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Health and economic benefits of clean air policies in China: A case study for Beijing-Tianjin-Hebei region

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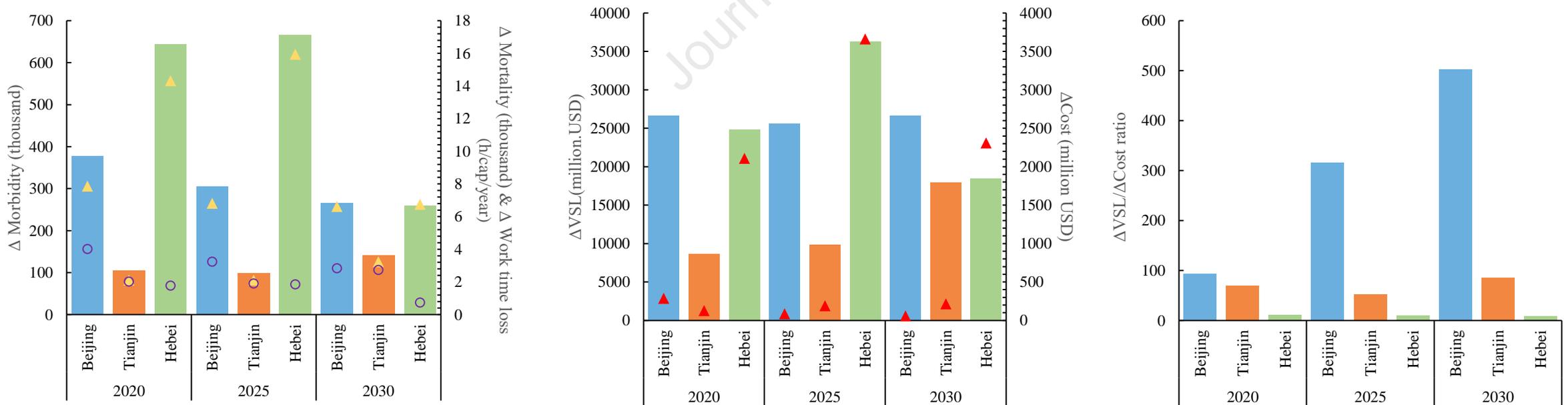
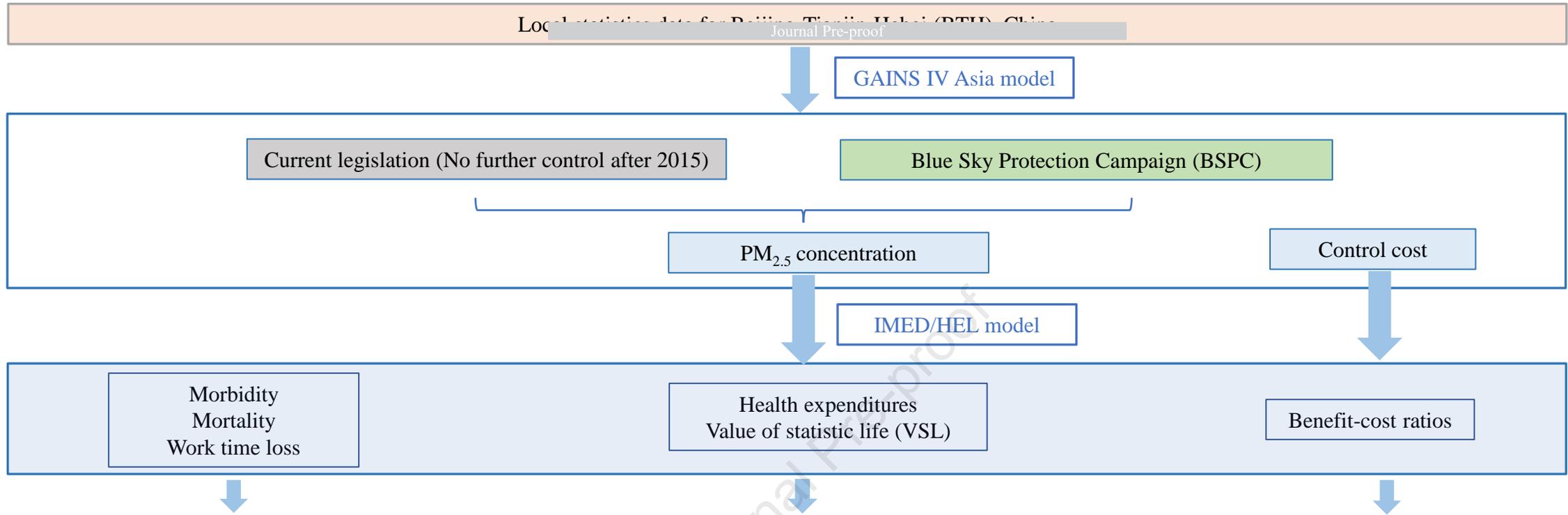
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Meng Xu and Yang Xie were responsible for data processing and model analysis.

Meng Xu and Shaohui Zhang provided research design. Meng Xu and Zhongfeng Qin jointly wrote the manuscript. Zhongfeng Qin and Yang Xie were responsible for revising the manuscript.

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Bars refer to the left Y-axis while triangles and circles refer to the right Y-axis



27 significantly, with 17 million USD (18%) and 63 billion USD (10%), respectively, in  
28 the BTH region. Besides, the economic benefits far exceed the policy costs of the  
29 BSPP, and the  $\Delta$  benefit/ $\Delta$  cost ratios of Beijing are significantly higher than those of  
30 Hebei. The BSPP in BTH has significant positive health and economic impacts. This  
31 study can provide a basis for future PM<sub>2.5</sub>-related health risk studies at an urban level  
32 in China.

33

34 **Keywords:**

35 Blue Sky Protection Campaign, Air quality, Health assessments, Economic effects,  
36 Integrated assessment methods.

37

38 **1 Introduction**

39 With the fast industrialization and urbanization in China, air pollution has  
40 become increasingly prominent (Cao et al., 2011; Dong et al., 2015; Hossain et al.,  
41 2021; Kuerban et al., 2020). Over 83% of the Chinese population live in areas with  
42 annual PM<sub>2.5</sub> concentrations that exceed the World Health Organization Interim  
43 Target-1 (WHO IT-1) of 35  $\mu\text{g}/\text{m}^3$ . Air pollution in China leads to serious health  
44 concerns (Brauer et al., 2012; Brauer et al., 2016; Liu et al., 2016). Epidemiological  
45 evidence has shown that PM<sub>2.5</sub> exposure is associated with increasing the risks of  
46 respiratory, cardiovascular, and cerebrovascular diseases and even premature deaths  
47 (Dockery et al., 1993; Pope et al., 2002; Pope et al., 2004). For instance, the  
48 premature deaths associated with PM<sub>2.5</sub> accounted for nearly 7.1% of worldwide  
49 deaths in 2010 (Evans et al., 2013). An analysis from Global Burden of Disease  
50 suggested that PM<sub>2.5</sub> pollution had caused 4.09 million mortalities in 2016, in which  
51 ischemic heart disease, chronic obstructive pulmonary disease, lower respiratory  
52 infections, and lung cancer accounting for 38.51%, 19.23%, 15.97%, and 6.83%,  
53 respectively (Roth et al., 2018). PM<sub>2.5</sub> pollution has been recognized as the prime  
54 environmental health risk (Cohen et al., 2005; Lim et al., 2012), predominantly in  
55 China and India (Anenberg et al., 2010; Ji et al., 2019; Lelieveld et al., 2015). Based

56 on the provincial PM<sub>2.5</sub> concentrations and health data, Rong and Wang (2016)  
57 suggested that PM<sub>2.5</sub> pollution in China caused 1,255,400 mortalities in 2010, 42%  
58 higher than the level in 2000. Further, mortalities associated with PM<sub>2.5</sub> pollution  
59 would be 2.3 million in China by 2030 (Xie et al., 2019). It suggests that PM<sub>2.5</sub>  
60 pollution in China has been an urgent issue needing been addressed.

61 The adverse effects of PM<sub>2.5</sub> pollution on health also lead to economic costs,  
62 including additional expenditures and labor productivity reduction. Tian et al. (2018)  
63 evaluated the economic impacts of PM<sub>2.5</sub>-related health impacts on China's road  
64 transport sector. They found air pollution from transportation led to 442.90 billion  
65 Yuan of the value of statistical life (VSL) loss and 2.09 hours/capita of work time loss  
66 in 2015. Besides, it would cause 737.15 billion Yuan of statistical life loss and 2.23  
67 hours/capita of work time loss by 2030 (Tian et al., 2018). Xie et al. (2016a)  
68 examined the health and economic effects of PM<sub>2.5</sub> at a provincial level in China,  
69 indicating that without any pollution controls, China would experience 2.00% GDP  
70 loss and 25.2 billion USD in health expenditures by 2030. Xie et al. (2016b)  
71 concluded that controlling PM<sub>2.5</sub> pollution could avoid 80.4% of health expenditure  
72 expenditures, and the benefit is much higher than the policy cost in the BTH region.  
73 Beijing, Tianjin, and Hebei would obtain 1.75%, 2.02%, and 1.46% GDP gains. Wu et  
74 al. (2017) evaluated the health and economic impacts of PM<sub>2.5</sub> exposure in Shanghai  
75 and suggested that controlling emissions could avoid 82.7% of mortality cases, 1.89%  
76 of GDP losses, and 2.63% of welfare losses.

