we may be able to address how various carbon flows in cities can be

78 concertedly managed according to urban economic and demographic changes.

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# 80 2. Materials and Methods

## 81 **2.1. System diagram and technical framework**

82 A system diagram of the Carbon Flow Network (CFN) is shown in Figure 1. Urban 83 carbon flows are embedded in an urban metabolic system wherein natural and artificial (human-84 dominated) compartments are interacted with each other. These 13 aggregated compartments 85 can be classified into four modules: 1) seven economic sectors, including Agriculture, forestry 86 and horticulture (Agr), Mining (Min), Manufacturing (Man), Electricity, gas, and water (Ele), 87 Construction (Con), Transportation (Tra), and Services (Ser); 2) two residential components, 88 domestic consumption (Dom) and governmental consumption (Gov); 3) two components 89 related to natural ecosystems: carbon stock variation (Sto) and biodegradable waste, such as 90 food residues and other biodegradable waste (Dwa); 4) two components of environmental 91 distribution: gaseous emissions (Ems) and non-biodegradable waste (Nwa). The CFN is 92 established based on a quantification of inter-component carbon flows through the integration 93 of material flow analysis (MFA), activity-based carbon inventory and life-cycle analysis (LCA). 94 We then assess the performance and pattern of a CFN and how they are related to the 95 socioeconomic attributes based on two categories of metrics (flow-based metrics and 96 interpretative network metrics).



98 **Figure 1** System diagram and technical framework for carbon flow network (CFN).

Note: Agr: Agriculture, forestry and horticulture; Min: Mining; Man: Manufacture; Ele: Electricity, gas
and water; Con: Construction; Tra: Transportation; Ser: Services; Dom: Domestic consumption; Gov:
Governmental consumption; Ems: gaseous emission; Dwa: biodegradable waste; Nwa: Nonbiodegradable waste; Sto: Stock variation. MFA: material flow analysis; LCA: life-cycle analysis; ENA:
ecological network analysis

### 104 **2.2. Carbon flow inventory**

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105 We used material flow analysis (MFA) to quantify the carbon flows and stock changes 106 of urban economic sectors. MFA plays a significant role in determining the growth of urban 107 metabolism, and can provide a strong foundation for assessing the impact of economic activities on natural ecosystems.<sup>18,40</sup> Additionally, MFA has great potential for linking with global 108 biogeochemical cycles.<sup>41</sup> In the carbon metabolic system of a city, it is important to consider 109 110 the physical fluxes embedded in products and gaseous emissions from economic sectors (including in-boundary energy-use related emission as well as those emitted from the 111 generation of imported electricity). These two parts form the major structure of a city's carbon 112 113 profile, and are both considered in this study.

114 First, because direct carbon flow data are rare for most cities, we converted the mass-based

flows in products (derived from local survey and published literature, see data compilation in Section 2.6) to carbon flows by multiplying the material mass with a ratio called carbon content factor (a) to obtain information on how much carbon is contained in assorted products. The sector-specific carbon appropriation can be calculated from the aggregation of the productspecific carbon contents, as follows:

120 
$$C_i = \sum_{x=1}^{n} \alpha^x M_i^x$$
 (1)

121 where  $C_i$  is the carbon appropriated by an urban component;  $M_i^x$  is the weight of a certain 122 type of product *x* consumed by component *i*;  $\alpha^x$  is the corresponding carbon content factor of 123 that product. The carbon content factor varies in different types of products, such as fuel and 124 biomass, agricultural and food products, and industrial and construction materials, as reported 125 in the literature.<sup>42-47</sup> Forestry products used by the cities such as industrial roundwood and 126 household wooden furniture are included in the carbon flow inventory, but the specific climatic 127 impact of land use change is not considered.

Secondly, it is widely recognized that both in-boundary carbon emissions and crossboundary emissions from electricity consumption should be considered in urban carbon accounting.<sup>10,18</sup> To calculate the flows of all urban components to Ems, we compiled an inventory of direct  $CO_2$  emissions from all economic sectors based on the approach recommended by the Intergovernmental Panel on Climate Change.<sup>48</sup> The electricity-related carbon emissions outside of the urban boundary are also quantified and combined based on the respective carbon coefficients. The carbon flow to Ems is formulated as follows:

135 
$$C_{Ems(i)} = \sum_{k=1}^{\infty} E_i^k \times \omega_i^k + U_i \times \omega_i^{ele}$$
(2)

where  $C_{Ems(i)}$  is the total amount of carbon emissions from economic sector *i*, and  $E_i^k$  is the energy combustion from a certain fuel type or the intensity of a certain industrial process (*k*);  $\omega_i^k$  denotes the respective CO<sub>2</sub> emissions coefficient for urban energy use or industrial processes;  $U_i$  is the import of external electricity;  $\omega_i^{ele}$  denotes the CO<sub>2</sub> emissions coefficient of electricity (depending on the energy mix in power generation).

141 **2.3 Establishment of carbon networks** 

Fath and his colleagues<sup>49</sup> proposed a step-by-step procedure for establishing ecological 142 143 networks. This procedure includes three key processes: 1) determining the nodes and allowing 144 to capture the interactions amongst different compartments; 2) quantifying the input, output, 145 and throughflows between different compartments; 3) finalizing the network using a flowbalancing technique that has been widely applied to different types of networks.<sup>30</sup> This can be 146 147 extended to the development of urban carbon flow networks. Here, a node of the carbon 148 network refers to the economic and ecological components of a city, while the arrow refers to 149 the carbon flows between components. The carbon imported to one component is equal to the 150 carbon transferred to other components through the production and consumption of products; 151 that is, the sum of all carbon inflows is equal to the sum of all carbon outflows (the stock change 152 is included as an outflow because it is considered as a component). In matrix terms, the row 153 sum and column sum are the same. The system balance of the carbon flow network is expressed 154 as follows:

155 
$$T_i^{in} \equiv z_i + \sum_{j=1}^{in} f_{ji}$$
 (3)

156 
$$T_i^{out} \equiv \sum_{j=1}^{i} f_{ij} + y_i$$
 (4)

$$157 T_i^{in} = T_i^{out} (5)$$

where  $T_i^{in}$  and  $T_i^{out}$  represent the total amount of flow input to and output from each urban component, respectively;  $f_{ij}$  is the carbon flow from component *i* to *j*;  $z_i$  denotes the boundary inflows (external import) to component *i*;  $y_j$  denotes the boundary outflows (export to other regions) from component *j*.

