

A Systems Approach to Measure Trade Dependencies in U.S. Food-Energy-Water Nexus

Nemi Vora^{1,2}, Brian D. Fath^{2,3*}, and Vikas Khanna^{1*}

¹ Department of Civil and Environmental Engineering, University of Pittsburgh, 3700 O'Hara Street, 742 Benedum Hall, Pittsburgh, Pennsylvania 15261, United States

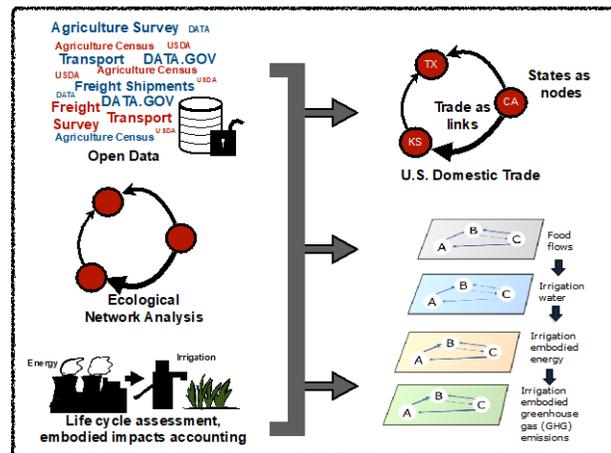
²Advanced Systems Analysis Program, International Institute for Applied Systems Analysis (IIASA), Schlossplatz 1, A-2361 Laxenburg, Austria.

³Department of Biological Sciences, Towson University, 8000 York Road, Towson, Maryland, 21252, United States

* Address correspondence to bfath@towson.edu and khannav@pitt.edu

ABSTRACT

We present a network model of the United States (U.S.) interstate food transfers to analyze the trade dependency with respect to participating regions and embodied irrigation impacts from a food-energy-water (FEW) nexus perspective. To this end, we utilize systems analysis methods including the



pointwise mutual information (PMI) measure to provide an indication of interdependencies by estimating probability of trade between states. PMI compares observed trade with a benchmark of what is statistically expected given the structure and flow in the network. This helps assess whether dependencies arising from empirically observed trade occur due to chance or preferential

25 attachment. The implications of PMI values are demonstrated by using Texas as an example, the
26 largest importer in the US grain transfer network. We find that strong dependencies exist not just
27 with states (Kansas, Oklahoma, Nebraska) providing high volume of transfer to Texas, but also
28 with states that have comparatively lower trade (New Mexico). This is due to New Mexico's
29 reliance on Texas as an important revenue source compared to its other connections. For Texas,
30 import interdependencies arise from geographical proximity to trade. As these states primarily rely
31 on the commonly shared High Plains aquifer for irrigation, over-reliance poses a risk for water
32 shortage for food supply in Texas. PMI values also indicate the capacity to trade more (the states
33 are less reliant on each other than expected), and therefore provide an indication of where the trade
34 could be shifted to avoid ground water scarcity. However, some of the identified states rely on
35 GHG emissions intensive fossil fuels such as diesel and gasoline for irrigation, highlighting a
36 potential tradeoff between crop water footprint and switching to lower emissions pumping fuels.

37

38 **KEYWORDS:** Food-Energy-Water nexus, food trade, irrigation, information theory, ecological
39 network analysis

40

41 **INTRODUCTION**

42 The United Nations General Assembly adopted the Sustainable Development Goals (SDGs) in
43 2015 to provide a roadmap for tackling seventeen distinct issues with the overarching theme of
44 human health and well-being, economic security, and environment sustainability. While diverse
45 in subjects, these goals are termed as an “indivisible whole”, and require managing for overlap in
46 policymaking to avoid suboptimal outcomes.¹ For instance, SDG 2 outlines ending hunger,
47 providing nutrition, achieving food security, and promoting sustainable agriculture. It directly ties

48 in with Goal 12 of sustainable production and consumption of resources, which in turn requires
49 planning for quality and plentiful supply of water (Goal 6), and renewable, affordable energy (Goal
50 7). As such, a single goal cannot be achieved in isolation while disregarding effects of others as it
51 may result in unintended consequences. Instead, a holistic systems perspective is required that
52 considers the complexity of interconnections. A crucial dilemma in applying a systems perspective
53 is to avoid falling into an abyss of an infinitely connected system. Therefore, an appropriate
54 boundary can help constrain the system and limit relevant interactions within and with the system.
55 The study of interactions within food, energy, and water resources, termed as food-energy-water
56 (FEW) nexus, can be seen as an example of drawing such a system boundary from many other
57 interwoven and equally important SDGs. Albeit, FEW nexus itself represents a complex web of
58 interconnections as energy and water are consumed across the entire food supply chain, energy is
59 needed for abstraction, treatment, and distribution of water, and a large amount of water is
60 consumed for power generation. Therefore, systems analysis needs to be complemented with a
61 context-specific study at specific geographic scales and sectors to understand effects of
62 interconnections. Recently, many such studies have adopted nexus approach to assess variety of
63 interactions at different spatial scales²⁻⁵ including wastewater management to simultaneously
64 reduce water-energy demand and boost nutrient cycling for London⁶, developing a scenario
65 analysis for competing water use in transboundary Brahmaputra River Basin⁷, impact of city-level
66 FEW nexus actions in Delhi⁸, and China's increasing environmental impacts due to focus on
67 international exports.⁹

68
69 The FEW nexus challenges associated with an agriculture-centric nation such as the United States
70 (U.S.) are different from developed countries that rely on agriculture imports or developing agro-

71 economies. For the U.S., one critical piece in understanding FEW nexus challenges is the energy
72 and greenhouse gas (GHG) emission burden of irrigated food production.¹⁰ Irrigation adds
73 significant value to food and feed production in the U.S.¹¹, providing a crucial link to study the
74 domestic FEW systems. Irrigation is the second largest freshwater withdrawal sector in the U.S.,¹²
75 while irrigation pumping accounts for substantial agricultural energy expenses.¹³ Additionally,
76 regional variation exists between agricultural resources availability and densely populated food
77 demand centers. For instance, the high plains in the U.S. is labeled the “breadbasket region” due
78 to significant grain production; and California provides a sizable portion of fruits, nuts, and
79 vegetables for domestic and international consumption. On the other hand, Illinois, Louisiana,
80 Texas, and Florida import a large amount of food due to their large population or geographically
81 strategic position as ports.¹⁴ As the imbalance between consumption and production increases,
82 understanding the patterns of trade dependencies becomes an important consideration for regional
83 food security.

84
85 Trading partner selections, and the subsequent dependencies, economic pressures, and
86 vulnerabilities of such preferences, have been discussed widely in the trade literature.¹⁵⁻¹⁷ Specific
87 to food trade, dependency is a complex issue as it may strengthen food security (through
88 diversifying trade partners) or harm food supply (reducing self-sufficiency). Prior work has
89 investigated dependencies arising from indirect resource use to produce traded food commodities
90 (referred to as virtual/embodied trade of resources).¹⁸⁻²⁰ Virtual resource trade (popularized by the
91 virtual water concept²¹) refers to the trade of resource that is not physically embedded but used in
92 producing the traded food commodity. Through virtual resource trade, regions can sustain greater
93 food demand than local production capacity by depending on external virtual water and land

94 imports to meet the demand.^{22, 23} Dependencies can also arise due to the structure and arrangement
95 of how trade links are formed. Prior work has investigated community patterns²⁴, central players^{10,}
96 ²⁵, robustness and resilience^{26, 27} and dynamics of the networks^{28, 29} by quantifying structural
97 properties of trade networks through graph theory based approaches. However, the dependencies
98 arising from interlinkages between food, energy and water resources and trading partners has been
99 understudied due to the complexity of the issue. Additionally, prior work addressing these issues
100 have focused on larger components and backbones,^{28, 30} central players¹⁰, and dominant flows in
101 the network.^{10, 31} However, little emphasis has been placed on examining weaker links and their
102 role in the network structure.