77 To control PM<sub>2.5</sub> pollution, China has implemented several air pollutant control  
78 plans. The Air Pollution Prevention and Control Action Plan (APPCAP) launched in  
79 2013 aimed to reduce urban PM<sub>10</sub> levels by 10% and decrease the PM<sub>2.5</sub>  
80 concentrations of the Beijing-Tianjin-Hebei (BTH), the Pearl River Delta (PRD), and  
81 the Yangtze River Delta (YRD) by 15%-25% from 2012 to 2017 (Ma et al., 2020).  
82 APPCAP made an outstanding achievement and reduced PM<sub>2.5</sub> concentrations by over  
83 30% in China (Zhang et al., 2018). Huang et al. (2018) assessed the health effects of  
84 the APPCAP in 74 cities of China and found 47,240 fewer mortalities and 710,020

85 fewer years of life lost in 2017 compared with 2013. Lu et al. (2019) conducted a  
86 provincial-level evaluation of the health impacts of PM<sub>2.5</sub> after implementing the  
87 APPCAP in China, and found that deaths decreased from 1,078,800 of 2014 to  
88 962,900 of 2017. The health cost benefits were estimated to be the USD 193,800 in  
89 2017. However, the BTH region is still the most over-polluted area in China. In 2017,  
90 the annual average PM<sub>2.5</sub> concentration in BTH was 58-67 µg/m<sup>3</sup>, much higher than  
91 the WHO standard (CNEMC, 2018). In 2018, the Chinese government launched a  
92 three-year plan, the Blue Sky Protection Campaign (BSPC), to control air pollution in  
93 BTH, YRD, and Fenwei Plain. One of the specific goals is to reduce PM<sub>2.5</sub>  
94 concentration by 18% compared with the 2015 level. As the BTH region is one of the  
95 most developed regions and one of the regions suffering from the most severe air  
96 pollution in China (Hao et al., 2018; Qi et al., 2017; Yang et al., 2019), it has been  
97 identified as a priority area for air pollution control and prioritized an investigation of  
98 the policy impacts. Systematic analysis of the BSPC impacts on air quality and air  
99 pollutant emissions in the BTH region has been evaluated in our previously published  
100 study (Xu et al., 2020). However, the study only focused on the environmental  
101 impacts of implementing the BSPC in BTH rather than the health and economic  
102 impacts. In addition, most previous studies have concentrated on the health and  
103 economic improvement about the APPCAP. Few studies have involved a system  
104 evaluation of the health and economic effects due to implementing the BSPC.

105 The previous studies mainly focused on the one-city level (Wu et al., 2017),  
106 individual-sector level (Zhang et al., 2019), or the whole country (Maji et al., 2017).  
107 We provide the first comprehensive systematic assessment on the health and  
108 economic impacts due to implementing the BSPC in the BTH region to the best of our  
109 knowledge. This study aims to solve two key questions. Firstly, this study focuses on  
110 evaluating the health and economic effects of BSPC by 2030 in the BTH region.  
111 Secondly, this study will analyze the costs and benefits of the BSPC toward 2030 in  
112 the BTH region. It contributes to these existing studies in three aspects. (1)  
113 Quantifying the health impacts of the BSPC in the BTH region. (2) Evaluating the

114 economic effects containing health expenditures and the VSLs. (3) Exploring the  
115 cost-benefits ratios of the clean air policies. This study will make a certain  
116 contribution to demonstrate the feasibility and necessity of the BSPC in BTH and  
117 provide a reference for policy-makers.

118 This paper is organized as follows. Section 2 describes the methodology. Section  
119 3 presents the results, including health effects, economic effects, and a cost-benefit  
120 analysis. Section 4 presents the discussion, and section 5 summarizes the conclusions.

121

## 122 **2 Methodology**

### 123 *2.1 Integrated assessment methods*

124 This study uses an integrated assessment approach by combining the Greenhouse  
125 Gas and Air Pollution Interactions and Synergies IV Asia (GAINS IV Asia) model and  
126 the Integrated Model of Energy, Environment and Economy for Sustainable  
127 Development/Health (IMED/HEL) model to evaluate the health and economic  
128 impacts of implementing the BSPC in the BTH region (Fig.1). This approach has  
129 been used in our previous studies (Xie et al., 2016a; 2016b; 2018; 2019; 2020). Firstly,  
130 we use the GAINS IV Asia model to predict the PM<sub>2.5</sub> concentration reductions and  
131 the control technologies cost due to the implementation of BSPC. The IMED/HEL  
132 model evaluates the health and economic effects of PM<sub>2.5</sub> pollution, including  
133 morbidity, mortality, work time loss, medical expenditures, and VSL.

#### 134 *2.1.1 The GAINS IV Asia model*

135 The GAINS IV Asia model is used to evaluate the effects of various control and  
136 policies. It takes into account the activity pathways (for example, power and heating  
137 plants, industry and transport) and air pollution control measures for different  
138 pollutants from different sectors at a five-year interval (Amann et al., 2011) and  
139 quantifies emissions and implementing costs. This study integrates and calibrates  
140 original parameters into the GAINS IV Asia model. Specifically, parameters such as  
141 energy consumption by fuel types and sectors, industrial activities, the utilization rates  
142 of air pollution control technology, and other values and assumptions are considered.

143 According to the policy restrictions, we change the scenario assumptions in the model  
144 to assess changes in pollutant emissions, PM<sub>2.5</sub> concentrations, and emission reduction  
145 costs under various scenarios. The effects of BSPC on energy structures, pollutant  
146 emissions, PM<sub>2.5</sub> concentrations, and control costs have been evaluated in our recent  
147 study (Xu et al., 2020). More details regarding the calculation principles, descriptions  
148 of the input parameters, calculation process, and parameters determinations for BSPC  
149 are shown in section 1 of supporting information.

### 150 2.1.2 The IMED/HEL model

151 The IMED/HEL model evaluates the health effects due to PM<sub>2.5</sub> exposure on six  
152 morbidity endpoints and chronic mortality using a concentration-response function  
153 (CRF) (Apte et al., 2015; Tian et al., 2018; Wu et al., 2017; Xie et al., 2016a; Xie et  
154 al., 2018). Besides, it can also quantify the monetary value of these health endpoints.  
155 The IMED/HEL model is continuously updated, and the documents of detailed  
156 information on the model are available online  
157 (<http://scholar.pku.edu.cn/hanchengdai/imedhel>). In our study, a non-linear  
158 concentration-response function is applied to assess the six morbidity endpoints,  
159 premature deaths with different causes (ischemic heart disease, stroke,  
160 cerebrovascular disease, chronic obstructive pulmonary disease, lung cancer, and  
161 acute lower respiratory infection), and work time loss. Work time loss includes  
162 premature deaths between 15–65 and morbidity for labor. This study uses the  
163 population data from the sixth census to conduct the calculation. Besides, to monetize  
164 the health effects, medical expenditures and the VSL losses attribute to PM<sub>2.5</sub>  
165 pollution are quantified by assuming a linear relationship with the PM<sub>2.5</sub>-related  
166 premature deaths by using health service price (Xie et al., 2016a). Health expenditures  
167 are calculated by multiplying outpatient and hospital admission prices with total  
168 health endpoints. VSL are evaluated through the willingness to pay method in China  
169 (Jin et al., 2020), calculated by certain regions' current GDP per capita values relative  
170 to the national average per capita GDP in 2010 and the income elasticity of 0.5. The  
171 model is set about our previous analyses (Wu et al., 2017; Xie et al., 2016a; Xie et al.,

172 2020). Details of the calculation principles and calculation process can be found in  
173 section 2 in the supporting information.

### 174 *2.1.3 Cost-benefit analysis*

175 This study also estimates the net benefits of implementing BSPC in the BTH  
176 region. The monetized benefits include health expenditure savings associated with  
177 reduced morbidities and economic savings due to avoided premature deaths. Total  
178 costs equal to emission control costs from BPSA estimated by the GAINS IV Asia  
179 model, including the implementation costs of all pollutant control measures (i.e.,  
180 upfront investments and operating costs).

### 181 *2.2 Scenarios setting*

182 GAINS IV Asia model simulates emission scenarios at five-year intervals. The  
183 baseline is 2015. The primary target year of BPSA is 2020. The years 2025 and 2030  
184 are also included as the secondary target years to show the long-term health and  
185 economic benefits of air quality improvements.

186 Two scenarios are developed to simulate policy effects: the baseline scenario and  
187 the policy scenario. The baseline scenario assumes there is no BPSA implementation  
188 for air quality improvement in the BTH region. Firstly, we obtain the original 2015  
189 data from the World Energy Outlook 2018 Current Policy Scenario (WEO-2018-CPS)  
190 projected by the GAINS IV Asia model and then update parameters of the default  
191 baseline scenario of 'WEO-2018-CPS'. The key updating parameters have been  
192 described in section 1.4.3 in supporting information. Besides, it's worth noting that  
193 the GAINS IV Asia model has some inaccuracies in its data on renewable energy for  
194 2015, we calibrate it using data from the renewable energy report for China and the  
195 BTH region. Secondly, based on the recalibrated data for 2015, we project the  
196 activities of energy, agricultural, and industrial processes for 2020 based on the trend  
197 from the 'WEO-2018-CPS' and the 13th Five-Year Plan (FYP) targets of some macro  
198 economy indexes published by the BTH regional government and China's renewable  
199 report (NDRC 2016, NEA 2017). The energy demands in 2025 and 2030 in the  
200 baseline scenario are estimated based on the growth rate in the scenarios from 2020 to

201 2030 of the ‘WEO-2018-CPS’. The policy scenario assumes that the BTH region  
202 implemented BSPC. The key sub-policy package of BSPC can be found in Table S1  
203 in supporting information. Specifically, the policy packages are divided into four  
204 categories: the power and industry sector (with core content such as improving the  
205 industrial structure and the associated distribution for steel, cement, coke, glass, and  
206 coal power plants); the transport sector (with core content such as increasing the  
207 number of new energy vehicles and implementing national vi (B) standard gasoline  
208 and diesel for traffic vehicles); the building sector (with core content such as replacing  
209 untreated coal of heating by households); and the cross-sector (such as the pollution  
210 controls for stationary sources). The policy packages of the BSPC and its  
211 corresponding emission control measures in the GAINS IV Asia model are  
212 summarized in Table S2 in the supporting information. From 2020 to 2030, each  
213 region of the BTH will maintain the same policy strictness levels, and the  
214 governments will continue to implement specific policy measures at the same rate.  
215 The setting of scenarios refers to our previous studies (Meng et al., 2020). The  
216 detailed parameters and default parameters of the GAINS IV Asia model used in our  
217 study are shown in sections 1.4.3 and 1.4.4 of the supporting information.

### 218 **2.3 Data sources**

219 Data for this study was obtained from the China Guidebook for Air Pollution  
220 Emission Inventory (MEEC 2015), the Provincial Economic Yearbooks (BSB 2016,  
221 HSB 2016, TSB 2016), the China Statistical Yearbook (NBS 2016), the 13th FYP of  
222 Energy Development (NEA 2017a), the Clean Heating Plan for Winter in North  
223 China (2017–2021) (NEA 2017b), the 13th FYP of Renewable Energy Development  
224 (NDRC 2016), the 13th FYP of Industrial Transformation and Upgrading (GOHB  
225 2016), the 13th FYP for the Comprehensive Development of Transportation Systems  
226 (GOBJ 2016, GOHB 2016, GOTJ 2016), the 13th FYP of Power Sector Development  
227 (PGC 2016), and several state-of-the-art studies (Xiong *et al* 2015, Zhang *et al* 2015,  
228 Su *et al* 2018).