#### 162 **2.4 Assessment of carbon network pattern and performance**

We adapted a set of system indicators from ecological network analysis (ENA)<sup>50</sup> and information theory<sup>51</sup> to identify the metabolic pattern of carbon flows and comprehensively assess the properties and functioning of carbon networks of cities. The application of networkbased indicators and tools in natural and human-dominated systems has been extensively discussed.<sup>39,52-55</sup> In this study, we employed two categories of metrics to uncover the system properties of carbon flow networks and how they are related to the socioeconomic development in cities.

170 (1) Flow-based metrics (FBMs)

FBMs are represented by total system throughflow, boundary flow, cycled flow and Finn cycling index, which are grounded on physical laws and are widely used in the description of natural and human human-dominated systems.

174 The total system throughflow (TST) accounts for the sum of throughflows of all 175 components. We used the TST of carbon to represent the size of a city's carbon metabolism, 176 which does not only include gaseous emissions but also other physical carbon flows. Thus, we 177 acquired a wider perspective with regard to how much carbon is appropriated by a city. The 178 boundary flow (BF) is a subset of TST that captures the import of carbon from outside of the 179 urban boundary, or the export of carbon to other regions or systems (in an equilibrium state 180 these two are equal). This clarifies the reliance of urban carbon metabolism on external markets and ecosystems. The cycled flow (CF) can be derived from the diagonal elements of the integral 181 flow matrix (N), and is used to investigate the carbon cycled in the urban ecosystem through 182 direct and indirect paths. The formulation of TST, BF, and CF is expressed as follows: 183

184 
$$TST \equiv \sum_{i=1}^{\infty} T_i^{in} = \sum_{i=1}^{\infty} T_i^{out}$$
(6)

185 
$$BF = \sum_{j=1}^{N} Z_j = \sum_{i=1}^{N} y_i$$
 (7)

186 
$$N = [n_{ij}] = \sum_{n=0}^{\infty} G^n = (\mathbf{I} - \mathbf{G})^{-1}$$
(8)

187 
$$CF = \sum_{j=1}^{\infty} \left( \frac{n_{jj} - 1}{n_{jj}} \mathbf{T}_j \right)$$
(9)

188

189 where N=[ $n_{ij}$ ] is the integral dimensionless matrix of metabolic flow, and G is the direct 190 dimensionless matrix of metabolic flow, <sup>50</sup> where  $g_{ii} = f_{ii}/T_i$ 

Finn cycling index (FCI)<sup>56</sup> was proposed to measure the amount of recycled flow compared with the total flow processed in a network, and was formulated based on the CF and TST results. Notably, FCI is not the recycling rate of carbon in the urban economy, but rather the carbon transferred amongst the components' circular supply chains.

195 
$$FCI = \sum_{j=1}^{\infty} \left(\frac{n_{jj} - 1}{n_{jj}} T_j\right) / \text{TST}$$
(10)

#### 196 (2) Interpretative network metrics (INMs)

197 INMs include centrality, control allocation and dependence allocation, ascendancy,
198 capacity, system robustness, synergism, which are based on theoretical ecological network
199 models that need to be interpreted for applications in human-dominated systems.

200 Network control analysis (NCA) has been proposed to quantify the dominance of one network component over another.<sup>26,59</sup> Previous studies have demonstrated that NCA can 201 effectively reveal inter-component relationships and dynamics, and identify the key processes 202 203 in urban metabolic networks.<sup>34,60</sup> This provides an advantage in targeting the most influential 204 activities in terms of carbon emissions and waste, and can therefore assist in designing a more 205 efficient method of urban decarbonization. In this study, we used the control metrics proposed by Chen and Chen,<sup>25</sup> namely, the control allocation (CA) and dependence allocation (DA) to 206 207 assess the control and dependence inter-component relationships with regard to urban carbon 208 exchanges.

209 
$$N' = (\mathbf{n}_{ij}) = \sum_{n=0}^{\infty} G^m = (\mathbf{I} - \mathbf{G}')^{-1}$$
 (11)

210 
$$CA = \left[ ca_{ij} \right] = \begin{cases} n_{ij} - n_{ji} > 0, ca_{ij} = \frac{n_{ij} - n_{ji}}{\sum_{i=1}^{m} n_{ij} - n_{ji}} \\ n_{ij} - n_{ji} \le 0, ca_{ij} = 0 \end{cases}$$
(12)

211 
$$DA = \left[ da_{ij} \right] = \begin{cases} n_{ij} - n_{ji} > 0, da_{ij} = \frac{n_{ij} - n_{ji}}{\sum_{j=1}^{m} n_{ij} - n_{ji}} \\ n_{ij} - n_{ji} \le 0, da_{ij} = 0 \end{cases}$$
(13)

where  $0 \le da_{ij}, ca_{ij} \le 1$ ;  $ca_{ij}$  indicates the control degree of compartment *j* on compartment *i* based on the controller's output environ;  $da_{ij}$  indicates the dependence degree of compartment *j* on *i* from the observer's input environ. In addition to N, the output-oriented integral matrix N' is also derived from a quantified CFN, wherein  $G' = (g'_{ij}), g'_{ij} = f_{ij}/T_i$ . The control allocation (CA) and dependence allocation (DA) are determined by the two pairwise integral flows N and N'. In addition, the formulation of component importance represented by centrality<sup>61-64</sup> is provided in Supporting Information.

219 The ascendancy (A) can quantify the network evolution and development built on the 220 inter-component flows, and has been widely applied in assessing the organization, efficiency, and sophistication of various systems.<sup>31,57</sup> Capacity (C) is often used to define the total volume 221 222 of information that a network contains based on its size and self-organized flow pattern. On this basis, the relative ascendancy, or the ratio of ascendancy to capacity ( $\alpha$ ) has been proposed.<sup>58</sup> 223 A higher A/C ratio value indicates a more developed, efficient, and organized system.<sup>24</sup> In this 224 study, the relative ascendancy represents the efficiency of carbon transfer and the 225 226 transformation amongst different components.