103 The importance of considering ties with weaker strength was outlined by Granovetter³² in
104 his essay on social networks. Granovetter noted that weak ties between individuals (i.e.,
105 acquaintances) are instrumental in maximum diffusion of information, mobility, and community
106 organization. **From a trade perspective, this translates to the fact that dependency exists in**
107 **both directions and weaker links may be important when all connections are considered.**
108 **Therefore**, we combine the resource and structure dependency narrative and examine the
109 importance of weak ties in the network. Specifically, we analyze the pattern of regional food trade
110 dependencies in the U.S. food trade. Here, a dependency denotes level of preferential attachment
111 (structural dependency) and reliance on resources (embodied resource dependency). We do this
112 by comparing observed trade to a null model of trade. The null model represents the most probable
113 trade given each state's import needs and export supply with no other specific preference in how
114 links are formed.³³ The emergent patterns in actual trade, not observed in the null model, provides
115 insights on dependence (level of preferential attachment) in the network. Additionally, we extend
116 the analysis to quantify virtual water (accounting for only irrigation), irrigation-related embodied

117 energy (referred to as embodied energy in the manuscript), and irrigation energy-related embodied
118 GHG emissions (referred to as embodied GHG emissions) to assess a state's [indirect dependency](#)
119 [on resources through trade](#). While trade typically refers to international exchanges, we limit the
120 analysis and discussion to the U.S. and refer to interstate trade as transfers.²⁵

121
122 Specifically, we leverage empirical data and compare existing patterns of domestic transfers with
123 calculated probabilities of association between participating states. To this end, we create four
124 distinct networks: 1) interstate physical food flows (US tons), 2) virtual water (m³),
125 3) embodied energy (MJ), and 4) embodied GHG emissions (kg CO₂ equivalent). Building on the
126 framework for the network analysis of physical food trade and embodied impacts first presented
127 in our previous work¹⁰, we limit the focus of the present study to grain and feed crop transfers with
128 states representing nodes in the network and volume of transfers and embodied environmental
129 impacts represented by links (edges) between nodes. In this study, we assess how much more often
130 than chance do two events occur together.³⁴ This is valuable information to gain for an extremely
131 well-connected network such as the U.S. domestic trade. Our previous analysis noted that on
132 average a state is connected to 36 other states out of 51 states.¹⁰ Therefore, if a state produces a
133 specific crop, unlike international trade, it is not restricted to trade with a particular state (no
134 political conflicts, trade agreements etc.).¹⁴ Therefore, by comparing observed trade connections
135 (empirical network) to those that may occur by chance (null model), we highlight the presence of
136 preferential attachment. Instead of purely empirical analysis, this provides statistical support to
137 understand significance of what we are observing and provides valuable contribution to the
138 literature. The rest of the article is organized as follows: material and methods section discusses
139 the data behind constructing four networks and introduces the PMI measure. Result and discussion

140 section applies the PMI measure to the system under study and discusses insights with the case of
141 Texas as an example. Details regarding the PMI measure, including relevant derivations are
142 provided in the supporting information (SI).

143

144 **Materials and Methods**

145 *Domestic food transfer network.* We built the domestic food transfer and embodied impact
146 networks using existing empirical datasets. The framework along with data sources are detailed in
147 the supporting information (SI) table S1. The bi-lateral domestic food transfer data were obtained
148 from the Freight Analysis Framework (FAFv4).³⁵ FAF provides estimates for tonnage and value
149 of freight transported by origin and destination, commodity type, and transportation mode. The
150 latest available data are for 2012 and serve as the base year for this analysis. FAF data are for
151 groups of commodities based on Standard Classification of Transported Goods (SCTG)
152 classification system. The US agriculture is quite oligopolistic in terms of mass producing select
153 agriculture crops, with cereal and animal feed alone constituting 53% of national agricultural
154 production.³⁶ Additionally, compared to fruits and vegetables, grains are widely produced by many
155 states, providing sufficient data to compare production practices and assess resulting dependencies
156 arising from embodied impacts. Therefore, in this work, we focused on commodities covered by
157 SCTG 02 (cereal grains) and SCTG 04 (animal feed, eggs, honey, and products of other origin).
158 For SCTG 04, we specifically focus on only the animal feed related commodities as they comprise
159 the majority of this group.^{25, 35} We included wheat, corn, rice, sorghum, rye, barley, and oats for
160 grains and corn silage, sorghum silage, alfalfa hay, and hay for animal feed. Corn diverted to
161 bioethanol production was excluded based on national corn use statistics for 2012.³⁷ We note that

162 some of the grains from the cereal grains category may end up as animal feed for non-ruminant
163 livestock, however, accounting for all final uses falls outside the scope of this study.

164

165 The embodied impacts are estimated for specific commodities, while the trade data exists for
166 aggregated groups of commodities. To disaggregate shipments data, we assumed that composition
167 of grains in a shipment is similar to composition of production at origin. Therefore, if rice
168 production in Arkansas was 80% of total grains production, the grain shipments coming out of
169 Arkansas would consist of 80% rice. While transport based surveys provide a best available
170 substitute for interregional transfers accounting, they suffer from several limitations such as over-
171 assigning inflows to transport hubs, and not distinguishing between point of production vs. point
172 of last value added.³⁸ We corrected for this limitation as follows: we limited the analysis to transfer
173 of raw grains, animal feed, and associated impacts and did not track processed products. Therefore,
174 food transfers to a particular location may not represent the final consumption of a food item, but
175 the first-set of consumers (e.g., processing plants) in the supply chain. As such, the discussion on
176 dependency still remains relevant but we avoid overestimating environmental impacts of processed
177 goods. Additionally, by disaggregating transfers based on state production data, we overcome the
178 possibility of incorrectly attributing production to non-producing states. Similar approach for
179 interregional disaggregation has been employed previously.^{10, 25, 39} A brief discussion on regional
180 commodity transfer limitations and reconciliation issues is provided in the SI Section S2. Next, we
181 constructed weighted and directed matrices of food transfer referred to as flow matrices (T). Each
182 matrix element (T_{ij}) represents flow of mass of grains and animal feed from origin (i) state to
183 destination (j) state. The focus of this work is limited to irrigation impacts of food trade. By
184 irrigation impacts, we specifically mean irrigation water, embodied energy, and embodied GHG

185 emissions related to irrigation. A discussion on GHG impacts of U.S. food transport can be found
186 elsewhere.⁴⁰⁻⁴²

187
188 ***Embodied energy and GHG emissions networks.*** First, we calculated the fraction of irrigated food
189 transfers by assuming proportional shares to irrigated production. We converted food transfer
190 matrices into distinct matrices of virtual water, embodied energy, and embodied GHG emissions
191 by using data from the Farm and Ranch Irrigation Survey⁴³, U.S. agriculture census⁴⁴, Energy
192 Information Administration data⁴⁵ combined with life cycle assessment methods. In particular, we
193 use cumulative energy demand⁴⁶ and IPCC 100 year global warming potential to calculate our life
194 cycle impacts.⁴⁷ The detailed methodology and assumptions were first described by framework
195 provided by Vora et al.¹⁰

196 ***Pointwise mutual information (PMI).*** We analyze state-wise trade dependencies through
197 pointwise mutual information (PMI) measure. The PMI measure is based on concepts from
198 information theory, graph theory, probability, and statistics.⁴⁸ Commonly applied in linguistics^{34,}
199 ^{49, 50}, PMI calculates the probability of co-occurrence or co-location of two words (events). A
200 classic example involves comparing two synonym adjectives “strong” and “powerful” from
201 English language. A set of specific words are used more commonly with one or the other. As an
202 example, “strong tea” and “powerful car” have a higher probability of appearing together than
203 “powerful tea” and “strong car”; although the adjectives convey the same message.⁵¹ In a set
204 containing these four, if the information of the first word being “strong” is known, then “tea” has
205 a higher probability of being the next word. Thereby, reducing indeterminacy of the system.⁵² We
206 extend the same logic to assess trade dependencies by asking, for example, if we know a state is
207 importing food, can we predict any information about its partners given the set of data? We

208 perform this exercise not to predict new links but as a way of assessing statistical significance of
 209 empirically observed data. PMI is defined by the following eq. 1). The complete derivation of PMI
 210 measure is provided in the SI,

$$PMI_{ij} = k \log_2 \frac{p_{ij}}{p_i \cdot p_j} \quad (1)$$

211 p_{ij} is the probability of i and j co-occurring. k is a scalar constant. If events i and j are independent
 212 of each other, then the probability of their co-occurrence is given by their marginal probability of
 213 occurrences. Marginal probability of occurrence for event i is p_i . (eq. 2) and for j is given as p_j
 214 (eq. 3)

215

$$p_i = \sum_j p_{ij} \quad (2)$$

$$p_j = \sum_i p_{ij} \quad (3)$$

216 For flow networks such as the system under consideration, we can replace the probabilities of
 217 occurrence with measured frequency of flow in the network. T_{ij} represents flow of trade from
 218 origin (i) to destination (j). A “dot” notation is used to represent summation over that index such
 219 that T_i gives the total outgoing flows of i , T_j gives the total incoming flows to j , $T_{..}$ gives the total
 220 trade in the network, referred to as total system throughput.