229

## 230 **3 Results**

### 231 **3.1 $PM_{2.5}$ concentration**

232 Table 1 shows the projected  $PM_{2.5}$  concentrations under different scenarios,  
233 which have been concluded in our previous study (Xu et al., 2020). During the last  
234 study,  $PM_{2.5}$  concentrations up to 2020 were quantified, and no longer-term  $PM_{2.5}$   
235 changes have been investigated. In this study, we extend it to 2030 to better evaluate  
236 the health and economic impacts.  $PM_{2.5}$  concentrations projected here are used as the  
237 basis for subsequent studies focusing on health and economic impacts. The results  
238 show that the emission of pollutants in Beijing in 2015 was not high, but Beijing and  
239 Tianjin are adjacent to Hebei, and a large number of pollutants disperse from Hebei to  
240 Beijing and Tianjin area. Also, the particular geographical locations of Beijing and  
241 Tianjin lead to the poor diffusion of air pollutants, so the  $PM_{2.5}$  concentrations of  
242 Beijing and Tianjin in 2015 were higher than that of Hebei. This study projected that  
243 the  $PM_{2.5}$  concentrations were  $78.7\mu\text{g}/\text{m}^3$ ,  $69.1\mu\text{g}/\text{m}^3$ , and  $65.2\mu\text{g}/\text{m}^3$  in 2015 for  
244 Beijing, Tianjin, and Hebei. According to the Environment Statement of 2015 in the  
245 BTH region (BJES, 2015; HBES 2015; TJES, 2015),  $PM_{2.5}$  concentrations were  
246  $80.6\mu\text{g}/\text{m}^3$ ,  $70\mu\text{g}/\text{m}^3$ , and  $77\mu\text{g}/\text{m}^3$  for Beijing, Tianjin, and Hebei, respectively. Our  
247 simulation for Beijing and Tianjin is comparable to the results from the Environment  
248 Statement.  $PM_{2.5}$  concentrations for Hebei is lower than Environment Statement.  
249 Because we aggregate and average the grid concentrations belonging to the region's  
250 longitude and latitude and the spatial resolution of concentration and area of the  
251 region may cause some differences in the results. Our projections are all comparable  
252 to the results from Zhang et al. (2019), which calculated the  $PM_{2.5}$  concentrations for  
253 Beijing, Tianjin, and Hebei in 2015 to be  $85\mu\text{g}/\text{m}^3$ ,  $72\mu\text{g}/\text{m}^3$ , and  $69\mu\text{g}/\text{m}^3$ ,  
254 respectively.

255 After implementing the BSPC,  $PM_{2.5}$  concentration in the BTH region will  
256 decrease significantly. By 2030, Tianjin will have the highest concentration of  $PM_{2.5}$ ,  
257 followed by Hebei and Beijing. Compared with the baseline scenarios,  $PM_{2.5}$   
258 concentration in policy scenarios will decrease by 8.7% (Hebei 2030) to 32.1%  
259 (Beijing 2020). However,  $PM_{2.5}$  concentration under the policy scenarios will still be

260 higher than the national standard of  $35 \mu\text{g}/\text{m}^3$  and will still have impacts the exposed  
261 population.

## 262 **3.2 Health impacts**

### 263 *3.2.1 Morbidity*

264 The main morbidities associated with  $\text{PM}_{2.5}$  pollution involve asthma,  
265 cardiovascular hospital admissions, cerebrovascular hospital admissions, chronic  
266 bronchitis, respiratory hospital admissions, and upper respiratory (Dockery et al.,  
267 1993; Iii et al., 2002; Pope et al., 2002), which are the primary health endpoints we  
268 evaluate in this study. Fig. 2A1-A3 (left Y-axis) presents the physical impacts on  
269 morbidities attributed to  $\text{PM}_{2.5}$  pollution in BTH. In 2015, the total morbidities due to  
270  $\text{PM}_{2.5}$  exposure amounted to 5208 thousand cases in BTH. The morbidity cases show  
271 a decreasing trend in both the baseline and the policy scenarios, and the more  
272 significant  $\text{PM}_{2.5}$  decreases under the policy scenario are more effective in reducing  
273 morbidity cases. By 2020, morbidity cases in the BTH region under the policy  
274 scenario will decrease significantly to 3561 (equivalent to a 24% reduction in the  
275 baseline scenario) thousand cases. It will reduce to 3242 (25%) and 2943 (18%)  
276 thousand cases by 2025 and 2030. Upper respiratory is the dominating morbidity  
277 endpoint, followed by asthma and chronic bronchitis. These three diseases account for  
278 more than 96% of total morbidity in all scenarios.

279 At the regional level, provinces with higher population density suffer more  
280 morbidities (Xie et al., 2019). Beijing presented the highest  $\text{PM}_{2.5}$  concentration in  
281 2015. Hebei had the highest  $\text{PM}_{2.5}$ -related morbidity cases, accounting for 68% of  
282 morbidity cases in BTH in 2015. Hebei has higher morbidity risks due to the aging  
283 population. Specifically, the aging population of Hebei has exceeded the national  
284 average level for the first time since 2015 (Li, 2019). By the end of 2017, it accounted  
285 for 18% of the total population in Hebei. The aging population will reach 15 million  
286 by 2020, indicating a moderately aging society on the horizon. The bars in Fig. 3  
287 indicate the absolute reductions (thousand cases) in morbidity between the baseline  
288 scenarios and the policy scenarios for each year. After implementing the BSPPC, 378

289 (38%), 105 (20%), and 644 (20%) thousand cases of morbidity will be avoided by  
290 2020 in Beijing, Tianjin, and Hebei, respectively. The avoided morbidities cases will  
291 amount to 266 (34%), 141 (29%), and 268 (11%) thousand by 2030. Zhang et al.  
292 (2019) evaluated the health benefits of the residential ‘coal-to-electricity’ policy in  
293 BTH. They found that with the implementation of the policy, 0.20 (19.7%), 0.07  
294 (13.1%), and 0.40 (11.2%) million cases of morbidity could be avoided in BTH,  
295 respectively. The avoided morbidity is higher than other studies of other clean air  
296 policies. Beijing improves the most in terms of  $PM_{2.5}$  concentrations and the  
297 proportions of reductions in morbidity cases. Hebei benefits the most in total  
298 morbidity population due to its large exposure population.

299 The number of morbidity cases is lower in the policy scenarios than in the  
300 baseline scenarios due to implementing the BSPC. The morbidity cases decrease in  
301 the 2020 baseline scenario compared to 2015 because of the continuous decline in  
302  $PM_{2.5}$  concentrations (Fan et al., 2020; Li et al., 2020; Maji et al., 2020; Meng and  
303 Zhou, 2020). Similar reductions in morbidity cases without implementing the BPC  
304 are also found in other researches in the BTH region (Huang and Zhang, 2013; Zhang  
305 et al., 2019) and other provinces (Liyang et al., 2014; Wu et al., 2017). For instance,  
306 Huang and Zhang (2013) presented that significant health improvements would be  
307 achieved since the implementation of Air Quality Standards in 2012. Therefore, it  
308 indicates that even if BPC is not implemented (as suggested in the baseline scenario),  
309 the morbidity cases in the 2020 baseline scenario will be lower than that of 2015. The  
310 morbidity reductions in Tianjin are lower than in Beijing and Hebei due to  $PM_{2.5}$   
311 concentration reduction potentials are more significant than those of Tianjin under  
312 BPC.

### 313 3.2.2 Mortality

314 The red triangles in Fig. 2A1-A3 indicate the impacts on mortalities associated  
315 with  $PM_{2.5}$  exposure in BTH. The mortalities were estimated to be 206 thousand in  
316 the BTH region of 2015. There will be slight downward trends of the mortality cases  
317 in both the baseline and policy scenarios. Mortality cases will significantly decrease

318 in the policy scenarios due to reductions in  $PM_{2.5}$  pollution under the BSPPC.  
319 Mortalities will amount to 204 thousand and 180 thousand in the baseline and policy  
320 scenarios, respectively. BSPPC will reduce 24 thousand premature deaths in BTH in  
321 2020. Mortality in BTH under the policy scenarios will decrease significantly by 20  
322 (12%) thousand and 20 (9%) thousand by 2025 and 2030, respectively.

323 It has been concluded that regions with a larger population have more mortalities  
324 (Xie et al., 2019). In Hebei, the mortality cases were 145 thousand in 2015. The  
325 yellow triangles in Fig. 3 indicate the absolute reductions (thousand cases) in  
326 mortality between the baseline scenarios and the policy scenarios for each year.  
327 Compared with the baseline scenario, the BSPPC will avoid approximately 14 (10%)  
328 thousand premature deaths by 2020 under the policy scenario. It will decrease by 16  
329 (11%) thousand and 8 (5%) thousand cases by 2025 and 2030. In Beijing and Tianjin,  
330 after implementing the BSPPC, the mortality cases will decrease by 8 (19%) thousand  
331 and 2 (9%) thousand by 2020. The avoided mortality in Beijing and Tianjin is 6.6  
332 (17%) thousand and 3.2 (14%) thousand cases of mortalities by 2030. Many studies  
333 have assessed premature mortality due to  $PM_{2.5}$  exposure. If the residential “coal-to  
334 electricity” policy in BTH is fully implemented, 6700, 2500, and 13200 cases of  
335 premature deaths can be avoided in 2020 (Zhang et al., 2019). Huang et al. (2018)  
336 assessed the health effects of the APPCAP in 74 cities of China and found 47,240  
337 fewer mortalities in 2017 compared with 2013. Our study shows implementing BSPPC  
338 can significantly reduce morbidity in the BTH. The avoided premature deaths in  
339 Hebei is much higher than in Beijing and Tianjin. Beijing benefits the most from  
340 BSPPC in the percentage of reductions in mortality cases.

341 The mortality cases in the 2020 baseline scenario will be comparable or even  
342 higher than the 2015 level. A slight decrease of  $PM_{2.5}$  concentration between 2015 and  
343 the 2020 baseline scenario will have little impact on the mortality cases. The results  
344 could also explain this from Wang et al. (2014). These volunteers with longer  
345 exposure histories present higher contaminant levels in serum while not cause acute  
346 deaths (Wang et al., 2014). The declining trends in mortality cases between the

347 reference year and the baseline scenario are consistent with the conclusions drawn  
348 from Zhang et al. (2019). Therefore, it demonstrates that even if BPSC is not  
349 implemented, the mortality cases in the 2020 baseline scenario will be even higher  
350 than that of 2015. Consequently, it is necessary to implement the BPSC to achieve  
351 health benefits to reduce mortality cases.