227 
$$A = TST_P^2 \sum_{i,j}^n \frac{f_{ij}}{TST_P} \log \frac{f_{ij}TST_P}{T_iT_j}$$
(14)

228 
$$C = -TST_P^2 \sum_{i,j}^n \frac{f_{ij}}{TST_P} \log \frac{f_{ij}}{TST_P}$$
(15)

 $229 \qquad \alpha = A/C \tag{16}$ 

where  $TST_P$  is the total system throughput of a city's entire CFN, and the sum of all carbon imports, inter-component flows, and exports; A, C, and  $\alpha$  denote the ascendancy, capacity, and relative ascendancy, respectively.

An ideal urban carbon network, arguably, should be both efficient in terms of the carbon exchanges amongst components and, at the same time, resilient against possible external disturbances (for example, lack of supply in certain carbon routes), which brings us to the fourth functional indicator called robustness (R). R measures the trade-off between efficiency and redundancy in a single metric. As the urban carbon networks move towards either extremes, i.e., overly efficient or overly redundant, the robustness of the carbon metabolic system falters.

$$R = -\alpha \log(\alpha) \tag{17}$$

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Network synergism<sup>65</sup> is an indicator extracted from utility analysis, and represents the ratio of the interaction effect between the benefit derived from net positive flows and the depression associated with net negative flows. In network utility analysis, the combination of element symbols in the integral utility matrix (U) can be used to determine the nature of interactions between two components, such as mutualism, competition, and so on. In this study, we focused on the ratio of positive conditions to negative conditions to obtain information on the health of and mutual benefit in an urban CFN.

247 
$$U = (I-D)^{-1}$$
 (18)

248 
$$D = (d_{ij}) = \frac{(f_{ij} - f_{ji})}{T_i}$$
(19)

249 
$$\frac{b}{c} = \frac{\sum +U}{\left|\sum -U\right|}$$
(20)

where U is the integral utility matrix with consideration to both the direct and indirect relative flow difference; D is the direct utility matrix with consideration only to the direct relative flow difference. The network synergism is calculated based on the ratio of the summing positiveintegral utilities to the summing negative integral utilities.

254 Based on these two categories of metrics, the correlations between system properties of 255 carbon metabolism and urban socioeconomic attributes are assessed. A set of widely-used 256 socioeconomic attributes that represent urban development are selected for correlation analysis, 257 including carbon emission (in total or per capita), population, population density, GDP (Gross 258 domestic products; in total or per capita), and urbanization rate. The significance of correlation 259 may indicate the degree of relevance of the network metrics to current urban socioeconomic 260 management and whether there is a synergy between carbon emission mitigation and urban 261 metabolism optimization.

#### 262 2.5 Case study and data

263 Eight global cities (at similar time point) were selected for case study: Vienna (2005), 264 Sydney (2008), Sao Paulo (2009), Los Angeles (2008), London (2005), Hong Kong (2006), Cape Town (2006) and Beijing (2008). The geographical and socioeconomic situation of these 265 eight cities are presented in Table S1. The selection of cities mainly because: (1) these cities 266 267 cover all major populated regions (North America, South America, Europe, Asia, Oceania, and 268 Africa), are currently in different development stages, and have sufficient geographical and 269 socioeconomical diversity to test the generic pattern of urban carbon flows; (2) they have 270 relatively reliable city-level energy and material data, which is a requirement for developing 271 valid carbon flow models. A detailed data description for the urban CFNs is provided in Table 272 S2, accompanied with the major sources of energy and material flow data for the eight cities.

- 273 3. Results and discussion
- 274 **3.1 Carbon flow networks of cities**

Figure 2 shows the inter-component flows in the carbon networks of the eight cities. The width of the ribbons indicates the amount of carbon exchanged between two urban components. The ribbons are colored according to the direction of the component from which the carbon is exported. However, because the total output is equal to the total input in the balanced networks, the number along each belt refers to the total carbon throughflow of each component. Note that these flows refer to direct flows controlled by urban components, that is, a direct exchange of carbon through trade or other linkages. We also show the per capita carbon throughflows of 8 global cities by component in Figure S1 in Supporting Information.

283 We find that import- and exported-related carbon flows contribute up to 70 percent of the 284 cities' total system throughflows. This indicates that urban carbon networks are highly open 285 systems in the sense that they rely on the external environment through frequent imports of 286 carbon as raw materials for manufacturing or household products for domestic consumption 287 (and therefore export the carbon emissions to the atmosphere following use of the imports). The 288 four dominant components inside the carbon networks of cities are Emission, Electricity, gas 289 and water, Construction and Services, although the component contribution to the carbon flows 290 is notably different for each city. Ems contributes the most to the total carbon throughflows in 291 most urban carbon networks (from 12% in Sao Paulo to 19% in Sydney). Two infrastructure-292 related economic sectors, namely Electricity, gas and water and Construction, play an important 293 part in directing the carbon exchanges in the cities, and are responsible for 10% and 9% of the 294 total carbon throughflow on average, respectively. The carbon emissions originating from 295 power generation have been shown to be a significant source of urban carbon flows for cities 296 in either developed or developing countries, although most flows originate outside of urban 297 boundaries. For example, in Vienna, flows from Electricity, Domestic consumption and 298 Transportation to Emission are the major carbon emission pathways. The carbon throughflow 299 of Construction is more diverse among cities. The construction activities in cities of developing 300 countries, such as Beijing and Sao Paulo, can contribute up to 12% of the total carbon 301 throughflow, while for cities in developed countries, such as Los Angeles and London, this 302 proportion is only 8%. This is mainly attributed to the higher demand of building materials 303 (wood, cement, and so on) during fast urbanization in developing countries. Interestingly, Stock

304 variation is a significant component for carbon networks in many cities, whose throughflow 305 accounting for 8% of the total carbon throughflow on average. In Sao Paulo, the amount of 306 carbon that ends up in stock (6229 kt C, 13% of total system throughflow) is higher than that 307 becoming emission (5756 kt C). These components are associated with the biggest carbon flows 308 in the cities. In Sydney, flows from Electricity and Transportation to Emission and from 309 Construction to Stock variation are significant in the network, and the same occurs to Los 310 Angeles, London, and Cape Town. For Sao Paulo, the pairs of Construction→Stock variation 311 and Domestic consumption  $\rightarrow$  Stock variation account for a large proportion of carbon 312 throughflows. The network analysis reveals important evidences proving that in addition to 313 gaseous emission, the change in urban stock may also have a significant impact on the whole 314 carbon networks of cities.

