

Contents lists available at ScienceDirect

# Global Environmental Change



journal homepage: www.elsevier.com/locate/gloenvcha

# Impact of the EU biodiversity strategy for 2030 on the EU wood-based bioeconomy

Fulvio Di Fulvio<sup>a,\*</sup>, Tord Snäll<sup>b</sup>, Pekka Lauri<sup>a</sup>, Nicklas Forsell<sup>a</sup>, Mikko Mönkkönen<sup>c,d</sup>, Daniel Burgas<sup>c,d</sup>, Clemens Blattert<sup>c,e</sup>, Kyle Eyvindson<sup>f,g</sup>, Astor Toraño Caicoya<sup>h</sup>, Marta Vergarechea<sup>i</sup>, Clara Antón-Fernández<sup>i</sup>, Julian Klein<sup>b</sup>, Rasmus Astrup<sup>i</sup>, Jani Lukkarinen<sup>j</sup>, Samuli Pitzén<sup>j</sup>, Eeva Primmer<sup>j</sup>

<sup>a</sup> Integrated Biosphere Futures (IBF) Research Group, Biodiversity and Natural Resources (BNR) Program, International Institute for Applied Systems Analysis (IIASA), Schlossplatz 1, 2361 Laxenburg, Austria

<sup>c</sup> Department of Biological and Environmental Sciences, University of Jyväskylä, P.O. Box 35, 40014 Jyväskylä, Finland

e Forest Resources and Management, Swiss Federal Institute for Forest, Snow and Landscape Research WSL, Zürcherstrasse 111, 8903 Birmensdorf, Switzerland

<sup>f</sup> Faculty of Environmental Sciences and Natural Resource Management, Norwegian University of Life Sciences, P.O. Box 5003, 1433 Ås, Norway

g Natural Resource Institute Finland (LUKE), Latokartanonkaari 9, FI-00790 Helsinki, Finland

h Chair of Growth and Yield Science, TUM School of Life Sciences, Hans-Carl-von-Carlowitz-Platz 2, 85354 Freising, Germany

<sup>1</sup> Department of Forest Management, Division of Forest and Forest Resources, National Forest Inventory, Norwegian Institute for Bioeconomy Research (NIBIO),

Høgskoleveien 8, 1433 Ås, Norway

<sup>j</sup> Finnish Environment Institute (Syke), Latokartanonkaari 11, FI-00790 Helsinki, Finland

#### ARTICLE INFO ABSTRACT Keywords: The EU Biodiversity Strategy (EUBDS) for 2030 aims to conserve and restore biodiversity by protecting large Bioeconomy areas throughout the European Union. A target of the EUBDS is to protect 30 % of the EU's land area by 2030, Wood harvest with 10 % being strictly protected (including all primary and old growth forests) and 20 % being managed 'closer Protection to nature'. Even though this will have a positive impact on biodiversity, it may negatively impact the EU's wood-Biodiversity based bioeconomy. In this study, we analyze how alternative interpretations and distributions of the EU's pro-Modelling tection targets may affect future woody biomass harvest levels, exports of wood commodities, and the spatial Forest sector distribution of managed areas under wood demands aligned with SSP2-RCP1.9. Using the model GLOBIOM-Leakage Forest, we simulate scenarios representing a variety of interpretations and geographic distributions of the EUBDS targets. The EUBDS targets would have a limited impact on EU harvest levels since the EU can still increase its wood harvest between 21 % and 24 % by 2100. With strict protection of 30 % of the area, the EU harvest level can still be increased by 10 %. Moreover, the most likely scenario (10 %/20 % protection within each MS) will result in increased net exports in the coming decades, but a slight decline after 2050. However, if

other biomes and mostly leaking into boreal regions.