### 352 *3.2.3 Work time loss*

353 Morbidities and mortalities lead to work time loss, as indicated by the bars below  
354 the X-axis in Fig. 2B1-B3. The purple circles in Fig. 3 indicate the absolute  
355 reductions (hour/cap/year) in work time loss between the baseline scenarios and the  
356 policy scenarios. The work time loss is calculated as the hours a person aged 15-64  
357 stays off work due to morbidities and mortalities. In 2015, the total annual work time  
358 loss in BTH was  $7.7 \times 10^8$  hours; this will drop to  $6.9 \times 10^8$  and  $5.3 \times 10^8$  hours by 2020  
359 and 2030 under the baseline scenarios. Under the policy scenarios, work time loss will  
360 be approximately  $5.2 \times 10^8$  hours and  $4.3 \times 10^8$  hours by 2020 and 2030. Work time loss  
361 presents a downward trend in the BTH region due to air quality improvement under  
362 all scenarios.  $PM_{2.5}$  decrease has a positive effect in decreasing work time loss by  
363  $1.7 \times 10^8$  hours (24%) in the BTH region by 2020.

364 At the provincial level, the annual per capita work time loss in 2015 was 11.98  
365 hours, 10.30 hours, and 9.63 hours in Beijing, Tianjin, and Hebei. Under the policy  
366 scenario, per capita work time loss in 2020 will drop to 6.69 hours (38%) in Beijing,  
367 8.11 hours (20%) in Tianjin, and 6.91 hours (20%) in Hebei. The per capita work loss  
368 will decrease to 5.52 hours (34%), 6.76 hours (29%), and 5.92 hours (11%). We  
369 calculated the percentages in the brackets as the absolute avoided work time loss ratio  
370 between the baseline scenario and the policy scenario to the absolute value of work  
371 time loss in the corresponding baseline scenario, representing the percentage of the  
372 avoided absolute work time loss caused by implementing the BPSC in the baseline  
373 scenario. Tian et al. (2018) evaluated  $PM_{2.5}$  pollution-related health impacts of road  
374 transport in China. They found the most strict control strategy scenario would  
375 decrease the hours of work time loss by 42.65% in China. Their result is comparable

376 to our study. In 2015, Beijing lost the most working hours in 2020. The regional  
377 disparities in work time loss are consistent with the regional disparities in PM<sub>2.5</sub>  
378 concentrations.

### 379 **3.3 Economic impacts**

#### 380 *3.3.1 Health expenditures*

381 PM<sub>2.5</sub> pollution also causes additional medical expenditures, as indicated by the  
382 bars above the X-axis in Fig. 2B1-B3. In 2015, additional medical expenses related to  
383 air exposure in the BTH region amounted to 134 million USD. In the baseline  
384 scenarios, the additional health expenditure for the BTH region will be 121 million  
385 USD and 92 million USD by 2020 and 2030. It will drop to 91 million USD and 75  
386 million USD under the policy scenarios by 2020 and 2030. The expenditures present  
387 slightly downward trends in the BTH region due to improved air quality under all  
388 scenarios. However, the reductions in policy scenarios are more significant than their  
389 respective baseline. The monetized benefits of avoided morbidity cases are 29 million  
390 USD and 17 million USD in the BTH region, accounting for 24% and 18% of their  
391 perspective baseline expenditures. The costs of treating respiratory symptoms are  
392 relatively low. Because of the large number of cases, respiratory symptoms are the  
393 first contributor to total morbidity expenditures, accounting for more than 50% of  
394 total expenditure in all the scenarios.

395 At the regional level, the additional medical expenditures in 2015 were 29  
396 million USD, 14 million USD, and 91 million USD for Beijing, Tianjin, and Hebei,  
397 respectively. Under the policy scenario, the additional medical expenditures in 2020  
398 will drop to 16 million USD (38%) in Beijing, 11 million USD (20%) in Tianjin,  
399 and 65 million USD (20%) in Hebei. The expenditures will be even lower by 2030,  
400 which will decrease to 13 million USD (34%), 8.9 million USD (29%), and 53 million  
401 USD (11%), respectively. Hebei encounters the highest additional medical spending in  
402 the BTH region and benefits the most in reducing health expenditures. Zhang et al.  
403 (2019) assessed the monetized benefits of morbidity reduction under the  
404 implementation of the coal-to-electricity policy in the residential sector in BTH. They

405 concluded that the morbidity benefits of reducing morbidity cases from PM<sub>2.5</sub>  
406 pollution are 0.23, 0.07, and 0.26 billion yuan in the BTH region, respectively, by  
407 2020, accounting for 22%, 18.2%, and 11.1% of their respective morbidity losses. The  
408 reduction of PM<sub>2.5</sub> concentration due to implementing the BSPC will help reduce the  
409 additional medical expenditures in the BTH region. The decreasing trends of  
410 additional medical spending on both the baseline scenarios and the policy scenarios  
411 are consistent with the declining trends of morbidity cases in different scenarios. Our  
412 result is consistent with the conclusions drawn from Xie et al. (2016b), which  
413 indicated that the medical expenditures would be lower in the scenarios with low  
414 PM<sub>2.5</sub> concentration.

### 415 3.3.2 VSL

416 The benefits of avoided deaths are monetized using VSL, shown by the scatter  
417 triangles in Fig. 2B1-B3. Mortality losses are significantly higher than morbidity  
418 losses. Mortality losses in 2015 were 395 billion USD in the BTH region. In the  
419 baseline scenarios, the mortality losses due to uncontrolled PM<sub>2.5</sub> pollution will  
420 increase to 473 billion USD and 624 billion USD in the BTH region by 2020 and  
421 2030. The mortality losses of implementing the BPSC will increase to 413 billion  
422 USD and 561 billion USD by 2020 and 2030. The benefits will be 60 billion USD and  
423 63 billion USD from reducing PM<sub>2.5</sub> concentration in 2020 and 2030. With the per  
424 capita GDP and income increase, total VSL will increase significantly in the future.

425 At the regional level, VSL is significantly higher in Hebei than in Beijing and  
426 Tianjin, which was estimated to be 198 billion USD, 125 billion USD, and 72 billion  
427 USD in 2015, respectively. Under the policy scenarios, VSL in 2020 will increase to  
428 81 billion USD (10%) and 222 billion USD (10%) in Tianjin and Hebei, with an  
429 absolute increase of 9 billion USD and 24 billion USD. By 2030, VSL will be 105  
430 billion USD and 331 billion USD under Tianjin and Hebei's policy scenarios. For  
431 Beijing, VSL also presents an increasing trend in both the baseline and policy  
432 scenarios. However, in the policy scenario, the VSL in 2020 will be lower than that in  
433 2015. It will present an upward trend from 2020 to 2030. Some studies have

434 monetized the benefits of mortality reduction due to PM<sub>2.5</sub> exposure. Zhang et al.  
435 (2019) suggested that if the “coal-to electricity” policy is fully implemented in BTH  
436 by 2020, there will be 20.10, 4.84, and 15.60 billion yuan in health benefits for  
437 Beijing, Tianjin, and Hebei, respectively. The health benefits of VSL in our study are  
438 higher than that study, and BSPC in BTH has higher economic benefits in our study.  
439 Our analysis also indicates that the economic benefits are dominated by VSL savings  
440 (avoided premature deaths) that far exceed the health expenditure savings. Our result  
441 is consistent with other studies (Garcia-Menendez et al., 2015; Turner et al., 2015;  
442 West et al., 2012).

### 443 *3.4 Cost-effectiveness analysis*

444 Our results indicate that the BSPC entails co-benefits in reducing morbidities,  
445 mortalities, work time loss, medical expenditures, and VSL. A cost-benefit evaluation  
446 of PM<sub>2.5</sub>-polluted health effects due to implementing the BSPC is shown in Fig. 4 and  
447 Fig. 5. The emission control costs are estimated using the GAINS IV Asia mode. The  
448 costs under different scenarios of each region are shown in Table 2 below. The health  
449 benefit mainly comprises two aspects of health expenditure savings attributed to  
450 reductions in morbidity and reductions in VSL associated with decreased mortality.  
451 VSL savings dominate the economic benefits. Therefore, the detailed benefit-cost  
452 evaluation of the BSPC is conducted by comparing the monetized health benefits  
453 (including VSL-related monetary savings) with pollutant terminal technology costs.

454 A closer look at the provincial level shows that the costs in policy scenarios are  
455 lower than those in the baseline scenarios from each region. The technological  
456 improvements in the BSPC will improve air quality and reduce the control costs  
457 simultaneously. In 2020, the cost and benefit ratio will amount to 94, 70, and 12 of  
458 Beijing, Tianjin, and Hebei, respectively. By 2030, the  $\Delta$  benefit/ $\Delta$  cost ratios are  
459 about 502, 85, and 8 in Beijing, Tianjin, and Hebei, respectively. Beijing has higher  
460 cost-effectiveness of BSPC. Hebei has a relatively lower ratio due to a large number  
461 of pollutant emissions and reduction pressure. Our result is consistent with the  
462 conclusions drawn from Xie et al. (2016b). They found the benefit-cost ratios in the

463 BTH region will be the highest in Beijing, followed by Tianjin and Hebei in 2020.

464 The benefit-cost ratios will increase from 2020 to 2030 in Beijing because of the  
465 cumulative effects (shown in Fig. 5), presenting increasing benefits obtained through  
466 implementing control measures. For Hebei, it is less significant due to its  
467 industrialization processes and the aging population society. Further, the efficiencies  
468 of pollutant controls under the BPSA in Beijing are higher than those of Tianjin and  
469 Hebei due to different sources of air pollutant emissions in the BTH region. Therefore,  
470 to effectively control air pollution, the BTH region needs unified planning and joint  
471 control.