315 It is widely recognized that anthropogenic gaseous emissions play a major role in the carbon cycles of natural-human complex systems such as cities.<sup>19,20</sup> From an urban metabolism 316 317 perspective, our study demonstrated that approximately one-fifth of the total carbon 318 throughflow is directly associated with carbon emissions into the atmosphere. Additionally, the 319 significant flows to carbon stock raise concerns with regard to potential future emissions, 320 although they are not currently considered as accounting for part of the emissions. The 321 inventory of all inter-component carbon throughflows can offer a broader view of the size and 322 structure of urban carbon metabolism compared with carbon emissions accounting and provide 323 a basis for further carbon network modelling.

Figure 2 Carbon flow networks of 8 global cities

Notes: The number along each belt refers to the carbon throughflow (in kiloton of C) of each component
in the cities, while the percentages are their contributions to TST. Agr: Agriculture, forestry and
horticulture; Min: Mining; Man: Manufacture; Ele: Electricity, gas and water; Con: Construction; Tra:
Transportation; Ser: Services; Dom: Domestic consumption; Gov: Governmental consumption; Ems:
gaseous emission; Dwa: Biodegradable waste; Nwa: Non-biodegradable waste; Sto: Stock variation;
Row: rest of the world. The figure is powered by Circos Table Viewer.

# 332 **3.2 Performances and patterns at system and component levels**

333 Figure 3a shows the correlations between carbon flows and urban socioeconomics in total values. 334 We found that three flow-based metrics, namely, the total system throughflow, boundary flow, and 335 cycled flow, are highly correlated with the magnitudes of carbon emissions from cities. These flow metrics do not represent the carbon footprint of the urban economy (e.g. <sup>12,15</sup>); instead, they act as 336 337 the carbon "metabolic intensity" and are affected by all carbon-related processes. Nonetheless, these 338 flow-based metrics are closely related with carbon emissions in at least two ways: 1) the carbon 339 emissions from various urban components are a significant part of total system throughflow, can 340 contribute to the cycled flow when entering cycled chains, and subsequently become a fraction of 341 the boundary flow; 2) more gaseous emissions often means higher consumption of energy or 342 frequent industrial activity, which in turn attracts carbon inflow to a city as fuels, construction 343 materials, and other products. More importantly, the deviation of these indicators from carbon 344 emissions is meaningful. These metrics can provide useful information on a city's total metabolism, 345 boundary metabolism, and cycled metabolism, which cannot be obtained by direct carbon 346 accounting. The cycled flow is also a good measure for the degree of circularity in the economy, a 347 concept that gaining traction as a way to both strengthen the economy and lower emissions.

348 The total system throughflow, boundary flow, and cycled flow have strong positively linear 349 correlations with the population. This indicates that the impact of the urban population on carbon 350 metabolism is unlikely to slow down as more people swarm into the city. An exception is Sao Paulo 351 with a population of 11.4 million, whose total system throughflow, boundary flow, and cycled flow 352 are lower than the values predicted by the regression model. Sao Paulo has a relatively low-carbon 353 economy from a carbon metabolic flow perspective. In contrast, Sydney and Los Angeles have a 354 higher level of carbon flow compared with the predicted level, which indicates a relatively high-355 carbon city profile given their carbon emissions related to transportation. Here, the correlations of 356 GDP with the total system throughflow and boundary flow are much weaker. The expansion of the 357 economic scale does not have a definitive impact on the urban carbon metabolism. Many other 358 factors may also be equal or more important, such as technology, scale of export, and so on. 359 Essentially, there is no significant correlation between GDP and cycled flow, in the sense that cycled 360 chains are often more related to the economic structure and compactness of urban industries and 361 services. This implies that, for cities with a larger amount of emission from transportation sector 362 like Los Angeles, the carbon metabolism is more intensive than other study cities.

363 Figure 3b shows the correlation between the carbon metabolic function (Finn cycling index, 364 synergism, system robustness) and the urban socioeconomics of the eight cities. The Finn 365 cycling index for the eight cities ranges from 0.06 to 0.10, which indicates that less than 10%366 of the carbon is cycled within the urban metabolic network. We found that Finn cycling index 367 has a positive correlation with the population density. This suggests that it is possible for the 368 recycling of carbon products in cities to increase with a denser urban form and more compact industrial network. In contrast, Finn cycling index has a negative correlation with GDP per 369 370 capita and carbon emissions per capita, although this correlation is loose. Higher GDP may 371 result in larger total system throughflow, but the increase of financial income typically 372 augments gaseous emissions that are not cycled back into the urban economy. This is an 373 important side effect caused by urban economic development. Therefore, a goal to increase the 374 service economy, because on face it might have lower direct emissions, might lead to higher 375 overall emissions both because GDP per capita increases and lower cycling. Similarly, the 376 network synergism of cities is positively correlated with the population density, and negatively 377 correlated with the carbon emissions per capita. Cities with a higher population density and 378 lower per capita carbon emissions also have a healthier carbon metabolic system with better 379 cooperation amongst components. This demonstrates that the objectives of urban 380 decarbonization and carbon metabolism optimization can be simultaneously achieved in a 381 systemic urban carbon management framework. The cities' relative ascendancy ( $\alpha$ ) ranges from

382 0.22 to 0.25, resulting in variations in the system robustness between cities (with up to 5%) 383 difference between Sydney and Vienna). The average system robustness of carbon networks is 384 0.34 of the study cities, which falls in the middle of natural ecosystems and artificial or 385 economic trade systems (as shown by Figure S3 in Supporting Information). This is mainly 386 because the carbon flow networks of cities represent the interface between natural processes 387 (such as waste decomposition and carbon sequestration by urban trees) and socioeconomic 388 activities (such as energy-related emission, carbon exchange in products and food consumption) 389 and can be influenced by both natural and human-dominated components in cities. We find that 390 network metrics used such as system robustness and synergy does not have a significant linear 391 correlation with either per capita GDP, population density, or per capita carbon emission.