221

$$p_{ij} = T_{ij}/T_{..}; \quad p_j = T_{.j}/T_{..}; \quad p_i = T_{i.}/T_{..} \quad (4)$$

222 Therefore, PMI can be re-written as,

$$PMI_{ij} = \log \frac{T_{ij}T_{..}}{T_{i.}T_{.j}} \quad (5)$$

223 In network trade studies, null modes or random networks have been used as a benchmark to
224 compare significance of structural properties of the observed/actual trade. If a random network can
225 generate higher order properties similar to those in observed trade, then observed structure of the
226 trade network is a result of random formation and estimating its properties does not give us any
227 useful information.³³ PMI measure essentially compares observed trade network with a pseudo-
228 random network (which is referred to as a null model). We use the term pseudo-random because
229 trade cannot be random, therefore comparing observed network to a completely random network
230 would not yield any meaningful insight. To make the null model comparable to the observed
231 network, some of the bare minimum properties of the observed network need to be preserved to
232 an otherwise randomly formed network. Here, the null model used to generate PMI values
233 constrains the network to keep the total inflow (demand) and outflow (supply) from each state
234 constant. This is an important constraint from sustainability perspective as it prevents states from
235 supplying more than their current reported capacity. This constraint results in a singular solution.
236 The flow matrix M , representing the null model of trade can be given by the following equation

$$M = F_{out}F_{in}T_{..} \quad (6)$$

$$F_{out} = \begin{bmatrix} T_{1\cdot}/T_{\cdot\cdot} \\ T_{2\cdot}/T_{\cdot\cdot} \\ \vdots \\ T_{n\cdot}/T_{\cdot\cdot} \end{bmatrix} \quad F_{in} = \begin{bmatrix} T_{\cdot 1}/T_{\cdot\cdot} & T_{\cdot 2}/T_{\cdot\cdot} & \dots & T_{\cdot n}/T_{\cdot\cdot} \end{bmatrix} \quad (7)$$

237 Here, F_{out} (51×1) and F_{in} (1×51) represent vectors of out-flows from and in-flows to each state
 238 respectively, normalized by the total flow in the system. Therefore, M is calculated by re-wiring
 239 network flows amongst each trade connection. A unique property of the null model is that it re-
 240 distributes flow in a way that the trade becomes more equitable (not equal) while considering
 241 current sending and receiving capacity of each state. Therefore, PMI values indicates how far each
 242 trade interaction is from being more equitable. An example of how the null model divides flow
 243 equitably is provided in SI section S4.

244
 245 The PMI measure can potentially take positive, negative, or zero values. If states i and j are
 246 completely independent (basis for null model), the value of PMI becomes 0. When i and j have a
 247 high probability of trading, but their actual trade is low, PMI values become negative (eq. 8).
 248 Similarly, a positive PMI indicates that states are more dependent than expected.

$$\log(p_{ij}) < \log(p_i p_j) \quad (8)$$

249 Previously, Kharrazi and Fath discussed the value of utilizing PMI measures to evaluate
 250 preferential trade policies within the context of international oil trade.⁵³ Based on PMI values, the
 251 aforementioned formulae can help evaluate policies for (un)desired trade relationships. It is to be
 252 noted that the goal is to not move towards a null model, as trade can never be random, but to
 253 understand more deeply the relations between dyads and to reverse the PMI value signs depending

254 on policy objectives, when desired. If a move from positive PMI to a negative PMI value is desired
255 (reduced trade) for a particular trade relationship, then trade volumes can be recalculated to
256 identify partners that can meet the additional demand. However, re-arranging even one pair would
257 alter the entire pattern of network flows indicating importance of considering interactions within
258 the entire system.

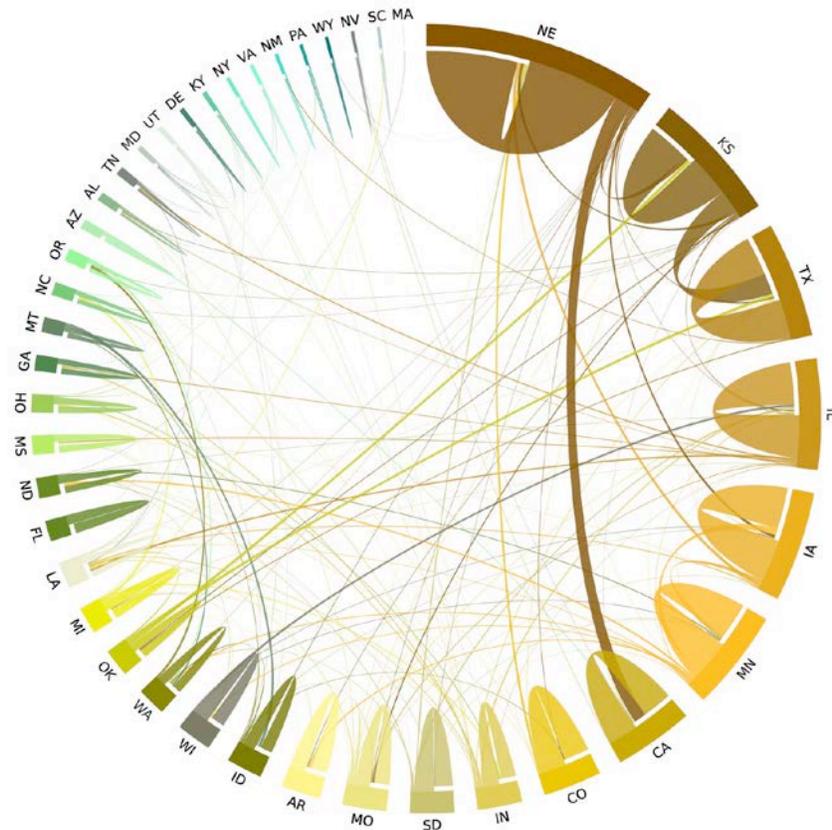
259

260 **RESULTS**

261 *Network Indicators.* We consider food transfers between 50 states plus District of Columbia,
262 creating a 51-node size (n) network. There are 1145 links (L) within these states dedicated to cereal
263 grains and animal feed trade. The density (L/n^2) of the network is 0.44 and reciprocity (proportion
264 of bi-directional connections (*links in both directions/total number of links*) of 0.64, indicating a
265 well-connected structure with high level of flow between states. The total flow in the network
266 amounts to 613 million US tons, with 166 billion m^3 of virtual water, 633 billion MJ of embodied
267 energy, and 42 billion kg CO_2 equivalents of GHG emissions embodied within the flows. Cereal
268 grains represent 75% of total food transfers by mass and subsequently represent a larger portion
269 of embodied irrigation impacts (SI Table S3). Figure 1 provides a visualization of irrigated
270 transfers within the U.S. The segments are arranged in a descending order based on their total out-
271 going activity. For a majority of the states, the highest volume of transfers are their within-state
272 flows. Nebraska's irrigated agriculture primarily includes corn for grain, corn silage, and alfalfa
273 hay. The large self-loop may indicate shipments going towards feeding the large cattle and hog
274 industry.⁵⁴ The largest (out of state) outgoing transfers are from Kansas, Nebraska, Minnesota,
275 Indiana, and Iowa. The largest inflows are to Texas, California, Nebraska, Illinois, and Iowa. The

276 largest out of state transfer is from Kansas to Texas of 18 million US tons and primarily consists
277 of corn, corn silage, alfalfa hay, and wheat in shipments.