# 1. Introduction

Forest-based bioeconomy aims to utilize forest resources to replace fossil-based raw materials and products (Wolfslehner et al., 2016). In the EU, forest resources and associated bioeconomy are controlled by several policy instruments, including those recently formulated under the European Green Deal (Aggestam and Giurca, 2021, Lier et al., 2021). As many of these instruments are still in the revision process, it would be difficult and premature to investigate the overall impact of the Green Deal on the EU forest-based bioeconomy. There are, however, some aspects of the Green Deal that are already well documented and could be investigated. One of them is the EU Biodiversity Strategy for 2030

protection is intended to also represent site productivity or to re-establish a green infrastructure, then EU net exports will also decline before 2050. With the decreased EU roundwood harvest, increased harvest will occur in

\* Corresponding author. *E-mail address:* difulvi@iiasa.ac.at (F.D. Fulvio).

https://doi.org/10.1016/j.gloenvcha.2025.102986

Received 30 January 2024; Received in revised form 18 November 2024; Accepted 28 February 2025 Available online 11 March 2025

<sup>&</sup>lt;sup>b</sup> SLU Swedish Species Information Centre, Swedish University of Agricultural Sciences, P.O. Box 7007, 75007 Uppsala, Sweden

<sup>&</sup>lt;sup>d</sup> School of Resource Wisdom, University of Jyväskylä, P.O. Box 35, 40014 Jyväskylä, Finland

<sup>0959-3780/© 2025</sup> The Authors. Published by Elsevier Ltd. This is an open access article under the CC BY license (http://creativecommons.org/licenses/by/4.0/).

(EUBDS), which addresses the current trends in biodiversity decline and sets a path to recovery (IPBES, 2019, Leclère et al., 2020). This plan includes a target to protect 30 % of EU land and sea areas by 2030 (EC, 2020a, b), of which one third (10 %) should be strictly protected and include all primary and old-growth forests. The remaining two thirds (20 %) should be under what is labelled as "closer-to-nature" management (Larsen et al., 2022). Currently, 23.3 % of EU forests are protected, including 4.7 % strictly protected and 18.6 % protected in other non-strict ways (WDPA, 2020).

To reach the EUBDS targets, many EU Member States (MS) need to adjust their land management and policies. However, the spatial distribution of protection within the EU has not been specified in the EUBDS. There are indeed ongoing scientific and political discussions concerning the spatial distribution of future protected areas (Hermoso et al., 2020) and the ways in which efforts to increase protection should be coordinated among the EU MS. In the current EU discussions, an equal allocation is considered at the level of countries or biogeographical regions (EC, 2022), but other distributions are also plausible since the EUBDS also includes objectives at the finer landscape scale. This is formulated as "green infrastructure", a "strategically planned network of natural and semi-natural areas with other environmental features, designed and managed to deliver a wide range of ecosystem services, while also enhancing biodiversity" (EC, 2013). Thus, MS are encouraged to conserve, manage and restore sufficient areas of functional habitat networks to support viable species populations and their migration. This should be achieved by protected areas, including the Natura 2000 network (EC, 2013, 2019a), sustainable forest management, as well as 'closer to nature' management. The latter is currently defined only at EU scale (EC, 2023a) and needs to be translated into MS guidelines and local management practices. Additionally, its implications in terms of providing wood, non-woody ecosystem services and biodiversity require thorough assessment.

Another key role of EU forests described in the EU Green Deal is the contribution to reach climate neutrality by 2050 (EC, 2019b, EC, 2021a), given their ability to absorb and store large amounts of carbon (EEA, 2020a). Combined, climate and conservation policies may create synergies and increase the overall environmental and socioeconomic sustainability. However, forests can be used for climate change mitigation in contrasting ways, namely to sequester carbon on the forest land (Griscom et al., 2017), store carbon in long-lived wood products (Mishra et al., 2022) and replace fossil fuel-based products and energy sources (Hurmekoski et al., 2023). The emphasis placed on the different climate mitigation functions vary across policy domains (Pitzén et al., 2023). Mitigation through woody biomass uses in the bioeconomy may conflict with environmental sustainability and biodiversity conservation (Blattert et al., 2023, Mazziotta et al., 2022), due to potential management intensification (Betts et al., 2021). Forest growth, harvest and management vary widely across the EU as well as the future CO2 sequestration trajectories of EU member states (Pilli et al., 2017, Pilli et al., 2022).