392 However, it is important to note that the complexity and evolution of urban metabolic 393 networks are not fully determined by social and economic conditions. This indicates that the 394 metabolic properties of carbon networks may not evolve in the same pace as urban 395 socioeconomic development, and the interpretive network metrics usually used in biological or 396 ecological systems, could not be interpreted in the same way as other mass-based network 397 indicators. In particular, system robustness has a clear linkage to ecosystem function (e.g. 398 biodiversity and abundance) and implications that may hard to find parallel in socioeconomic 399 systems like urban economy. Thus, caution should be used when applying system robustness 400 to assess urban carbon metabolism or possibly other urban systems that are driven by 401 socioeconomic activities rather than ecological processes.

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Figure 3 Correlations (a) between carbon metabolic flows and urban socioeconomics, and (b)
between carbon metabolic function and urban socioeconomics. Extended regression results
are provided in Table S3 and Table S4 in Supporting Information.

Based on a network metrics called centrality, we assessed the role each component plays for the eight cities with consideration to both direct and indirect flows (Figure S2). Generally, Agriculture, Mining, Transportation, and Biodegradable waste are more important from an input perspective, in the sense that their input environ centralities are notably higher than their 412 output environ centralities. This phenomenon is common for all study cities and may indicate 413 the similarity of carbon metabolizing behaviors in specific components regardless of the city's 414 development stage. However, Emission component, Stock variation, and Non-biodegradable waste are dominant ways of exporting carbon with relatively high output environ centralities. 415 416 The Emission component is more important in some cities such as Sydney, Los Angeles, and 417 London, while Stock variation seems to contribute more in Sao Paulo. This confirms our 418 conjecture that Gaseous emission and Urban stock are two significant destinations of the 419 integral carbon flow. Emission, Construction, and Stock variation have the largest throughflow 420 centrality in Vienna, Sydney, and Sao Paulo, while in Los Angeles, London, Hong Kong, and 421 Cape Town, Stock variation and Service are dominant in addition to Emission. A further 422 evaluation of these two components is needed to clarify how they are controlled by other 423 components and what is the most efficient approach toward regulating them.

424 The control and dependence relationships between components in urban carbon networks 425 reveal the potential mechanism of efficient carbon management (Figure 4). By considering all 426 direct and indirect interactions, we found that the control allocation (CA) amongst the 427 components are diverse and uneven. The silence (low degree of control relations) amongst the 428 economic sectors does not mean that there are no interactions, but rather that the interactions 429 between the urban economy and the environmental distributions are more intensive from a 430 network control perspective. There are significant differences between the control regimes of 431 the eight global cities. However, various generic patterns can be derived. By targeting the 432 dominant components and key processes in the carbon networks, it is possible to obtain efficient 433 carbon mitigation pathways in cities. Many urban economic sectors such as Electricity, gas and 434 water, Transportation, and Construction have a strong control over Emission and Stock 435 variation. In some cases, more control in these economic sectors is exerted over Emission. For 436 example, in Sydney, 32% of Electricity, gas and water control in carbon exchange is allocated 437 to Emission, which is much higher than that received by Stock variation. However, in Sao Paulo, 438 more than 70% of the control in the economic sectors is allocated to Stock variation, owing to

the dominance of stock in the carbon network (as revealed by previous results). The economic sectors also show some extent of control over Domestic consumption, while in most cities, Domestic consumption has significant control over Emission and Stock variation, but no control over the industry and Service. For example, 35% and 51% of Domestic consumption control goes to Emission and Stock variation in Vienna, respectively, while the proportions in Sydney are 26% and 45%, respectively.

445 The dependence allocation (DA) shows the inter-component control relationships in the 446 carbon networks from a receiver's viewpoint. Manufacturing, Service, and Domestic 447 consumption depend on many other components in the urban economy to derive carbon. For 448 example, the Manufacturing in Sao Paulo depends on Construction, Mining, and Service by 449 65%, 6%, and 5%, respectively. Service is greatly dependent on Electricity, gas and water and 450 Construction, with a total dependence degree of 81–91%. In London, Domestic consumption is 451 dependent on Electricity, gas and water, Construction, Transportation, and Service by 13%, 452 20%, 11%, and 44%, respectively. However, this varies from city to city. For example, Hong 453 Kong's Domestic consumption is dependent on these components by 6%, 35%, 12%, and 31%, 454 respectively. In London, the dependence of domestic activities on Service is up to 44% owing 455 to the major role of the commercial activities in the city's economy. We can see that Emission 456 depend on a range of urban economic sectors, namely Electricity, gas and water, Construction, 457 Transportation, and Service. In our sample of cities, the Emission is controlled by these 458 economic sectors by 16–22%, 7–12%, 9–16%, 8–15%, and 6–18%, respectively. Additionally, 459 domestic activities also have a considerable impact on Emission with a dependence degree of 460 6-10% amongst the cities. We also found that Stock variation is very dependent on 461 Construction, Service, and Domestic consumption in the sense that these three components are 462 amongst the major sources of carbon stored in the urban economy. It is essential to have a clear 463 understanding of the full carbon flow chains before they end up in emissions, including both direct and indirect pathway between sectors.<sup>25,34</sup> These results can provide a systemic 464 465 perspective on how carbon emissions are controlled by urban economic sectors through



466 tracking each pair-wise network relationship and how it ripples through the urban network.

Figure 4 Control allocation (CA) and dependence allocation (DA) between urban components 468 in exchanging carbon flows. Note: CA should be read from row to column, i.e. the control of 469 470 column components over the row components; DA should also be read from row to 471 column, but the meanings change to the dependence of column components on the row 472 components. Both CA and DA are within the range between 0 and 1, with larger number 473 representing higher control or dependence over others. The highlighted areas are the 474 control of key urban economic sectors over others from the CA perspective, and the 475 dependence of carbon emission and stock on other sectors from the DA perspective.

476 Cities must address climate change in every possible way.<sup>36</sup> Yet, there is a big knowledge

477 gap between the inventory of carbon emissions/sinks and the modelling of carbon flows within 478 the context of urban metabolism. A number of methodological frameworks and guidelines have 479 been proposed for city-level carbon emission inventory.<sup>12,18,66-68</sup> The analyses centered around 480 emission dominates current discussion on urban decarbonization, while other non-emission 481 carbon being exchanged in urban economy are largely disregarded. This is mainly because how 482 these flows will end up in emission is not clearly understood and there is no sufficient and 483 reliable data to do just so.