278



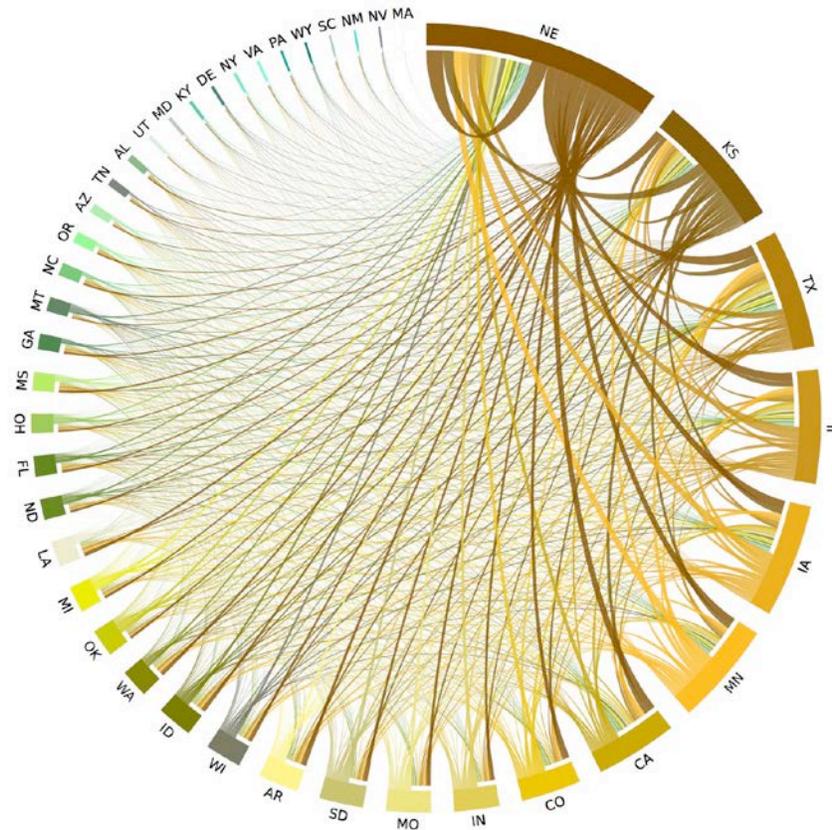
279

280 Figure 1. Cereal and feed grains transfer amongst the U.S. states. For visualization purpose, links
281 with at least 1% of maximum link weight are shown.¹⁹ Each circular segment represents
282 participating states. The white gap indicates in-coming transfers, while the same colored links
283 originating from the segment represents out-going transfers. The segments are arranged in
284 descending order based on their total out-going (both within state and out of state) transfers. The
285 figure is prepared using circos visualization tool.⁵⁵

286

287 Next, we visualize flow values according to null model in the system (figure 2). These values are
288 re-arranged in a more uniform fashion considering mass of the product of total flow going and
289 coming out of states. It should be noted that the flows are not re-distributed to become equal in
290 volume but based on equity in distribution. The degrees (number of connections) distribution and

291 weighted degree distributions for the observed flow and null model are provided in Figure S4 and
292 indicate maximum connectivity of the null model while preserving total throughput from each
293 state. Additionally, the density of the null model network is 0.9 with reciprocity of 0.79, indicating
294 an overly connected structure with more flows being reciprocated. When we compare the structure
295 of observed flow with the null model, the observed flow presents a preference in their transfers.
296 As there are no political boundaries compared to international trade,¹⁴ the preference represents
297 presence of “additional information” in how ties are formed.



298
299 Figure 2. Cereal and feed grains trade between U.S. states for null model(zero dependency). The
300 flow structure is redistributed considering network flow constraints such that total throughput
301 (both incoming and outgoing transfers) in each state remains constant. For consistency, links with
302 at least 1% of maximum link weight are shown. Each circular segment represents participating
303 states. The white gap indicates in-coming transfers, while same colored links emanating from the
304 segment represent out-going transfers.
305

306 ***Dependencies in the network.*** Generally, direct dependencies of trade relationships are identified
307 listing top importers/exporters for each trading partners. However, direct relationships do not
308 incorporate the role of considered relationship in the context of other relationships out of the two
309 states. This translates to how overall connections in the network (the system) affects one
310 relationship being studied. Additionally, a large volume of inflows may not translate to a higher
311 dependency for the pair, but low inflows may be more valuable to the network.^{32, 53} This is
312 explained in more detail next.

313 PMI values are calculated for each interaction between the dyads and therefore result in a 51×51
314 matrix for each network. As an example, we focus on Texas- the largest importer and its trading
315 partners to demonstrate the usefulness of considering system dependencies. Texas received
316 incoming transfers amounting to 49 million US tons from 34 states including a large chunk of
317 within-state transfers. Texas's largest inflows (apart from within-state flows) are from Kansas,
318 Oklahoma, Nebraska, Louisiana, and Indiana. Therefore, in a conventional sense, Texas highly
319 depends on these states for food flows. We rank PMI values from Texas's top ten import partners
320 in a descending order and compare with ranks of direct incoming transfer volume (Table 1).
321 Mismatches between PMI ranks and direct trade volume ranks show that associating dependencies
322 based on direct trade observations may not account for important, but less visible states. The PMI
323 value for New Mexico borders on zero, indicating the observed flow's proximity to null model
324 behavior. Considering all transfers from New Mexico, a substantial portion is already being
325 transferred to Texas, with a little room for increase (negative PMI), indicating a higher dependency
326 of the connection. On the other hand, Nebraska has a lower PMI rank and negative PMI value,
327 denoting that despite substantial volume of flows already going in to Texas, Nebraska has the
328 ability to send more, resulting in a lower bi-lateral dependence than possible. Kansas and

329 Oklahoma have the largest PMI values as Texas' exporting partners, indicating Texas's over
330 reliance on these two states. As observed from Table 1, majority of connections have negative PMI
331 values compared to positive values. This is consistent across the network in both import and export
332 connections for majority of states (SI Figure 3) indicating that at the network level, a few states
333 control the throughput of flow. This has important implications for local network structural
334 resiliency as reliance on a few states makes a state more prone to effect of shocks. Additionally,
335 some of the PMI rankings are consistent with mass/volume-based rankings denoting that the high
336 flows empirically observed are not by chance but statistically significant. A visualization of the
337 null model and observed flows along with extended PMI table for Texas is provided in the SI
338 section S5. We emphasize that by providing comparison of rankings, our motive is not to
339 recommend PMI method over traditional approaches, but to provide complementary insights along
340 with other commonly used measures.

341
342 Negative PMI values indicate a state's capacity to trade more (as the states are less reliant on each
343 other than expected), and therefore provide a first indication of where the trade could be rewired
344 without extensive economic and physical system modeling (such as used in crop displacement
345 studies ^{56, 57}

346

347

348

349

350

351 Table 1. Texas' top 10 importing partners ranked by their PMI value in a descending order,
 352 compared with observed incoming transfers and respective rank. Positive PMI indicates higher
 353 than expected dependency and negative PMI indicates lower than expected dependency.

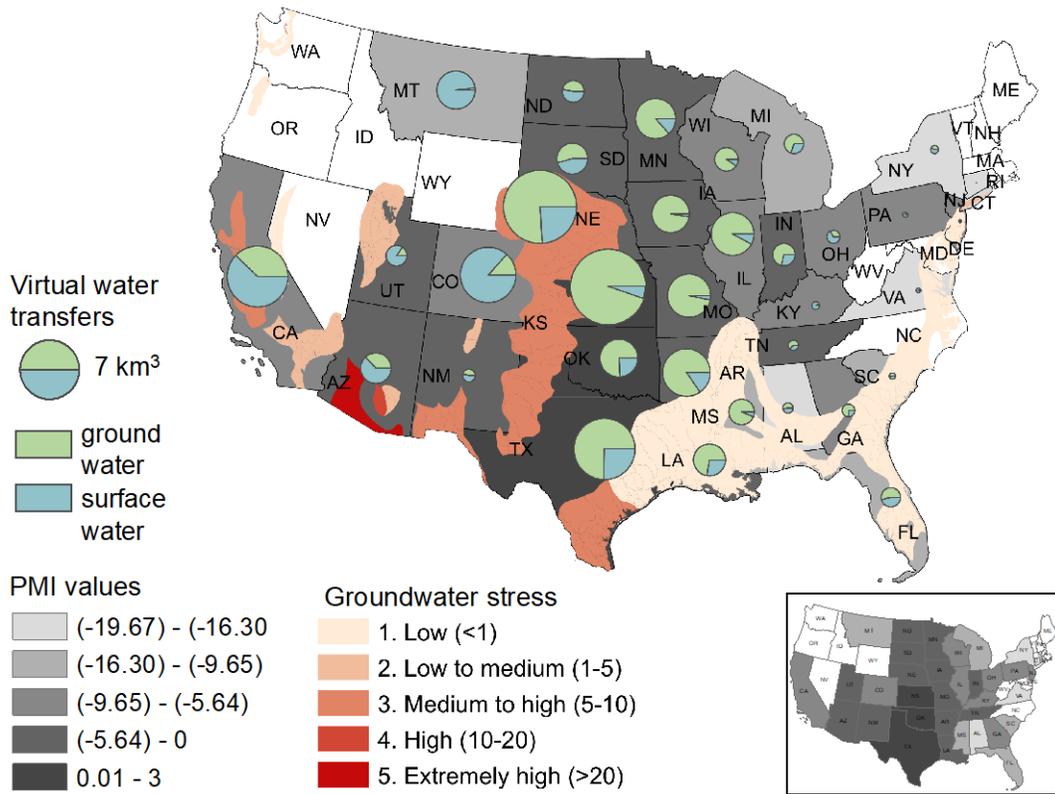
354

Incoming flow	PMI	PMI Rank	Flow (US tons)	Flow Rank
Texas	3.31	1	3.23E+07	1
Kansas	1.61	2	1.77E+07	2
Oklahoma	1.10	3	2.76E+06	3
Louisiana	0.23	4	9.38E+05	5
New Mexico	-0.05	5	1.19E+05	11
Indiana	-1.59	6	6.60E+05	6
Missouri	-2.17	7	4.06E+05	7
Tennessee	-2.51	8	5.99E+04	16
Nebraska	-2.76	9	1.37E+06	4
Arizona	-2.81	10	6.01E+04	15

355

356 *Embodied Impacts and Implications for FEW Nexus.* Next, we analyze trade interactions and
 357 dependencies within a FEW nexus context focusing on virtual water, embodied energy, and
 358 embodied GHG emissions.