472

#### 473 **4 Discussion**

474 Our results clearly show that the BPSA can significantly benefit air quality and  
475 health improvement in BTH. In 2015, Beijing presented the highest  $PM_{2.5}$   
476 concentration, followed by Tianjin and Hebei, comparable with previous studies  
477 (Zhang et al., 2019). After implementing the BPSA, in 2020, Tianjin will have the  
478 highest  $PM_{2.5}$  concentration, followed by Hebei and Beijing, with  $55.1 \mu g/m^3$ ,  $48.4$   
479  $\mu g/m^3$ , and  $47.1 \mu g/m^3$ , respectively. Even in 2030, the average annual  $PM_{2.5}$   
480 concentrations from Beijing, Tianjin, and Hebei in policy scenarios far exceed the  
481 national standard of  $35 \mu g/m^3$ . BPSA is not enough to achieve air quality standards in  
482 the BTH region. More attention and efforts are needed to achieve better air quality.

483 Air pollution will inevitably pose severe health impacts on people (Dockery et al.,  
484 1993 ; Iii et al., 2002; Pope et al., 2002). The health and economic impacts of air  
485 pollution cannot be ignored (Wu et al., 2017). The results show that the  $PM_{2.5}$   
486 concentration reductions due to implementing BPSA will reduce morbidity cases and  
487 health expenditures in BTH. After implementing the BPSA, by 2020, the reduced  
488 morbidity cases and health expenditures will be 378 thousand cases and 10 million  
489 USD in Beijing, 105 thousand cases and 2.7 million USD in Tianjin, and 644  
490 thousand cases and 16.6 million USD in Hebei. The provinces with higher population  
491 density and more heavy industry could benefit the health and economy after  
492 implementing the clean air policies (Xie et al., 2020). By 2030, these reduced

493 morbidity cases and health expenditures would be 266 thousand cases and 6.8 million  
494 USD in Beijing, 141 thousand cases and 3.6 million USD in Tianjin, and 260  
495 thousand cases and 6.7 million USD in Hebei, which are equivalent to 34.0 % and  
496 34.0%, 28.8% and 28.8%, and 11.1% and 11.1% of the 2015 levels in Beijing, Tianjin,  
497 and Hebei, respectively.

498 The most concerning health effects are mortalities (Zhang et al., 2019). In 2020,  
499 the reduced mortality cases and VSL due to implementing the BSPC will be 7.6  
500 thousand cases and 26.7 billion USD in Beijing, 2.1 thousand cases and 8.7 billion  
501 USD in Tianjin, and 14.3 thousand cases and 24.8 billion USD in Hebei. In 2030, the  
502 reduced mortality cases and VSL will be 6.6 thousand cases and 26.6 billion USD in  
503 Beijing, 3.2 thousand cases and 17.9 billion USD in Tianjin, and 6.7 thousand cases  
504 and 18.5 billion USD in Hebei, which are equivalent to 16.3 % and 21.2%, 15.4% and  
505 24.8%, and 4.7% and 9.4% of the 2015 levels for Beijing, Tianjin, and Hebei,  
506 respectively. It is worth noting that the absolute mortality cases have been keeping  
507 relatively stable from 2020 to 2030 under the policy scenarios, while the VSL has  
508 shown a significant increasing trend from 2020 to 2030 under the policy scenarios.  
509 Our results show that the benefits are dominated by VSL savings that far exceed the  
510 benefits in health expenditure savings, consistent with many other studies (West et al.,  
511 2013; Garcia-Menendez et al., 2015).

512 In 2020, the reduced work loss time due to implementing the BSPC will be 4.0  
513 hour/cap/year of Beijing, 2.0 hour/cap/year of Tianjin, and 1.8 hour/cap/year of Hebei,  
514 which are equivalent to 33.5%, 19.6%, and 18.4% of the 2015 levels in Beijing,  
515 Tianjin, and Hebei, respectively. In 2030, the reduced work loss time would be 2.8  
516 hour/cap/year of Beijing, 2.7 hour/cap/year of Tianjin, and 0.74 hour/cap/year of  
517 Hebei, which are equivalent to 23.7%, 28.8%, and 7.7% of the 2015 levels in Beijing,  
518 Tianjin, and Hebei, respectively.

519 On the other hand, controlling air pollution needs a lot of capital and technology  
520 investments. The costs in our study are calculated based on the GAINS IV Asia model,  
521 which represents the costs of terminal treatment technology (Amann et al., 2011).

522 Overall, compared with the 2015 level, the pollutant emission controls cost in 2020  
523 and 2030 will be about 1.3 times and 2.4 times higher for Beijing, 1.4 times and 1.9  
524 times higher for Tianjin, 1.4 times and 1.6 times higher for Hebei, respectively.

525 Based on the consideration of health benefits and investments in 2020, the  
526 benefit-cost ratios of air pollution control in BTH will be the highest in Beijing (94),  
527 followed by Tianjin (70) and Hebei (12). Similar trends are for the year 2030. Our  
528 results reveal the benefits of pollution controls in Beijing are significantly higher than  
529 those in Tianjin and Hebei. This is in line with a prior study which suggests that the  
530 benefit-cost ratios of air pollution controls in BTH in 2020 is the highest in Beijing,  
531 followed by Tianjin, and Hebei is the lowest (Xie et al., 2016a). The formulation of  
532 BSPC is most beneficial to Beijing, but the net benefit to Hebei is low. But this does  
533 not mean that pollution in Hebei will not be controlled. To effectively control air  
534 pollution, the BTH region needs unified planning and joint control. Cooperation with  
535 the surrounding areas to actively introduce air pollution control technology should be  
536 encouraged (Xie et al., 2020).

537

## 538 **5 Conclusion**

539 In this study, we combine GAINS IV Asis model and IMED/HEL model to  
540 quantify the benefits of implementing the BSPC in the BTH region. Our integrated  
541 assessment reveals that the BSPC can significantly benefit air quality and health  
542 improvement in the BTH region. Substantial reduction of mortalities and morbidity  
543 related to  $PM_{2.5}$  will be achieved in the BTH region, and Hebei will have the majority  
544 in Hebei.  $PM_{2.5}$  reductions due to BSPC are highly effective in reducing work time  
545 loss.

546 We also conduct a detailed benefit-cost evaluation of the BSPC in the BTH  
547 region by comparing the monetized health benefits (VSL-related savings) with  
548 pollutant control technology costs. Substantial health gains can be obtained due to  
549  $PM_{2.5}$  reductions, and the benefits can offset the control costs in the BTH region. It  
550 presents a solid framework for evaluating the effectiveness of control measures from a

551 public health perspective. The applicability can be broadened to other areas to assist  
552 in policy-making and determine the strategies needed to achieve ambitious air quality  
553 objectives.

554 There are some limitations to this study. The BSPC has co-benefits on CO<sub>2</sub>  
555 mitigation, which have been identified by our previous analysis (Meng et al., 2020).  
556 Therefore, the CO<sub>2</sub> reduction due to the implementation of BSPC may have other  
557 benefits. Secondly, we only consider the effects on labor supply for the health  
558 evaluation, while different conclusions have shown that PM<sub>2.5</sub> pollution can influence  
559 labor productivity. If we can assess productivity effects in the future, the benefits will  
560 be higher than our calculation. However, our conclusion will still hold if we take into  
561 account productivity effects. Thirdly, we ignore the impact of indoor PM<sub>2.5</sub> pollution.  
562 Therefore, future researches are necessary to achieve comprehensive policy insights.

563

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567

#### 568 **References**

- 569 Amann M, Bertok I, Borcken-Kleefeld J, Cofala J, Heyes C, Höglund-Isaksson L, et al.,  
570 2011. Cost-effective control of air quality and greenhouse gases in Europe:  
571 Modeling and policy applications. *Environmental Modelling & Software*. 26,  
572 1489-1501.
- 573 Anenberg SC, Horowitz LW, Tong DQ, West JJ., 2010. An Estimate of the Global  
574 Burden of Anthropogenic Ozone and Fine Particulate Matter on Premature  
575 Human Mortality Using Atmospheric Modeling. *Environmental Health*  
576 *Perspectives*. 118, 1189-1195.
- 577 Apte JS, Marshall JD, Cohen AJ, Brauer M., 2015. Addressing Global Mortality from  
578 Ambient PM<sub>2.5</sub>. *Environmental Science & Technology*. 49, 8057-8066.
- 579 Bai R, Lam JCK, Li VOK., 2018. A review on health cost accounting of air pollution

- 580 in China. *Environment International*. 120, 279-294.
- 581 BJES. 2015. Beijing Environmental Statement 2015. Beijing Municipal  
582 Environmental Protection Bureau.  
583 [http://sthjj.beijing.gov.cn/bjhrb/resource/cms/2018/04/2018042409544236953.](http://sthjj.beijing.gov.cn/bjhrb/resource/cms/2018/04/2018042409544236953.pdf)  
584 pdf
- 585 Brauer M, Amann M, Burnett RT, Cohen A, Dentener F, Ezzati M, et al., 2012.  
586 Exposure Assessment for Estimation of the Global Burden of Disease  
587 Attributable to Outdoor Air Pollution. *Environmental Science & Technology*.  
588 46, 652-660.
- 589 Brauer M, Freedman G, Frostad J, van Donkelaar A, Martin RV, Dentener F, et al.,  
590 2016. Ambient Air Pollution Exposure Estimation for the Global Burden of  
591 Disease. 2013. *Environmental Science & Technology*. 50, 79-88.
- 592 BSB 2016 Beijing Statistics Yearbook 2016 Beijing Statistical Bureau.
- 593 Cao J, Yang C, Li J, Chen R, Chen B, Gu D, et al., 2011. Association between  
594 long-term exposure to outdoor air pollution and mortality in China: A cohort  
595 study. *Journal of Hazardous Materials*. 186, 1594-1600.
- 596 CNEMC. China Environmental State Communique. China National Environmental  
597 Monitoring Center,  
598 [http://www.cnemc.cn/jcbg/zghjzkqb/201905/t20190529\\_704755.shtml](http://www.cnemc.cn/jcbg/zghjzkqb/201905/t20190529_704755.shtml), 2018.
- 599 Cohen AJ, Ross Anderson H, Ostro B, Pandey KD, Krzyzanowski M, Künzli N, et al.,  
600 2005. The Global Burden of Disease Due to Outdoor Air Pollution. *Journal of*  
601 *Toxicology and Environmental Health, Part A*. 68, 1301-1307.
- 602 Dockery DW, Pope CA, 3rd, Xu X, Spengler JD, Ware JH, Fay ME, et al., 1993. An  
603 association between air pollution and mortality in six U.S. cities. *N Engl J*  
604 *Med*. 329, 1753-9.
- 605 Dong H, Dai H, Dong L, Fujita T, Geng Y, Klimont Z, et al., 2015. Pursuing air  
606 pollutant co-benefits of CO<sub>2</sub> mitigation in China: A provincial leveled analysis.  
607 *Applied Energy*. 144, 165-174.
- 608 Evans J, van Donkelaar A, Martin RV, Burnett R, Rainham DG, Birkett NJ, et al.,