484 On the other hand, there has been increasing interest in using a nature-based method to alleviate the environmental burden carried by urban development.<sup>69,70</sup> The concept of 485 metabolism fits right into this research initiative. Urban metabolism has been developed as a 486 487 methodological framework for investigating various energy and material flows associated with urban growth.<sup>71</sup> The human impact on urban carbon metabolism is highly complex and 488 489 interrelated with various natural and economic components. To better understand this impact, 490 all carbon imports, exports and inter-component exchanges comprising the network should be examined. Current inventory-based approaches are mainly directed to an intensity-based 491 492 analysis, and yet the structural and functional aspects can be better understood by modelling 493 techniques that include indirect effects. The integration of intensity, structural and functional 494 information of carbon metabolism is needed to fill in current knowledge gap and provide a broader understanding of cities' impact on climate change.<sup>10,18,33,72</sup> 495

496 The categorization of carbon flow metrics in this study (flow-based metrics and 497 interpretative network metrics) can provide a promising scheme for linking carbon emission 498 inventories to metabolic modelling efforts. Flow-based metrics are grounded on conservation 499 of mass and so these results can be directly used in making carbon management policies, which 500 are not very different from indicators like carbon emissions and sinks for this matter. In contrast, 501 interpretative network metrics can be used to understand the mechanism of network functioning 502 or inter-component relations that cannot be shown by flow-based metrics. They are appropriate 503 for interpreting the carbon-related interactions in the urban metabolic system and for comparing

504 the system performances of carbon metabolic networks among cities, though whether they can 505 be employed in the regulation of sectoral activities and behaviors needs further inspection. This 506 merit is well-reflected by the control analysis it provides. Regarding carbon analyses, one can 507 account for direct carbon emissions from all economic sectors and households using material 508 flow analysis and embodied carbon emission from input-output analysis, but how the carbon 509 emission of a sector is controlled by activities of a bunch of other sectors and how these 510 activities are further induced by other activities could be only be fully addressed using network 511 analysis. Another merit is that it can provide potential goal functions for system evolution and optimization such as maximum ascendency, maximum cycling, among others.<sup>73</sup> Some of these 512 513 goal functions are showing potential of applications in socioeconomic systems as well, such as 514 the information-based carbon modelling in.<sup>74</sup> Our results indicate that they align favorably with 515 other common standard metrics already in place, but give a richer insight into how the network 516 patterns lead to these outcomes. Within the framework of urban metabolism, these two 517 categories of metrics can be combined to quantify the influence of urbanization and economic transition on carbon network connectivity and diversity,<sup>31,74</sup> and foster system-oriented 518 519 strategies for urban carbon reduction that supplement current mitigation actions.

520 Globally, comparative studies among cities are called for to disentangle the interactions 521 of human activities and to find strategies and roadmap to mitigate climate change.<sup>36,75,76</sup> The 522 analyses of eight global cities suggest there is a large difference in city-level carbon balance, 523 but a similarity in inter-component relationships and general metabolic characteristics can be 524 found. An integration of flow-based metrics, interpretative network metrics and other 525 socioeconomic models will convey important information about how future carbon flows 526 should be managed according to the urban economic and demographic changes. A major 527 limitation to the comparative results is the relatively small sample used (eight cities). Provided 528 the metabolic data are more available and frequently updated at a city level, a global 529 comparation with a large sample may renew our current understanding. Still, the current eight-530 city study is able to demonstrate how the carbon metabolic patterns can be identified and 531 compared among different cities. Policy makers can acquire the carbon metabolism knowledge 532 from other cities to help them select their own strategies and countermeasures and guide cities 533 toward more rational and concerted climate actions. In turn, this will increase the importance 534 of determining the key metabolic characteristics of different cities and using them as a reference 535 during the adaptation of available mitigation techniques.

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#### References

- Grimm, N. B.; Faeth, S. H.; Golubiewski, N. E.; et al. Global change and the ecology of cities. *Science* 2008, 319(5864), 756–760.
- [2] Gamba, P.; Herold, M. Global mapping of human settlement: experiences, datasets and prospect; CRC Press, Boca Raton, 2009.
- [3] Seto, K. C.; Dhakal, S.; Bigio, A.; et al. *Human Settlements, Infrastructure and Spatial Planning. In: Climate Change 2014: Mitigation of Climate Change*. Contribution of Working Group III to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change [Edenhofer, O., R. et al. (eds.)]. Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA, 2014.
- [4] Seto, K. C.; Güneralp, B.; Hutyra, L. R. Global forecasts of urban expansion to 2030 and direct impacts on biodiversity and carbon pools. *Proceedings of the National Academy of Sciences* 2012, 109, 16083-16088.
- [5] American Association for the Advancement of Science, Rise of the city. *Science* 2016, 352(6288), 906–907.
- [6] United Nations Climate Change. The Paris Agreement. Available at: https://unfccc.int/process-

and-meetings/the-paris-agreement/the-paris-agreement.