Achieving the EUBDS protection targets in the EU may have spillover effects in other parts of the world where wood is imported from (Rosa et al., 2023, Cerullo et al., 2023), with an increased risk of failing to achieve the targets of the Kunming-Montreal Global Biodiversity Framework (GBF, 2022). The reason is so-called leakage when decreased harvest in one region results in increased harvest elsewhere to meet the same wood demands. For example, Kallio et al. (2018) identified leakage when constraining EU wood harvest levels. However, they solely studied leakage given different country-level harvest, thus not varying neither management regimes nor areas protected within and among countries, both of which are key components of the EUBDS and further likely affecting leakage.

Protection according the EUBDS targets could have considerable impacts on the EU forest sector. Lotze-Campen et al. (2018) suggested a 13 % reduction in EU roundwood harvest potential when expanding the network of EU protected areas. However, they limited their analyses to year 2040 and simulated scenarios without any link to the EUBDS targets or the EU climate neutrality. Indeed, Schier at al. (2022) disclosed that EU roundwood harvest may be reduced by 9 %-48 % by 2030 and 11 %-58 % by 2050 when considering scenarios aligned with the EUBDS. However, they studied solely Germany and extrapolated the results to the rest of the EU without considering the very large spatial variation in allocation of protected areas and wood production.

Forest sector models provide an appropriate tool to investigate the impacts of polices on global resources use and bioeconomy while studying the complex dynamics of forest sector. Such models include aspects of forest management, forest industries, spatial explicit wood production costing as well as trade patterns, and enable the connection between wood supply and demands (Toppinen and Kuuluvainen, 2010, Latta et al., 2013, Riviere and Caurla, 2020). The forest sector Global Biosphere Management Model GLOBIOM (Havlík et al., 2011, Lauri et al., 2021) simulates management regimes and protection areas spatially explicitly at a 50 km<sup>2</sup> scale in the EU and at a coarser resolution globally. The model accounts for forest growths in each simulation unit and intensity of harvest according to management regimes applied within each unit. Hence, this model structure enables to simulate how changing distribution of management regimes would affect harvest, forest industry production and global trade in a consistently connected way. Further, GLOBIOM allows to capture potential wood harvest leakages in other global regions through wood trade.

The overall aim of this study is to investigate effects of alternative approaches to protect forests under policy targets aligned to the EUBDS on woody biomass harvest, considering also leakage outside the EU on different spatiotemporal scales. We contrast six scenarios of increased protection against a baseline scenario of unchanged protection. We quantify the leakage outside EU resulting from increased protection, including biomes that can be expected to compensate for the consequential decreased harvest within the EU. The spatially explicit scenarios differ in terms of how the protection is distributed among: EU member states, biogeographic regions, forest productivity across the green infrastructure, and combinations thereof. We further conduct a sensitivity analysis assuming a stricter interpretation of the conservation policy where all 30 % of the forest land is strictly protected. We use the state-of-art forest sector model GLOBIOM for simulating forest management and the resulting impacts on wood harvest and trade under each of the seven forest protection scenarios until year 2100.

# 2. Material and methods

# 2.1. Modelling approach

The scenario simulations were conducted using the Global Biosphere Management Model (GLOBIOM) (Havlík et al., 2011, 2014), a global spatially-explicit agricultural and forest sector model. The model is solved recursively for each 10-year time step by maximizing the economic welfare, defined as the sum of the producer and consumer economic monetary surplus (S.M4). Here we specifically used GLOBIOMforest, which includes forestry, forest industry and bioenergy modules, as described in Lauri et al. (2019, 2021, S.M1), but where the agricultural sector is simplified to include just one product, energy crops (S. M5). The wood supply in the EU region is modelled on a  $0.5^{\circ}$  grid (ca. 50 km<sup>2</sup>) while the demand and trade are modelled using 57 economic global regions (including 28 EU Member states as single regions, see Table S.M3.2), where wood supply in regions outside EU is modelled at the resolution of 2.0° (ca. 200 km<sup>2</sup>). Land-use and land-management change is modelled according to linear land-use and land-management change costs functions following the approach of Havlík et al. (2011).