359



360

361 Figure 3. PMI values for virtual water transfers to Texas (also included in inset for clearer
 362 visualization). The pie chart indicates portion of virtual surface and groundwater in food trade.
 363 The scale of pie chart represents total virtual water transfer out of each state (within-state flows
 364 included). The states colored in white represent absence of virtual water transfer to Texas. The
 365 primary groundwater aquifers of USA are overlaid in the graph with associated groundwater stress
 366 obtained from Gleeson et al. Aqueduct water risk atlas.^{58, 59}

367

368 A spatial display of the PMI values for virtual water transfers to Texas shows the pattern of near
 369 neighbors being higher ranked (figure 3). The dark grey shaded states represent high PMI values,
 370 and therefore higher dependence. Previous work has discussed the prevalence of gravity law^{60, 61}
 371 based relationship of distance enabling trade in international virtual water trade.⁶² The size of the
 372 pie chart represents total virtual water transfers out of each state. The scale of the pie chart accounts
 373 for irrigation intensity of crops (m³/ton) as well as volume of transfers. Statewide irrigation
 374 intensities are provided in the SI. Nebraska, Kansas, Louisiana, and Missouri have lower irrigation
 375 water application intensity, but overall higher volume of transfers. This may be attributed to

376 metering of groundwater due to regulations⁶³ along with high crop yields in the area. However,
377 high PMI ranked states New Mexico, Arizona, Colorado, and Utah have high water application
378 intensities, indicating virtual water hotspots in Texas' imports.

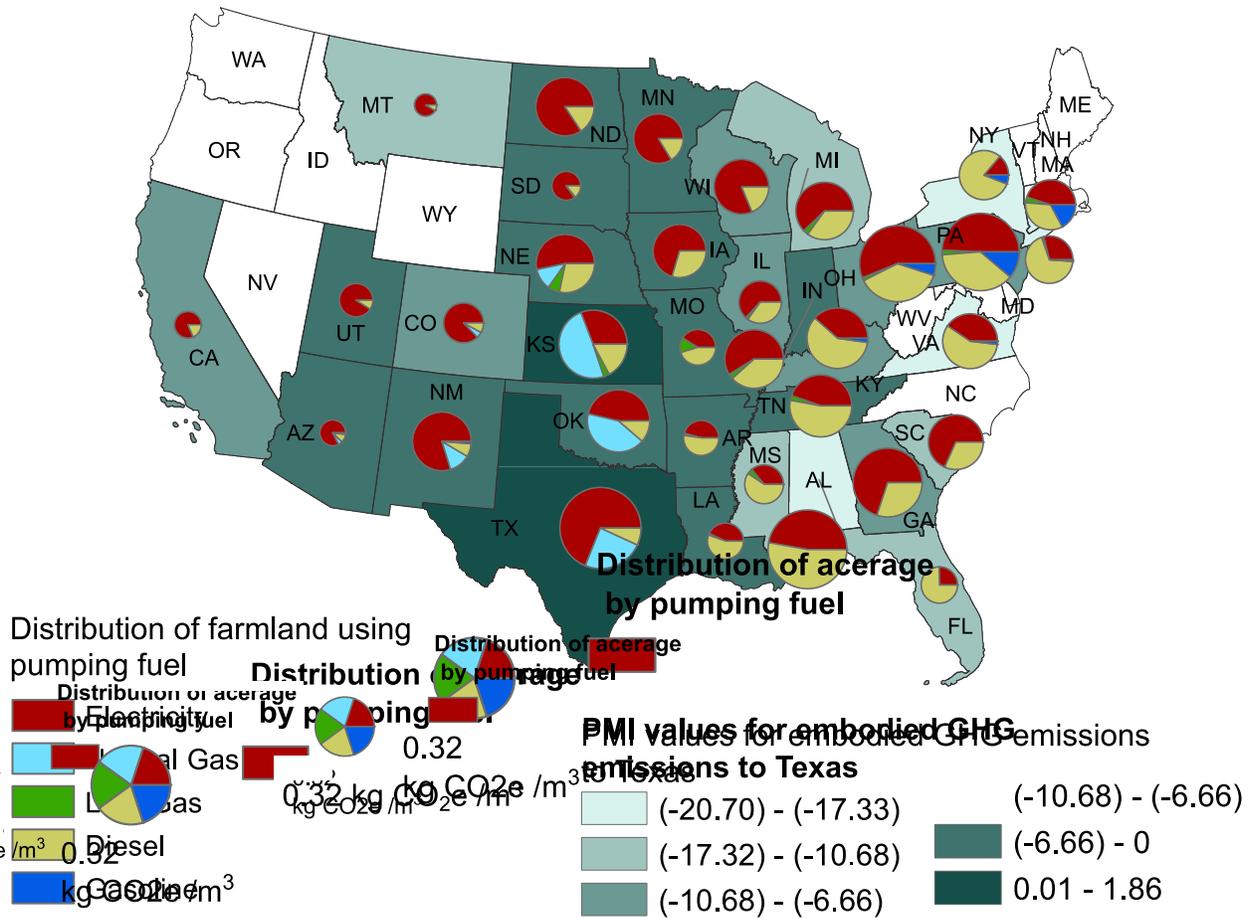
379

380 The pie charts show distribution of virtual groundwater and surface water used for production of
381 food transfers. A majority of Texas' exporters, and Texas, rely on groundwater for food imports.
382 Therefore, groundwater depletion is an important aspect in considering regional virtual water flow
383 dependencies. We overlay the PMI map with a layer of groundwater stress in major groundwater
384 basins, derived from Gleeson et al.⁵⁸ and Aqueduct database.⁵⁹ Groundwater stress represents
385 groundwater footprint over total aquifer area and is computed by setting up a water balance
386 between groundwater withdrawal, recharge, and environmental flows.⁵⁸ From South Dakota to
387 Texas, eight states heavily depend on the Ogallala aquifer as an important common groundwater
388 source for irrigation. The Ogallala aquifer's current use exceeds natural recharge with significant
389 decline in Kansas and Texas.⁶⁴ Scanlon et al.⁶⁵ estimate that if the current depletion rate continues,
390 then 35% of the southern plains would not be able to support irrigation in the next 30 years.
391 Therefore, despite lower water application intensity for some states, virtual water imports to Texas
392 from within-state flows, and neighbors Kansas, Oklahoma, and New Mexico may be affected by
393 groundwater depletion in the long run, especially as pressure on the shared Ogallala aquifer
394 increases from population demand and changing climate.⁶⁶

395

396 From a demand side, the possibility of groundwater shortage can be managed by re-structuring
397 existing trade to explore alternate states that have a higher potential to trade by looking at negative
398 PMI values. In such cases, states with policies that support sustainable irrigation can be given a

399 preference to build a water-scarcity resilient food supply chain. For example, lighter grey shaded
400 states such as Alabama, South Carolina, Florida, Kentucky, and Ohio have lower PMI value, low
401 water application intensity, and a balanced use of irrigation water sources, making them potential
402 candidates for increasing trade. However, the marginal environmental impact of increasing trade,
403 specifically on water quality in gulf states would have to be examined. From a supply side, majority
404 of Ogallala states have implemented state level groundwater management plans, along with some
405 moving beyond conservation and planning for depletion targets.⁶⁷ Schipanski et al.⁶⁸ note that the
406 next set of Ogallala strategies will require managing adaptation challenges for all the stakeholders
407 involved. In such cases, the mutual dependence due to regional trade can act as an incentive for
408 negotiations towards sustainable management of common source.