- 609 2013. Estimates of global mortality attributable to particulate air pollution  
610 using satellite imagery. *Environmental Research*. 120, 33-42.
- 611 Fan Y, Ding X, Hang J, Ge J., 2020. Characteristics of urban air pollution in different  
612 regions of China between 2015 and 2019. *Building and Environment*. 180,  
613 107048.
- 614 Garcia-Menendez F, Saari RK, Monier E, Selin NE., 2015. U.S. Air Quality and  
615 Health Benefits from Avoided Climate Change under Greenhouse Gas  
616 Mitigation. *Environmental Science & Technology*. 49, 7580-7588.
- 617 GOBJ 2016 13th Five-Year Plan for the Comprehensive Development of  
618 Transportation System in Beijing The People's Government of Beijing.  
619 <https://wenku.baidu.com/view/4c9d0d7351e79b89690226d0.html>
- 620 GOHB 2016 13th Five-Year Plan of Industrial Transformation and Upgrading in  
621 Hebei The People's Government of Hebei.  
622 <http://info.hebei.gov.cn/eportal/ui?pageId=1962757&articleKey=6672412&columnId=329982>  
623
- 624 GOTJ 2016 13th Five-Year Plan for the Comprehensive Development of  
625 Transportation System in Tianjin The People's Government of Tianjin.  
626 <https://max.book118.com/html/2018/0316/157508421.shtm>
- 627 Hao Y, Peng H, Temulun T, Liu L-Q, Mao J, Lu Z-N, et al., 2018. How harmful is air  
628 pollution to economic development? New evidence from PM<sub>2.5</sub> concentrations  
629 of Chinese cities. *Journal of Cleaner Production*. 172, 743-757.
- 630 HBES. 2015. Hebei Environmental Statement 2015. Hebei Municipal Environmental  
631 Protection Bureau.  
632 [http://hbepb.hebei.gov.cn/root8/auto454/201606/W020160613320961712384.](http://hbepb.hebei.gov.cn/root8/auto454/201606/W020160613320961712384.pdf)  
633 pdf
- 634 Hossain, M. S., Frey, H. C., Louie, P. K. K., and Lau, A. K. H., 2021. Combined  
635 effects of increased O<sub>3</sub> and reduced NO<sub>2</sub> concentrations on short-term air  
636 pollution health risks in Hong Kong. *Environmental Pollution*. 270, 116280.
- 637 HSB. 2016. Hebei Economic Yearbook 2016. Hebei Statistical Bureau, China

- 638 Statistics Press
- 639 Huang DS, Zhang S., 2013. Health benefit evaluation for PM<sub>2.5</sub> pollution control in  
640 Beijing-Tianjin-Hebei region of China. *China Environmental Science*. 33,  
641 166-174.
- 642 Huang J, Pan X, Guo X, Li G., 2018. Health impact of China's Air Pollution  
643 Prevention and Control Action Plan: an analysis of national air quality  
644 monitoring and mortality data. *The Lancet Planetary Health*. 2, 313-323.
- 645 Iii, C.A.P., Burnett, R.T., Thun, M.J., Calle, E.E., Krewski, D., Ito, K., Thurston, G.D.,  
646 2002. Lung cancer, cardiopulmonary mortality, and long-term exposure to fine  
647 particulate air pollution. *JAMA*. 287, 1132
- 648 Ji, W., Zhou, B., and Zhao, B., 2019. Potential reductions in premature mortality  
649 attributable to PM<sub>2.5</sub> by reducing indoor pollution: A model analysis for  
650 Beijing-Tianjin-Hebei of China. *Environmental Pollution*. 245, 260-271.
- 651 Jin, Y., Andersson, H., Zhang, S., 2020. Do preferences to reduce health risks related  
652 to air pollution depend on illness type? Evidence from a choice experiment in  
653 Beijing, China. *Journal of Environmental Economics and Management*, 103,  
654 102355.
- 655 Kuerban, M., Waili, Y., Fan, F., Liu, Y., Qin, W., Dore, A. J., Peng, J., Xu, W., and  
656 Zhang, F., 2020. Spatio-temporal patterns of air pollution in China from 2015  
657 to 2018 and implications for health risks. *Environmental Pollution*. 258,  
658 113659.
- 659 Lelieveld J, Evans JS, Fnais M, Giannadaki D, Pozzer A., 2015. The contribution of  
660 outdoor air pollution sources to premature mortality on a global scale. *Nature*.  
661 525,367-371.
- 662 Li MM., 2019. Analysis on the influencing factors of population aging in hebei  
663 province. *Fujian Quality Management*. 22:229.
- 664 Li W, Shao L, Wang W, Li H, Wang X, Li Y, et al., 2020. Air quality improvement in  
665 response to intensified control strategies in Beijing during 2013–2019. *Science*  
666 of The Total Environment. 744, 140776.

- 667 Lim SS, Vos T, Flaxman AD, Danaei G, Shibuya K, Adair-Rohani H, et al., 2012. A  
668 comparative risk assessment of burden of disease and injury attributable to 67  
669 risk factors and risk factor clusters in 21 regions, 1990–2010: a systematic  
670 analysis for the Global Burden of Disease Study 2010. *The Lancet*. 380,  
671 2224-2260.
- 672 Liu J, Han Y, Tang X, Zhu J, Zhu T., 2016. Estimating adult mortality attributable to  
673 PM<sub>2.5</sub> exposure in China with assimilated PM<sub>2.5</sub> concentrations based on a  
674 ground monitoring network. *Science of The Total Environment*. 568,  
675 1253-1262.
- 676 Liying MA, Zeqin D, Kejia WU, Jun P., 2014. Assessing the Healthy Benefits and  
677 Cost from Household Solid Fuel Intervention in Rural Guizhou.  
678 *Environmental Science & Technology*. 37, 112-117.
- 679 Lu X, Lin C, Li W, Chen Y, Huang Y, Fung JCH, et al., 2019. Analysis of the adverse  
680 health effects of PM<sub>2.5</sub> from 2001 to 2017 in China and the role of  
681 urbanization in aggravating the health burden. *Science of The Total  
682 Environment*. 652, 683-695.
- 683 MEEC 2015 China Guidebook for Air Pollution Emission Inventory 2015 The  
684 Ministry of Ecology and Environment of China.  
685 [http://www.mee.gov.cn/gkml/hbb/bgth/201603/t20160315\\_332883.htm](http://www.mee.gov.cn/gkml/hbb/bgth/201603/t20160315_332883.htm)
- 686 Ma X, Li C, Dong X, Liao H., 2020. Empirical analysis on the effectiveness of air  
687 quality control measures during mega events: Evidence from Beijing, China.  
688 *Journal of Cleaner Production*. 271, 122536.
- 689 Maji K J , Dikshit A K , Arora M , Deshpande A., 2017. Estimating premature  
690 mortality attributable to PM<sub>2.5</sub> exposure and benefit of air pollution control  
691 policies in China for 2020. *Science of the Total Environment*, 612:683-693.
- 692 Maji KJ, Li VOK, Lam JCK., 2020. Effects of China's current Air Pollution  
693 Prevention and Control Action Plan on air pollution patterns, health risks and  
694 mortalities in Beijing 2014–2018. *Chemosphere*. 260, 127572.
- 695 Meng M, Zhou J., 2020. Has air pollution emission level in the Beijing–Tianjin–

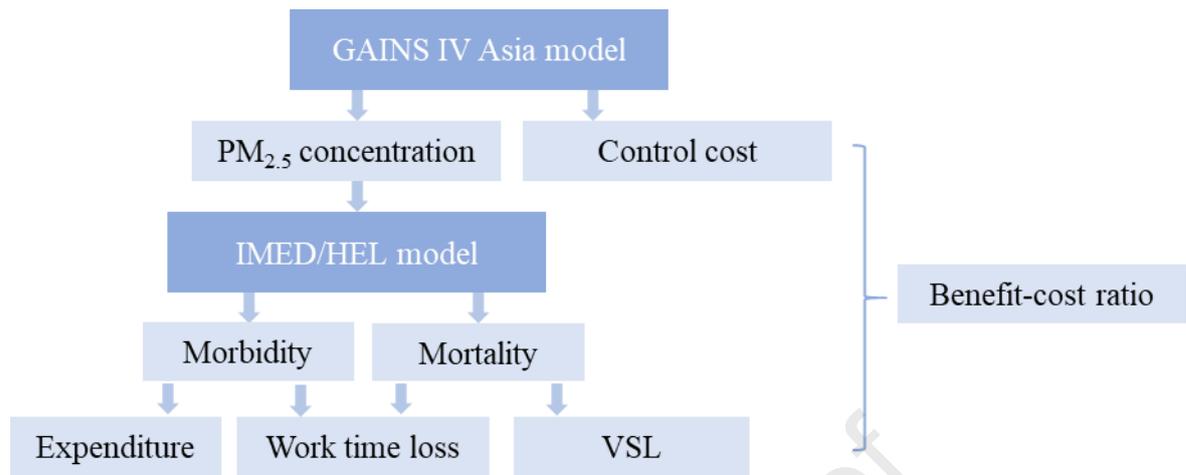
- 696 Hebei region peaked? A panel data analysis. *Ecological Indicators*. 119,  
697 106875.
- 698 Meng X, Zhongfeng Q, Shaohui Z., 2021. Carbon dioxide mitigation co-effect  
699 analysis of clean air policies: lessons and perspectives in China's  
700 Beijing-Tianjin-Hebei region. *Environmental Research Letters*. 16, 015006.
- 701 NBS 2016 China Statistical Yearbook 2016 National Bureau of Statistics.  
702 <http://www.stats.gov.cn/tjsj/ndsj/2016/indexch.htm>
- 703 NDRC. 2016. 13th Five-Year Plan of Renewable Energy Development. National  
704 Development and Reform Commission,  
705 <http://energy.people.com.cn/n1/2016/1219/c71661-28959415.html>
- 706 NEA. 2017a. 13th Five-Year Plan of Energy Development. National Energy  
707 Administration. [http://www.nea.gov.cn/2017-01/17/c\\_135989417.htm](http://www.nea.gov.cn/2017-01/17/c_135989417.htm)
- 708 NEA 2017b Clean Heating Plan for Winter in North China (2017-2021) National  
709 Energy Administration. [http://www.nea.gov.cn/2017-12/27/c\\_136854721.htm](http://www.nea.gov.cn/2017-12/27/c_136854721.htm)
- 710 Pope CA, 3rd, Burnett RT, Thun MJ, Calle EE, Krewski D, Ito K, et al., 2002. Lung  
711 cancer, cardiopulmonary mortality, and long-term exposure to fine particulate  
712 air pollution. *Jama*. 287, 1132-41.
- 713 PGC 2016 13th Five-Year Plan of Industrial Transformation and Upgrading. The  
714 People's Government of China.  
715 [http://www.gov.cn/xinwen/2016-11/07/content\\_5129638.htm](http://www.gov.cn/xinwen/2016-11/07/content_5129638.htm)
- 716 Pope CA, Burnett RT, Thurston GD, Thun MJ, Calle EE, Krewski D, et al., 2004.  
717 Cardiovascular Mortality and Long-Term Exposure to Particulate Air Pollution.  
718 *Circulation*. 109, 71-77.
- 719 Qi J, Zheng B, Li M, Yu F, Chen C, Liu F, et al., 2017. A high-resolution air pollutants  
720 emission inventory in 2013 for the Beijing-Tianjin-Hebei region, China.  
721 *Atmospheric Environment*. 170, 156-168.
- 722 Rong X, Wang H., 2016. Spatio-temporal Variations and Source Contributions of  
723 China's Premature Deaths Attributable to Ambient PM<sub>2.5</sub>. AGU Fall Meeting  
724 Abstracts.