- [7] UN-Habitat; UNESCO; World Health Organization; UNISDR; UN Women; UNEP and UNDP. SDG Goal 11 Monitoring Framework. Accessed at <u>http://unhabitat.org/sdg-goal-11-monitoring-framework/</u>; 2016.
- [8] Wigginton, N. S.; Fahrenkamp-Uppenbrink, J.; Wible, B.; et al. Cities are the future. *Science* 2016, 352 (6288), 904-905.
- [9] Reckien, D.; Creutzig, F.; Fernandez, B.; et al. Climate change, equity and the Sustainable Development Goals: an urban perspective. *Environment and urbanization* 2017, 29(1), 159-182.
- [10] Creutzig, F.; Baiocchi, G.; Bierkandt, R.; et al. Global typology of urban energy use and potentials for an urbanization mitigation wedge. *Proceedings of the National Academy of Sciences* 2015, 112 (20), 6283-6288.
- [11] Chen, S. Q.; Chen, B. Coupling of carbon and energy flows in cities: A meta-analysis and nexus modelling. *Applied energy* 2017, 194, 774-783.
- [12] Chavez, A.; Ramaswami, A. Articulating a trans-boundary infrastructure supply chain greenhouse gas emission footprint for cities: Mathematical relationships and policy relevance. *Energy Policy* 2013, 54, 76-384.
- [13] Ramaswami, A.; Tong, K.; Fang, A.; et al. Urban cross-sector actions for carbon mitigation with local health co-benefits in China. *Nature Climate Change* **2017**, 7(10), 736.
- [14] Liu, Z.; Feng, K.; Hubacek, K.; et al. Four system boundaries for carbon accounts. *Ecological modelling* 2015, 318, 118-125.
- [15] Chen, S. Q.; Liu, Z.; Chen, B.; et al. Dynamic Carbon Emission Linkages Across Boundaries. *Earth's Future* 2019,7(2), 197–209.
- [16] Lin, J.; Hu, Y.; Cui, S.; et al. Tracking urban carbon footprints from production and consumption perspectives. *Environmental Research Letters* 2015, 10(5), 054001.
- [17] Ramaswami, A.; Hillman, T.; Janson, B.; et al. A Demand-Centered, Hybrid Life-Cycle Methodology for City-Scale Greenhouse Gas Inventories. *Environmental Science & Technology* 2008, 42(17), 6455-6461.
- [18] Kennedy, C.; Steinberger, J.; Gasson, B.; et al, Greenhouse gas emissions from global cities. *Environmental Science & Technology* 2009, 43, 7297-7302.
- [19] Pataki, D. E.; Alig, R. J.; Fung, A. S.; et al. Urban ecosystems and the North American carbon cycle. *Global Change Biology* **2006**, 12, 2092-2102.
- [20] Churkina, G.; Brown, D. G.; Keoleian, G. Carbon stored in human settlements: the conterminous United States. *Global Change Biology* 2010, 16, 135-143.
- [21] Churkina, G. Modeling the carbon cycle of urban systems. *Ecological Modelling* **2008**, 216, 107-113.
- [22] Peters, G. P.; Davis, S. J.; Andrew, R. A synthesis of carbon in international trade.

Biogeosciences 2012, 9, 3247-3276.

- [23] Schramski, J. R.; Dell, A. I.; Grady, J. M.; et al. Metabolic theory predicts whole-ecosystem properties. Proceedings of the National Academy of Sciences 2015, 112(8), 2617-2622.
- [24] Pizzol, M.; Scotti, M.; Thomsen, M. Network Analysis as a tool for assessing environmental sustainability: Applying the ecosystem perspective to a Danish Water Management System. *Journal of environmental management* 2013, 118, 21-31.
- [25] Chen, S.; Chen, B. Network environ perspective for urban metabolism and carbon emissions: A case study of Vienna, Austria. *Environmental Science & Technology* 2012, 46(8), 4498–4506.
- [26] Schramski, J. R.; Gattie, D. K.; Patten, B. C.; et al. Indirect effects and distributed control in ecosystems: Distributed control in the environ networks of a seven-compartment model of nitrogen flow in the Neuse River Estuary, USA—Steady-state analysis. *Ecological Modelling* 2006, 194(1), 189-201.
- [27] Hines, D. E.; Singh, P.; Borrett, S. R. Evaluating control of nutrient flow in an estuarine nitrogen cycle through comparative network analysis. *Ecological Engineering* 2016, 89, 70-79.
- [28] Rakshit, N.; Banerjee, A.; Mukherjee, J.; et al. Comparative study of food webs from two different time periods of Hooghly Matla estuarine system, India through network analysis. *Ecological Modelling* 2017, 356, 25-37.
- [29] Layton, A.; Bras, B.; Weissburg, M. Designing industrial networks using ecological food web metrics. *Environmental science & technology* 2016, 50(20), 11243-11252.
- [30] Borrett, S. R; Sheble, L.; Moody, J.; et al. Bibliometric review of ecological network analysis: 2010–2016. *Ecological Modelling* 2018, 382, 63-82.
- [31] Fath, B. D. Quantifying economic and ecological sustainability. *Ocean & Coastal Management* **2015**, 108, 13-19.
- [32] Lu, Y.; Chen, B.; Feng, K.; et al. Ecological network analysis for carbon metabolism of ecoindustrial parks: a case study of a typical eco-industrial park in Beijing. *Environmental Science* & Technology 2015, 49(12), 7254-7264.
- [33] Xia, C.; Li, Y.; Xu, T.; et al. Quantifying the spatial patterns of urban carbon metabolism: A case study of Hangzhou, China. *Ecological Indicators* 2018, 95, 474-484.
- [34] Chen, S. Q.; Chen, B. Tracking inter-regional carbon flows: a hybrid network model. *Environmental Science & Technology* **2016**, 50 (9), 4731–4741.
- [35] Xia, L.; Zhang, Y.; Wu, Q.; et al. Analysis of the ecological relationships of urban carbon metabolism based on the eight nodes spatial network model. *Journal of Cleaner Production* 2017, 140, 1644-1651.
- [36] Bai, X.; Dawson, R. J.; Ürge-Vorsatz, D.; et al. Six research priorities for cities and climate change. *Nature* 2018, 555, 23-25.
- [37] Falkowski, P.; Scholes, R. J.; Boyle, E.; et al. The global carbon cycle: a test of our knowledge of earth as a system. *Science* **2000**, 290(5490), 291-296.