409

410 Figure 4. PMI values for embodied GHG emissions in imports to Texas. The pie chart
 411 indicates distribution of acreage using specific pumping fuel for on-farm irrigation pumps.
 412 The size of the pie chart indicates GHG emissions intensity in kg CO₂ equivalent per m³
 413 of water abstracted. The states colored in white represent absence of GHG transfer to
 414 Texas.

415

416 Figure 4 indicates PMI values for embodied GHG emissions transfers to Texas. Each pie chart
 417 represents the distribution of pumping fuels used in every state with all states employing
 418 electricity, and diesel-based pumps with a handful using natural gas (Texas, Oklahoma, Kansas,
 419 Nebraska), gasoline (Pennsylvania, Ohio, Rhode Island, New York) and LPG (Nebraska,
 420 Missouri) based pumps. The size of the pie chart indicates GHG emissions intensity in kg CO₂
 421 equivalent per m³ of water abstracted. Barring electricity, natural gas-based pumps have the lowest

422 embodied GHG emissions intensity amongst all four fuels considered. Life cycle emissions
423 attributable to electricity-based pumping differ considerably across states due to differences in
424 regional grid mixes. Apart from electricity, all the states use diesel-based pumps in some capacity,
425 with eastern states using diesel pumps on significant acreage. In addition to fuel mix, pumping
426 energy requirements depend on other factors such as type of irrigation system (gravity vs. pressure
427 based), system pressure, depth to water for lift, velocity, and pipe losses.⁶⁹ Contrarily to water
428 intensity for crops, California, Colorado, Arizona, Arkansas, and Utah have lower GHG emissions
429 intensity per m³ of water withdrawn. These states primarily use gravity-based irrigation or rely on
430 lower to medium pressure systems. Many of the Ogallala states, despite using substantial natural
431 gas in their pumping mix, have higher GHG emissions per m³ of water withdrawn. This could be
432 attributed to high coal-based electricity mix in their grid (e.g., Kansas, Nebraska, Oklahoma have
433 more than 60% coal-based generation), water depth for groundwater pumping, and use of water
434 efficient but energy intensive pressurized sprinkler systems. High use of diesel and/or gasoline-
435 based pumps combined with pressurized irrigation systems could be contributing to high GHG
436 emissions intensity of states such as Pennsylvania, Ohio, Alabama, and Kentucky.⁴³ These states
437 represent a clear example of water scarcity vs. GHG emissions tradeoff and denote an area of farm
438 conservation policy focus for improving pumping energy and emissions profile of irrigation by
439 upgrading fuel pumps. As part of Ogallala conservation efforts, several programs have been
440 underway since 2008 to reduce irrigation withdrawals and, as a result have also reduced energy
441 requirements of farms suggesting that groundwater conservation and irrigation emissions
442 reductions may not be mutually exclusive goals.⁷⁰

443

444

445 **DISCUSSION**

446 This work provides a systems-level perspective in analyzing domestic food-energy-water
447 interactions (within regional transfers and between embodied systems) through interdisciplinary
448 methods spanning information theory, graph theory, water footprint, embodied energy, and
449 emissions quantification. We demonstrate the usefulness of considering interactions at a network
450 level to provide a comprehensive indication of trade dependencies. Using Texas as an example,
451 we show that major importing partners of Texas by volume may not rank high in expected trading
452 as expressed here in the index of PMI values and vice versa. A bi-lateral trade relationship consists
453 of an interaction between a dyad, with both partners playing an equally important role. Ranking
454 Texas' exporters by volume only showcases Texas' dependency of the transfer but not of its
455 partners. As PMI accounts for overall transfer activity and the potential to increase (or decrease)
456 activity between a dyad, it provides a thorough accounting of their mutual dependency. This is
457 clearly exhibited in importance of Texas-New Mexico trade connection despite being of a lower
458 volume, and reiterates the importance of also considering weak ties.³²

459
460 When we compare the visual difference between flow in a null model and actual trade, the
461 heterogenous distribution in trade concertation becomes apparent with a few links/states
462 dominating the network (figure S3). Another visible trend is the importance of geographical
463 distance in forming trade relationships. Our results indicate that distance drives the grain and
464 animal feed trade preference for Texas, specifically as a significant portion may be dedicated to
465 providing cost-effective animal feed for Texas' sizable cattle industry or for food and beverage
466 manufacturing. By combining PMI results and a ground water stress indicator, we highlight the
467 regional reliance of Texas' and neighboring states on Ogallala aquifer for irrigation while engaging

468 in substantial transfer amongst themselves and discuss alternate potential states with less stressed
469 irrigation systems. In fact, dependence through regional trade can serve as a motivation to manage
470 common water resources and help avoid water allocation disputes such as the recent one between
471 New Mexico and Texas⁷¹ and between users of Colorado River basin.⁷² Further, considerable
472 geographic variation exists in recharge rates across the Ogallala aquifer due to its sub-surface
473 hydrology.⁶⁵ Therefore, our estimates can be improved in the future by characterizing the portion
474 of domestic food consumption attributed to nonrenewable groundwater withdrawals from U.S.
475 aquifers.⁷³

476

477 The analysis presented in this work has its limitations. An important limitation of this work is the
478 FAF dataset's inability to trace the final point of consumption (e.g., household consumption). This
479 would require integration and reconciliation of a larger scale of datasets to accurately track the
480 supply chain, such as the recent study of corn supply chain by Smith et al.⁷⁴ Additionally, future
481 domestic trade analysis should involve employing origin tracing algorithms⁷⁵ used in international
482 trade studies to remove re-exports from the data. From a systems-level analysis, we emphasize that
483 no one method is universally superior over other methods including techniques such as life cycle
484 assessment, material flow analysis, network analysis etc. Additionally, we note that while PMI
485 provides information on structural dependency based on trade data, it cannot differentiate between
486 a (un)desirable option based on embodied impacts such as type of water resource, water scarcity,
487 and fossil fuel used as this information is not inherently built-into snapshot of trade. Therefore, it
488 needs to be supplemented with footprint approaches, life cycle assessment methods to provide a
489 complete picture.

490

491 Furthermore, we do not account for energy and emissions associated with off-farm water supply
492 (prevalent in the western U.S.)⁷⁶ due to lack of national data, making our estimates conservative
493 and likely to increase. Therefore, if future policies internalize the cost of GHG emissions in trade,
494 states may look for cost-effective and cleaner energy options with natural gas currently being one
495 of the easily accessible choice. As our results demonstrate, this may be at odds with other equally
496 important goals to achieve a sustainable and resilient food supply. Specific policies have long been
497 in place under the U.S. Farm Bill to subsidize switching to water-efficient irrigation systems, but
498 a rebound effect of over-pumping may lead to water depletion⁷⁷ and salinization.⁷⁸ At the same
499 time, the discussion on FEW nexus should incorporate electric utilities and authorities that can
500 devise demand-response programs for farmers to offer electricity at lower prices off-peak and
501 potentially manage the emissions profile of generators.⁷⁹⁻⁸¹ Finally, PMI values demonstrate the
502 potential to trade less (positive PMI) or more (negative PMI) given the existing network constraints
503 compared to the situation of no preference. Therefore, it may serve as a valuable policy aid in
504 building sustainable and resilient food systems by indicating overall effect of potential trade
505 (dis)preferences for diversifying trade partners.

506

507 **ASSOCIATED CONTENT**

508 Additional information regarding data sources, code for PMI, and the modeling approach is
509 provided in the Supporting Information.