- 725 Roth GA, Abate D, Abate KH, Abay SM, Abbafati C, Abbasi N, et al., 2018. Global,  
726 regional, and national age-sex-specific mortality for 282 causes of death in  
727 195 countries and territories, 1980–2017: a systematic analysis for the Global  
728 Burden of Disease Study 2017. *The Lancet*. 392, 1736-1788.
- 729 Su, C., H. Madani, and B. Palm 2018 Heating solutions for residential buildings in  
730 China: Current status and future outlook *Energy Conversion and Management*.  
731 177 493-510
- 732 TJES. 2015. Tianjin Environmental Statement 2015. Tianjin Municipal Environmental  
733 Protection Bureau.  
734 <http://sthj.tj.gov.cn/YWGZ7406/HJZL9827/HJZKGB866/TJSLNHJZKGB665>  
735 [3/202010/W020201021409644639275.pdf](http://sthj.tj.gov.cn/YWGZ7406/HJZL9827/HJZKGB866/TJSLNHJZKGB665)
- 736 Tian X, Dai H, Geng Y, Wilson J, Wu R, Xie Y, et al., 2018. Economic impacts from  
737 PM<sub>2.5</sub> pollution-related health effects in China's road transport sector: A  
738 provincial-level analysis. *Environment International*. 115, 220-229.
- 739 TSB 2016 Tianjin Statistic Yearbook Tianjin Statistical Bureau
- 740 Turner MC, Jerrett M, III CAP, Krewski D, Gapstur SM, Diver WR, et al., 2015.  
741 Long-Term Ozone Exposure and Mortality in a Large Prospective Study. *Am J*  
742 *Respir Crit Care Med*. 193, 1134-1142.
- 743 Wang Y, Xu M, Jin J, He S, Li M, Sun Y., 2014. Concentrations and relationships  
744 between classes of persistent halogenated organic compounds in pooled  
745 human serum samples and air from Laizhou Bay, China. *Science of The Total*  
746 *Environment*. 482-483, 276-282.
- 747 West JJ, Smith SJ, Silva RA, Naik V, Emmons LJE., 2012. Co-benefits of Global  
748 Greenhouse Gas Mitigation for Future Air Quality and Human Health via Two  
749 Mechanisms. *Epidemiology*. 23, 1.
- 750 West JJ, Smith SJ, Silva RA, Naik V, Zhang Y, Adelman Z, et al., 2013. Co-benefits of  
751 mitigating global greenhouse gas emissions for future air quality and human  
752 health. *Nature Climate Change*. 3, 885-889.
- 753 Wu R, Dai H, Geng Y, Xie Y, Masui T, Liu Z, et al., 2017. Economic Impacts from

- 754 PM<sub>2.5</sub> Pollution-Related Health Effects: A Case Study in Shanghai.  
755 Environmental Science & Technology. 51, 5035-5042.
- 756 Xie Y, Dai H, Dong H, Hanaoka T, Masui T., 2016a. Economic Impacts from PM<sub>2.5</sub>  
757 Pollution-Related Health Effects in China: A Provincial-Level Analysis.  
758 Environmental Science & Technology. 50, 4836-4843.
- 759 Xie Y, Dai H, Tatsuya H, Toshihiko M., 2016b. Health and economic impacts of PM<sub>2.5</sub>  
760 pollution in Beijing-Tianjin-Hebei area. China Population, Resources and  
761 Environment. 26, 19-27.
- 762 Xie Y, Dai H, Xu X, Fujimori S, Hasegawa T, Yi K, et al., 2018. Co-benefits of  
763 climate mitigation on air quality and human health in Asian countries.  
764 Environment International. 119, 309-318.
- 765 Xie Y, Dai H, Zhang Y, Wu Y, Hanaoka T, Masui T., 2019. Comparison of health and  
766 economic impacts of PM<sub>2.5</sub> and ozone pollution in China. Environment  
767 International. 130, 104881.
- 768 Xie, Y., Wu, Y., Xie, M., Li, B., Zhang, H., Ma, T., and Zhang, Y., 2020. Health and  
769 economic benefit of China's greenhouse gas mitigation by 2050.  
770 Environmental Research Letters. 15, 104042.
- 771 Xiong, W., Y. Wang, B. V. Mathiesen, H. Lund, and X. Zhang 2015 Heat roadmap  
772 China: New heat strategy to reduce energy consumption towards 2030 Energy.  
773 81 274-285
- 774 Xu M, Qin Z, Zhang S., 2020. Integrated assessment of cleaning air policy in China: a  
775 case study for Beijing-Tianjin-Hebei region. Journal of Cleaner Production.  
776 296, 126596.
- 777 Yang, H., Tao, W., Liu, Y., Qiu, M., Liu, J., Jiang, K., Yi, K., Xiao, Y., and Tao, S.,  
778 2019. The contribution of the Beijing, Tianjin and Hebei region's iron and steel  
779 industry to local air pollution in winter. Environmental Pollution. 245,  
780 1095-1106.
- 781 Zhang N-N, Ma F, Qin C-B, Li Y-F., 2018. Spatiotemporal trends in PM<sub>2.5</sub> levels from  
782 2013 to 2017 and regional demarcations for joint prevention and control of

- 783 atmospheric pollution in China. *Chemosphere*. 210, 1176-1184.
- 784 Zhang, S., E. Worrell, and W. Crijns-Graus 2015 Mapping and modeling multiple  
785 benefits of energy efficiency and emission mitigation in China's cement  
786 industry at the provincial level *Applied Energy*. 155 35-58
- 787 Zhang X, Jin Y, Dai H, Xie Y, Zhang S., 2019. Health and economic benefits of  
788 cleaner residential heating in the Beijing–Tianjin–Hebei region in China.  
789 *Energy Policy*. 127, 165-178.
- 790
- 791

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792 **Figures**

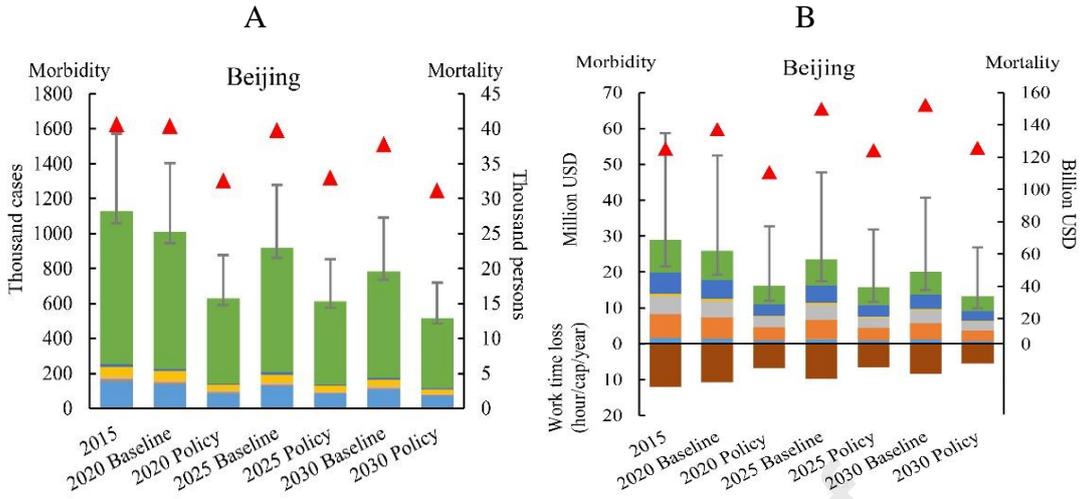
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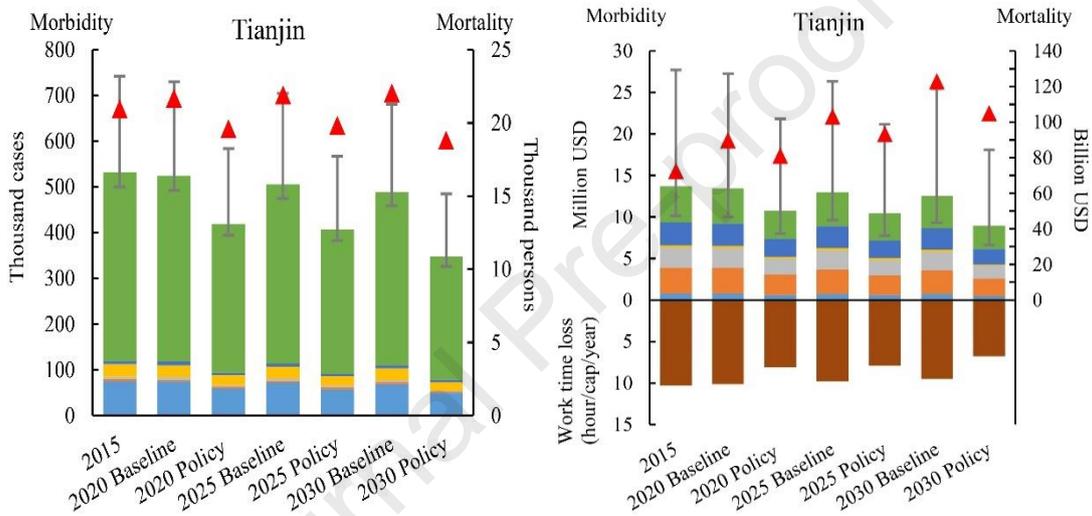
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Fig. 1. Research framework.