- [38] Luo, Y.; Keenan, T. F.; Smith, M. Predictability of the terrestrial carbon cycle. *Global change biology* 2015, 21(5), 1737-1751.
- [39] Chen, S.; Xu, B.; Chen, B. Unfolding the interplay between carbon flows and socioeconomic development in a city: what can network analysis offer? *Applied Energy* **2018**, 211, 403–412.
- [40] Piña, W. H. A.; Martínez, C. I. P. Urban material flow analysis: An approach for Bogotá, Colombia. *Ecol Indicat* 2013. http://dx.doi.org/10.1016/j.ecolind.2013.10.035
- [41] Bai, X. Industrial ecology and the global impacts of cities. *Journal of Industrial Ecology* 2007, 11(2), 1-6.
- [42] Hao, Y.; Su, M.; Zhang, L.; et al. Integrated accounting of urban carbon cycle in Guangyuan, a mountainous city of China: the impacts of earthquake and reconstruction. *Journal of Cleaner Production* 2015, 103, 231-240.
- [43] DAFF. Australian State of the Forests Report: Five yearly report 2008. Australian Bureau of Agricultural and Resource Economics and Sciences. Department of Agriculture, Forestry and Fisheries, 2008.
- [44] Lamlom, S. H.; Savidge, R. A. A reassessment of carbon content in wood: variation within and between 41 North American species. *Biomass and Bioenergy* 2003, 25, 381-388.
- [45] Stockmann, K. D.; Anderson, N. M.; Skog, K. E.; et al. Estimates of carbon stored in harvested wood products from the United States forest service northern region, 1906-2010. *Carbon Balance and Management* 2012, 7, 1.
- [46] Moriarty, D. J.; Barclay, M. C. Carbon and Nitrogen Content of Food and the Assimilation Efficiencies of Penaeid Prawns in the Gulf of Carpentaria. Aust. J. Mar. *Freshwater Res* 1981, 32, 245-251.
- [47] Aguilera, J. A.; Aragon, C.; Campos, J. Determination of carbon content in steel using laserinduced breakdown spectroscopy. *Applied Spectroscopy* 1992, 46, 1382-1387.
- [48] IPCC, Revised 1996 Guidelines for National Greenhouse Gas Inventories: Workbook (1997). <u>http://www.ipcc-nggip.iges.or.jp/public/gl/invs1.html</u>, 1997.
- [49] Fath, B. D.; Scharler, U. M.; Ulanowicz, R. E.; et al. Ecological network analysis: network construction. *Ecological Modelling* 2007, 208(1), 49-55.
- [50] Fath, B. D.; Patten, B. C. Review of the foundations of network environ analysis. *Ecosystems* **1999**, 2(2), 167-179.
- [51] Ulanowicz, R. E. Mass and energy flow in closed ecosystems. *Journal of Theoretical Biology* 1972, 34(2), 239-253.
- [52] Fath, B. D.; Borrett, S. R. A Matlab® function for network environ analysis. *Environmental Modelling & Software* 2006, 21(3), 375-405.
- [53] Kazanci, C. EcoNet, a new software for ecological model simulation and network analysis. *Ecological Modelling* 2007, 208 (1), 3-8.
- [54] Schramski, J. R.; Kazanci, C.; Tollner, E. W. Network environ theory, simulation, and EcoNet®

2.0. Environmental Modelling & Software 2011, 26(4), 419-428.

- [55] Borrett, S. R.; Lau, M. K. enaR: an R package for ecosystem network analysis. *Methods in Ecology and Evolution* 2014, 5(11), 1206-1213.
- [56] Finn, J. T. Measures of ecosystem structure and function derived from analysis of flows. J. Theor. Biol. 1976, 56, 363-80.
- [57] Ulanowicz, R. E.; Norden, J. S. Symmetrical overhead in flow networks. Int J Syst Sci 1990, 21(2), 429-37.
- [58] Ulanowicz, R. E.; Goerner, S. J.; Lietaer, B.; et al. Quantifying sustainability: resilience, efficiency and the return of information theory. *Ecological Complexity* **2009**, 6 (1), 27–36.
- [59] Fath, B. D. Distributed control in ecological networks. Ecol Model 2004, 179, 235-46.
- [60] Chen, S. Q.; Fath, B. D.; Chen, B. Information-based Network Environ Analysis: A system perspective for ecological risk assessment. *Ecol Indic* **2011**, 11 (6), 1664–1672.
- [61] Borgatti, S. P. Centrality and network flow. Soc Networks 2005, 27, 55–71.
- [62] Wasserman, S.; Faust, K. Social Network Analysis: Methods and Applications. Cambridge University Press, Cambridge, New York, 1994.
- [63] Borrett, S. R. Throughflow centrality is a global indicator of the functional importance of species in ecosystems. *Ecological Indicators* 2013, 32, 182-196.
- [64] Fann, S. L.; Borrett, S. R. Environ centrality reveals the tendency of indirect effects to homogenize the functional importance of species in ecosystems. (Report). *Journal of Theoretical Biology* 2012, 294, 74-86.
- [65] Fath, B. D.; Patten, B. C. Network synergism: emergence of positive relations in ecological systems. *Ecological Modelling* **1998**, 107(2-3), 127-143.
- [66] Kennedy, A.; Ibrahim, N.; Hoornweg, D. Low-carbon infrastructure strategies for cities. *Nature Climate Change* 2014, 4, 343-346.
- [67] WRI and WBCSD. The Greenhouse Gas Protocol: A Corporate Accounting and Reporting Standard, World Resources Institute and World Business Council for Sustainable Development: Washington, DC. <u>http://www.ghgprotocol.org/sites/default/files/ghgp/standards/ghg-protocolrevised.pdf</u>, 2004.
- [68] ICLEI, WRI and C40. *Global Protocol for Community-Scale GHG Emissions*. http://www.iclei.org/activities/agendas/low-carbon-city/gpc.html, 2014.
- [69] Collins, J. P.; Kinzig, A.; Grimm, N. B.; et al. A New Urban Ecology Modeling human communities as integral parts of ecosystems poses special problems for the development and testing of ecological theory. *American Scientist* 2000, 88(5), 416-425.
- [70] Kabisch, N.; Frantzeskaki, N.; Pauleit, S.; et al. Nature-based solutions to climate change mitigation and adaptation in urban areas: perspectives on indicators, knowledge gaps, barriers, and opportunities for action. *Ecology and Society* 2016, 21(2), 39.
- [71] Kennedy, C.; Cuddihy, J.; Engel-Yan, J. The changing metabolism of cities. Journal of

Industrial Ecology 2007, 11(2), 43-59.

- [72] Chen, S.; Chen, B.; Feng, K.; et al. Physical and virtual carbon metabolism of global cities. *Nature Communications* **2020**, 11, 182.
- [73] Fath, B. D.; Patten, B. C.; Choi, J. S. Complementarity of ecological goal functions. *Journal of theoretical biology* 2001 208(4), 493-506.
- [74] Fang, D.; Chen, B. Information-based ecological network analysis for carbon emissions. *Applied Energy* **2019**, 238, 45-53.
- [75] Bai, X. Eight energy and material flow characteristics of urban ecosystems. *Ambio.* 2016, 45(7), 819-830.
- [76] C40 Cities Climate Leadership Group. https://www.c40.org/.