510

511 **ACKNOWLEDGMENTS**

512 This research is supported by the National Science Foundation (award number CBET 1803527).
513 NV was supported by the 2017 Young Scientists Summer Program (YSSP) at the International
514 Institute for Applied Systems Analysis with the financial assistance provided by Ferrero Trade

515 Lux S.A. Opinions, findings, or recommendations expressed in this material do not necessarily
516 reflect the views of these organizations.
517

518 REFERENCES

- 519 1. Nilsson, M.; Griggs, D.; Visbeck, M., Map the interactions between sustainable
520 development goals. *Nature* **2016**, *534*, (7607), 320-323.
- 521 2. Liu, J.; Hull, V.; Godfray, H. C. J.; Tilman, D.; Gleick, P.; Hoff, H.; Pahl-Wostl, C.; Xu,
522 Z.; Chung, M. G.; Sun, J., Nexus approaches to global sustainable development. *Nature*
523 *Sustainability* **2018**, *1*, (9), 466.
- 524 3. D'Odorico, P.; Davis, K. F.; Rosa, L.; Carr, J. A.; Chiarelli, D.; Dell'Angelo, J.; Gephart,
525 J.; MacDonald, G. K.; Seekell, D. A.; Suweis, S., The global food-energy-water nexus. *Reviews*
526 *of geophysics* **2018**, *56*, (3), 456-531.
- 527 4. Scanlon, B. R.; Ruddell, B. L.; Reed, P. M.; Hook, R. I.; Zheng, C.; Tidwell, V. C.;
528 Siebert, S., The food-energy-water nexus: Transforming science for society. *Water Resources*
529 *Research* **2017**, *53*, (5), 3550-3556.
- 530 5. Cai, X.; Wallington, K.; Shafiee-Jood, M.; Marston, L., Understanding and managing the
531 food-energy-water nexus—opportunities for water resources research. *Advances in Water*
532 *Resources* **2018**, *111*, 259-273.
- 533 6. Walker, R. V.; Beck, M. B.; Hall, J. W.; Dawson, R. J.; Heidrich, O., The energy-water-
534 food nexus: Strategic analysis of technologies for transforming the urban metabolism. *Journal of*
535 *environmental management* **2014**, *141*, 104-115.
- 536 7. Yang, Y. E.; Wi, S.; Ray, P. A.; Brown, C. M.; Khalil, A. F., The future nexus of the
537 Brahmaputra River Basin: climate, water, energy and food trajectories. *Global environmental*
538 *change* **2016**, *37*, 16-30.
- 539 8. Boyer, D.; Ramaswami, A., What Is the Contribution of City-Scale Actions to the Overall
540 Food System's Environmental Impacts?: Assessing Water, Greenhouse Gas, and Land Impacts
541 of Future Urban Food Scenarios. *Environmental science & technology* **2017**, *51*, (20), 12035-
542 12045.
- 543 9. White, D. J.; Hubacek, K.; Feng, K.; Sun, L.; Meng, B., The Water-Energy-Food Nexus
544 in East Asia: A tele-connected value chain analysis using inter-regional input-output analysis.
545 *Applied Energy* **2018**, *210*, 550-567.
- 546 10. Vora, N.; Shah, A.; Bilec, M. M.; Khanna, V., Food-Energy-Water Nexus: Quantifying
547 Embodied Energy and GHG emissions from Irrigation through Virtual Water Transfers in Food
548 Trade. *ACS Sustainable Chemistry & Engineering* **2017**.
- 549 11. Schaible, G.; Aillery, M., Water conservation in irrigated agriculture: Trends and
550 challenges in the face of emerging demands. *USDA-ERS Economic Information Bulletin* **2012**,
551 (99).
- 552 12. Maupin, M. A.; Kenny, J. F.; Hutson, S. S.; Lovelace, J. K.; Barber, N. L.; Linsey, K. S.
553 *Estimated use of water in the United States in 2010*; 2330-5703; US Geological Survey: 2014.
- 554 13. Sands, R.; Westcott, P. C.; Price, M.; Beckman, J.; Leibtag, E.; Lucier, G.; McBride, W.;
555 McGranahan, D.; Morehart, M.; Roeger, E., *Impacts of higher energy prices on agriculture and*
556 *rural economies*. United States Department of Agriculture: 2011.
- 557 14. Lin, X.; Dang, Q.; Konar, M., A network analysis of food flows within the United States
558 of America. *Environmental science & technology* **2014**, *48*, (10), 5439-5447.

- 559 15. Hirschman, A. O., *National power and the structure of foreign trade*. Univ of California
560 Press: 1980; Vol. 105.
- 561 16. Gasiorowski, M.; Polachek, S. W., Conflict and interdependence: East-West trade and
562 linkages in the era of detente. *Journal of Conflict Resolution* **1982**, 26, (4), 709-729.
- 563 17. Dixon, W. J., Trade concentration, economic growth, and the provision of basic human
564 needs. *Social Science Quarterly* **1984**, 65, (3), 761.
- 565 18. Qu, S.; Liang, S.; Konar, M.; Zhu, Z.; Chiu, A. S. F.; Jia, X.; Xu, M., Virtual Water
566 Scarcity Risk to the Global Trade System. *Environmental Science & Technology* **2018**, 52, (2),
567 673-683.
- 568 19. Dalin, C.; Wada, Y.; Kastner, T.; Puma, M. J., Groundwater depletion embedded in
569 international food trade. *Nature* **2017**, 543, (7647), 700-704.
- 570 20. Marston, L.; Konar, M.; Cai, X.; Troy, T. J., Virtual groundwater transfers from
571 overexploited aquifers in the United States. *Proceedings of the National Academy of Sciences*
572 **2015**, 112, (28), 8561-8566.
- 573 21. Allan, J., *Fortunately there are substitutes for water otherwise our hydro-political futures*
574 *would be impossible*. Overseas Development Administration: London, 1993; Vol. 13, p 26.
- 575 22. Fader, M.; Gerten, D.; Krause, M.; Lucht, W.; Cramer, W., Spatial decoupling of
576 agricultural production and consumption: quantifying dependences of countries on food imports
577 due to domestic land and water constraints. *Environmental Research Letters* **2013**, 8, (1),
578 014046.
- 579 23. Suweis, S.; Rinaldo, A.; Maritan, A.; D'Odorico, P., Water-controlled wealth of nations.
580 *Proceedings of the National Academy of Sciences* **2013**, 110, (11), 4230-4233.
- 581 24. D'Odorico, P.; Carr, J.; Laio, F.; Ridolfi, L., Spatial organization and drivers of the
582 virtual water trade: A community-structure analysis. *Environmental Research Letters* **2012**, 7,
583 (3), 034007.
- 584 25. Dang, Q.; Lin, X.; Konar, M., Agricultural virtual water flows within the United States.
585 *Water Resources Research* **2015**, 51, (2), 973-986.
- 586 26. Fang, D.; Chen, B., Ecological network analysis for a virtual water network.
587 *Environmental science & technology* **2015**, 49, (11), 6722-6730.
- 588 27. Kharrazi, A.; Rovenskaya, E.; Fath, B. D., Network structure impacts global commodity
589 trade growth and resilience. *PloS one* **2017**, 12, (2), e0171184.
- 590 28. Dupas, M.-C.; Halloy, J.; Chatzimpiros, P., Time dynamics and invariant subnetwork
591 structures in the world cereals trade network. *PloS one* **2019**, 14, (5), e0216318.
- 592 29. Gephart, J. A.; Pace, M. L., Structure and evolution of the global seafood trade network.
593 *Environmental Research Letters* **2015**, 10, (12), 125014.
- 594 30. Konar, M.; Dalin, C.; Suweis, S.; Hanasaki, N.; Rinaldo, A.; Rodriguez-Iturbe, I., Water
595 for food: The global virtual water trade network. *Water Resources Research* **2011**, 47, (5).
- 596 31. Salmoral, G.; Yan, X., Food-energy-water nexus: A life cycle analysis on virtual water
597 and embodied energy in food consumption in the Tamar catchment, UK. *Resources,*
598 *Conservation and Recycling* **2018**, 133, 320-330.
- 599 32. Granovetter, M. S., The strength of weak ties. *American journal of sociology* **1973**, 78,
600 (6), 1360-1380.
- 601 33. Fagiolo, G.; Squartini, T.; Garlaschelli, D., Null models of economic networks: the case
602 of the world trade web. *Journal of Economic Interaction and Coordination* **2013**, 8, (1), 75-107.
- 603 34. Church, K. W.; Hanks, P., Word association norms, mutual information, and
604 lexicography. *Computational linguistics* **1990**, 16, (1), 22-29.