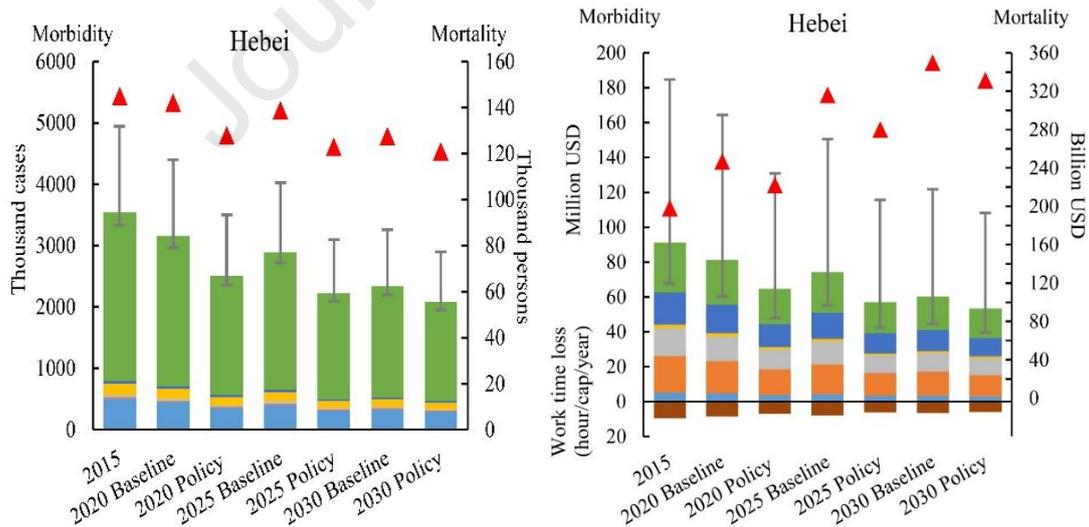
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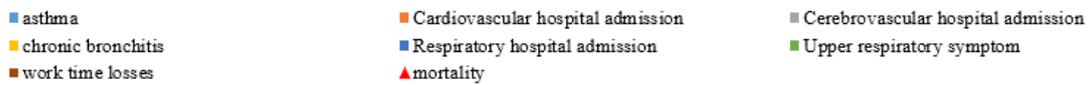
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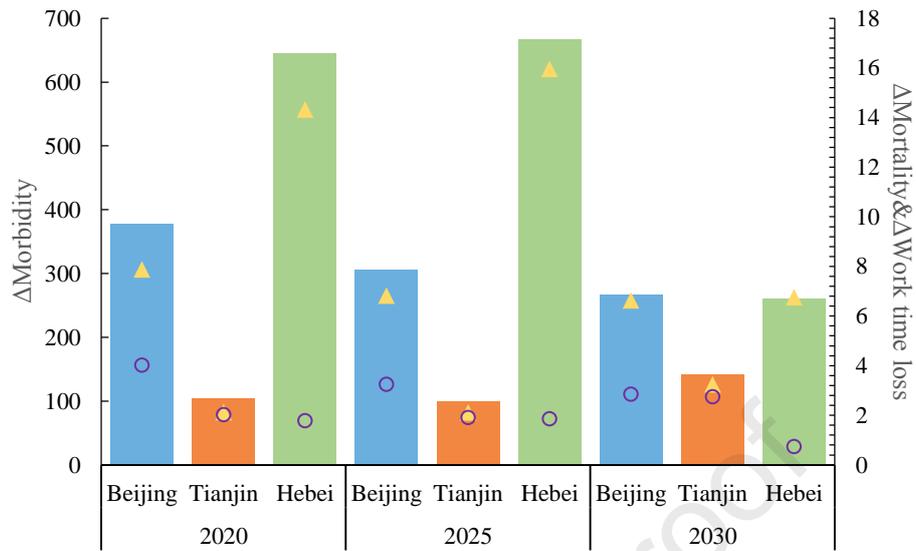
801 Fig. 2. Health effects of morbidities and mortalities (A) and expenditures, VSL and work time loss

802 (B) in the BTH region. (Bars refer to the left Y-axis and scatter triangles refer to the right Y-axis.

803 Lines in the middle of bars are error lines of morbidity and expenditures acquired from the 95%  
804 confidence interval of response functions.)  
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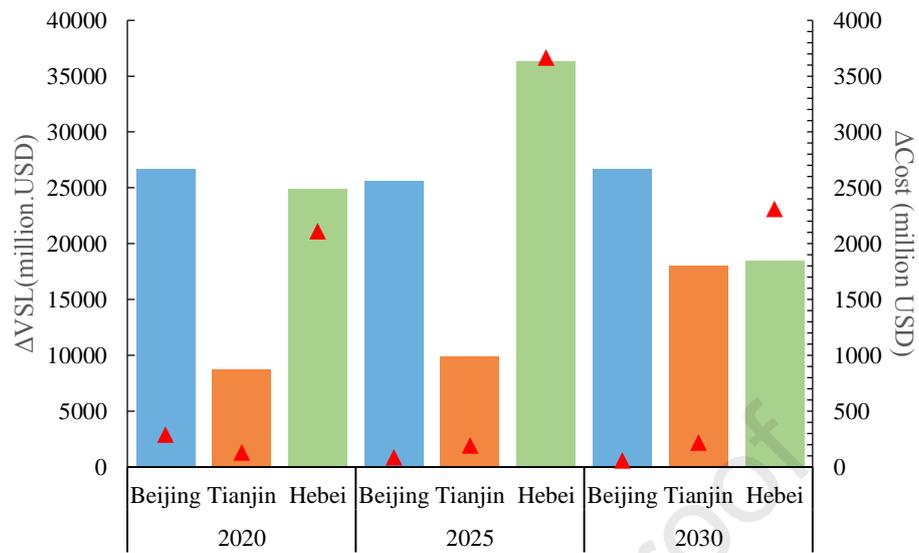
808 Fig. 3. Comparisons of absolute reductions (Thousand cases) in morbidity and  
 809 mortality and absolute reductions (hour/cap/year) in work time loss between the  
 810 baseline scenarios and the policy scenarios for 2020, 2025, and 2030, respectively.

811 (Bars refer to the left Y-axis while triangles and circles refer to the right Y-axis.)

812 (Detailed results are presented as supplementary tables in Table S4.)

813

814

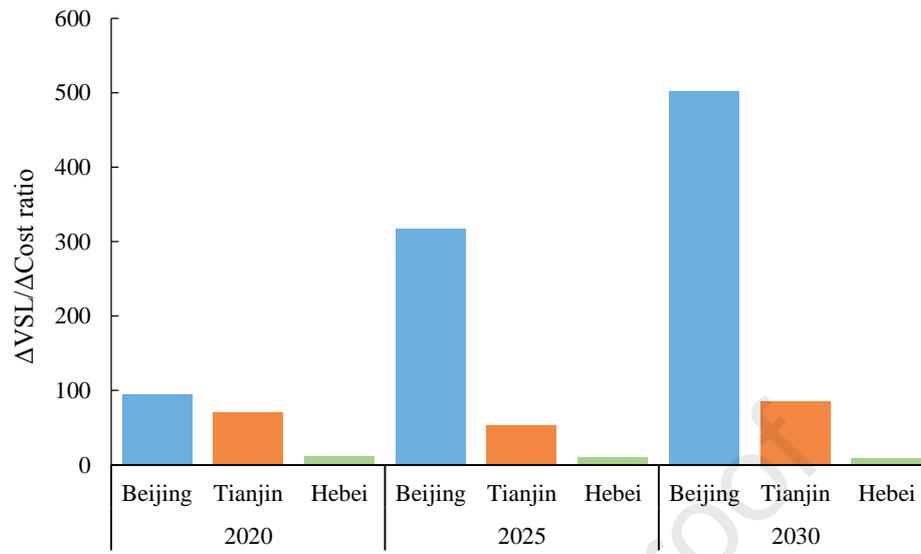


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816 Fig. 4. Co-benefit analysis of  $\Delta VSL$  and  $\Delta Cost$  in the BTH region. (Bars refer to the  
 817 left Y-axis while triangles refer to the right Y-axis.) (Detailed results are presented as  
 818 supplementary tables in Table S5.)

819

820



821

822 Fig. 5. The  $\Delta VSL/\Delta Cost$  ratio in the BTH region. (Detailed results are presented as

823 supplementary tables in Table S6.)

824

825

826 **Tables**

827

828 Table 1 PM<sub>2.5</sub> concentrations in different scenarios in the BTH region ( $\mu\text{g}/\text{m}^3$ ).

Regions	2015	2020		2025		2030	
		Baseline	Policy	Baseline	Policy	Baseline	Policy
Beijing	78.7	69.5	47.1	63.2	45.5	54.5	39.4
Tianjin	69.1	66.3	55.1	63.3	52.9	60.6	46.0
Hebei	65.2	58.3	48.4	54.0	43.8	45.4	41.5

829

830

831 Table 2 Emission control costs in different scenarios in the BTH region (million

832 USD/yr)

Regions	2015	2020		2025		2030	
		baseline	policy	baseline	policy	baseline	policy
Beijing	970.8	1506.2	1222.0	2075.1	1994.1	2411.3	2358.3
Tianjin	1476.3	2158.9	2034.9	2660.0	2473.1	2963.0	2751.9
Hebei	6951.3	11558.8	9454.1	13778.4	10119.2	13252.5	10947.0

833

- Implementing Blue Sky Protection Campaign has positive health and economic impacts.
- The PM<sub>2.5</sub> reductions are highly effective in reducing work time loss.
- A sum of 60 billion USD in value of statistical life will be avoided in BTH by 2020.
- The  $\Delta$  benefit/ $\Delta$  cost ratio in Beijing is the highest, followed by Tianjin and Hebei.

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**Declaration of interests**

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

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