- 605 35. Hwang, H.-L.; Hargrove, S.; Chin, S.-M.; Wilson, D. W.; Davidson, D. *Freight Analysis*
606 *Framework Version 4-Building the FAF4 Regional Database: Data Sources and Estimation*
607 *Methodologies*; Oak Ridge National Laboratory (ORNL), Oak Ridge, TN (United States): 2016.
- 608 36. FAO, Food balance sheets. In 2012.
- 609 37. O'Donoghue, E.; Hansen, J., *USDA Agricultural Projections to 2026*. 2017.
- 610 38. Sargento, A. L.; Ramos, P. N.; Hewings, G. J., Inter-regional trade flow estimation
611 through non-survey models: An empirical assessment. *Economic Systems Research* **2012**, *24*,
612 (2), 173-193.
- 613 39. Marston, L.; Konar, M., Drought impacts to water footprints and virtual water transfers of
614 the Central Valley of California. *Water Resources Research* **2017**, *53*, (7), 5756-5773.
- 615 40. Taptich, M. N.; Horvath, A., Freight on a Low-Carbon Diet: Accessibility, Freightsheds,
616 and Commodities. *Environmental Science & Technology* **2015**, *49*, (19), 11321-11328.
- 617 41. Heller, M. C.; Keoleian, G. A., Exploring a water/energy trade-off in regional sourcing of
618 livestock feed crops. *Environmental science & technology* **2011**, *45*, (24), 10619-10626.
- 619 42. Weber, C. L.; Matthews, H. S., Food-miles and the relative climate impacts of food
620 choices in the United States. *Environmental science & technology* **2008**, *42*, (10), 3508-3513.
- 621 43. USDA, Farm and Ranch Irrigation Survey. In 2013.
- 622 44. USDA, Census of Agriculture. In 2012.
- 623 45. Energy Information Administration, U. S. Monthly Energy Review 2014.
- 624 46. Jungbluth, N.; Frischknecht, R., Cumulative energy demand. *LCIA Implementation. CD*
625 *ROM. Final report ecoinvent* **2000**, (3).
- 626 47. Stocker, T.; Qin, D.; Plattner, G.; Tignor, M.; Allen, S.; Boschung, J.; Nauels, A.; Xia,
627 Y.; Bex, B.; Midgley, B., IPCC, 2013: climate change 2013: the physical science basis.
628 Contribution of working group I to the fifth assessment report of the intergovernmental panel on
629 climate change. **2013**.
- 630 48. Ulanowicz, R. E.; Goerner, S. J.; Lietaer, B.; Gomez, R., Quantifying sustainability:
631 resilience, efficiency and the return of information theory. *Ecological complexity* **2009**, *6*, (1),
632 27-36.
- 633 49. Bullinaria, J. A.; Levy, J. P., Extracting semantic representations from word co-
634 occurrence statistics: A computational study. *Behavior research methods* **2007**, *39*, (3), 510-526.
- 635 50. Recchia, G.; Jones, M. N., More data trumps smarter algorithms: Comparing pointwise
636 mutual information with latent semantic analysis. *Behavior research methods* **2009**, *41*, (3), 647-
637 656.
- 638 51. Halliday, M. A., Lexis as a linguistic level. *In memory of JR Firth* **1966**, *148*, 162.
- 639 52. Goerner, S. J.; Lietaer, B.; Ulanowicz, R. E., Quantifying economic sustainability:
640 Implications for free-enterprise theory, policy and practice. *Ecological Economics* **2009**, *69*, (1),
641 76-81.
- 642 53. Kharrazi, A.; Fath, B. D., Measuring global oil trade dependencies: An application of the
643 point-wise mutual information method. *Energy Policy* **2016**, *88*, 271-277.
- 644 54. NASS Nebraska state agriculture overview.
645 https://www.nass.usda.gov/Quick_Stats/Ag_Overview/stateOverview.php?state=NEBRASKA
- 646 55. Krzywinski, M.; Schein, J.; Birol, I.; Connors, J.; Gascoyne, R.; Horsman, D.; Jones, S.
647 J.; Marra, M. A., Circos: an information aesthetic for comparative genomics. *Genome research*
648 **2009**, *19*, (9), 1639-1645.
- 649 56. Davis, K. F.; Seveso, A.; Rulli, M. C.; D'Odorico, P., Water savings of crop
650 redistribution in the United States. *Water* **2017**, *9*, (2), 83.

651 57. Davis, K. F.; Rulli, M. C.; Seveso, A.; D'Odorico, P., Increased food production and
652 reduced water use through optimized crop distribution. *Nature Geoscience* **2017**, *10*, (12), 919-
653 924.

654 58. Gleeson, T.; Wada, Y.; Bierkens, M. F.; van Beek, L. P., Water balance of global aquifers
655 revealed by groundwater footprint. *Nature* **2012**, *488*, (7410), 197-200.

656 59. Gassert, F.; Landis, M.; Luck, M.; Reig, P.; Shiao, T., Aqueduct global maps 2.0. *Water*
657 *Resources Institute (WRI): Washington, DC* **2013**, 202011-2012.

658 60. Bergstrand, J. H., The gravity equation in international trade: some microeconomic
659 foundations and empirical evidence. *The review of economics and statistics* **1985**, 474-481.

660 61. Anderson, J. E., A theoretical foundation for the gravity equation. *The American*
661 *Economic Review* **1979**, *69*, (1), 106-116.

662 62. Tamea, S.; Carr, J.; Laio, F.; Ridolfi, L., Drivers of the virtual water trade. *Water*
663 *Resources Research* **2014**, *50*, (1), 17-28.

664 63. OECD, *Drying Wells, Rising Stakes: Towards Sustainable Agricultural Groundwater*
665 *Use*. Paris, 2015.

666 64. McGuire, V., Water-level changes in the High Plains aquifer, predevelopment to 2007,
667 2005-06, and 2006-07. *Publications of the US Geological Survey* **2009**, 17.

668 65. Scanlon, B. R.; Faunt, C. C.; Longuevergne, L.; Reedy, R. C.; Alley, W. M.; McGuire, V.
669 L.; McMahon, P. B., Groundwater depletion and sustainability of irrigation in the US High
670 Plains and Central Valley. *Proceedings of the national academy of sciences* **2012**, *109*, (24),
671 9320-9325.

672 66. Little, J. B., The Ogallala aquifer: saving a vital US water source. *Scientific American*,
673 *March* **2009**.

674 67. Schipansky, M.; Auvermann, B.; Gowda, P.; Guerrero, B.; Kremen, A.; Porter, D.; Rice,
675 C.; Sanderson, M.; Wagner, K.; Warren, J.; West, C.; Waskom, R., *The Future of the Ogallala*
676 *Aquifer*. 2017.

677 68. Jarvis, T.; Wolf, A., Managing water negotiations and conflicts in concept and in
678 practice. *Transboundary Water Management: Principles and Practice* **2010**, 125-141.

679 69. Kahn, E. Characterization of Uncertainty and Variability of Freshwater Consumption
680 Impacts in Life Cycle Assessment. Ph.D. Dissertation, University of Washington, 2013.

681 70. Gollehon, N.; Winston, B., Groundwater irrigation and water withdrawals: the Ogallala
682 aquifer initiative. *USDA Economic Series* **2013**, *15*.

683 71. Tory, S., A Southwest water dispute reaches the Supreme Court. *High Country News*
684 *Jan.23*, 2018.

685 72. Rothberg, D., States accuse Arizona water agency of gaming Lake Mead, undermining
686 Colorado River drought plans. *The Nevada Independent* April 17, 2018, 2018.

687 73. Wada, Y.; Beek, L.; Bierkens, M. F., Nonsustainable groundwater sustaining irrigation:
688 A global assessment. *Water Resources Research* **2012**, *48*, (6).

689 74. Smith, T. M.; Goodkind, A. L.; Kim, T.; Pelton, R. E.; Suh, K.; Schmitt, J., Subnational
690 mobility and consumption-based environmental accounting of US corn in animal protein and
691 ethanol supply chains. *Proceedings of the National Academy of Sciences* **2017**, *114*, (38), E7891-
692 E7899.

693 75. Kastner, T.; Kastner, M.; Nonhebel, S., Tracing distant environmental impacts of
694 agricultural products from a consumer perspective. *Ecological Economics* **2011**, *70*, (6), 1032-
695 1040.

- 696 76. Tidwell, V. C.; Moreland, B.; Zemlick, K., Geographic footprint of electricity use for
697 water services in the Western US. *Environmental science & technology* **2014**, *48*, (15), 8897-
698 8904.
- 699 77. Nixon, R., Farm Subsidies Leading to More Water Use. *The New York Times* June 6,
700 2013.
- 701 78. Schoups, G.; Hopmans, J. W.; Young, C. A.; Vrugt, J. A.; Wallender, W. W.; Tanji, K.
702 K.; Panday, S., Sustainability of irrigated agriculture in the San Joaquin Valley, California.
703 *Proceedings of the National Academy of Sciences* **2005**, *102*, (43), 15352-15356.
- 704 79. Siler-Evans, K.; Azevedo, I. s. L.; Morgan, M. G., Marginal emissions factors for the US
705 electricity system. *Environmental science & technology* **2012**, *46*, (9), 4742-4748.
- 706 80. Chambers, A.; Kline, D.; Vimmerstedt, L.; Diem, A.; Dismukes, D.; Mesyanzhinov, D.,
707 Comparison of methods for estimating the NO_x. **2005**.
- 708 81. Marks, G.; Wilcox, E.; Olsen, D.; Goli, S. *Opportunities for demand response in*
709 *California agricultural irrigation: A scoping study*; Lawrence Berkeley National Lab.(LBNL),
710 Berkeley, CA (United States): 2013.
711