Woody biomass supply is based on spatially explicit harvest potentials, harvest costs, transportation costs and forest-/management-regime specific land-use change costs. The harvest costs are based on G4M output and transportation costs on Di Fulvio et al. (2016). The initial forest conditions for starting our simulations (year 2020) and economic optimizations match FRA (2020) country level data on forest area, standing biomass stocks, the World Database of protect areas (WDPA, 2020) grid level data, Global Forest Management Map (Lesiv et al., 2022) grid level data. These constitute the basis for the forest types generation and assignment of forest management regimes (S.M1). Sustainable harvest potentials are based on MAI (Mean Annual Increment) for each grid cell, as simulated with the Global Forest Model G4M (Kindermann et al., 2006, 2008; Gusti and Kindermann, 2011). Moreover, for the historical periods 2000-2020, management areas are scaled to match with FAOSTAT country level harvest volumes (FAO, 2023). Using G4M data and global age-class database (Besnard et al., 2021), the model generates initial grid level biomass growth curves and structures, that are subsequently calibrated to the country level biomass stocks. Within each grid cell, the model can have different management regimes with different age-class dynamics, mortality and harvest intensity. After 2020, the age-class dynamics develops endogenously and the model is not allowed to harvest more than long term biomass growth, represented by the mean annual increment (MAI).

GLOBIOM-forest applies three management regimes (close to nature, multifunctional, high intensity) and three main forest types (primary, managed and secondary forests), each characterized by two species groups (conifer/broadleaf). Primary forests have not been used historically for wood production, managed forests are currently used for production while secondary forests are abandoned managed forests. The management regimes in GLOBIOM forest differ in the proportion of MAI that can be harvested. In high intensity management, the whole MAI can be harvested while in multifunctional and close to nature management, only a part of it can be harvested, up to a maximum threshold set for each management regime (Table 1). Consequently, harvest volume can be changed in each simulation grid cell by changing the managed forest area, the forest management regime or by changing the share of MAI harvested within each management regime.

#### Table 1

Management	regimes	calibration	and	parametrization	in	the	simulated	EU
protection sce	enarios ( <mark>S</mark>	.M1-2).						

Management regime	Area, year 2020 (Mha)	Area calibration method	Maximum allowed roundwood volume harvest (% MAI)	Maximum allowed logging residues harvest volume (% Potential)
Strict protection	7.8	Primary forests allocated according to Lesiv et al. (2022), National level area matching to FRA, 2020. Strict protection forest areas according to IUCN categories Ia, Ib, II, III (WDPA, 2020).	0	0
Closer to nature	30.5	Non-strictly protected forests (IUCN classes IV- V-VI, S.M2) within the Natura2000 network (WDPA, 2020)	50	0
Multifunctional	92.7	Residual after assigning all the other classes.	75	25
High intensity	32.6	Planted forests according to Lesiv et al. (2022)	100	50

In all scenarios simulated, the EU (EU27 and UK, hereafter referred as EU) forest area was fixed at 164 M ha, following FRA (2020) forest cover harmonization. For Norway, the forest area was fixed at 12.8 M ha following FRA (2020). The total area available for management was kept constant over time.

After the initialization of starting year 2000, the allocation of area shares to the four management classes in each grid cell 2000–2100 was controlled endogenously by the model and based on the economic optimization (i.e. maximization of economic surplus), aiming to fulfil demands for wood at national level under each scenario.

GLOBIOM-forest includes 26 wood-based products, five harvested products and deadwood. These includes as an example four paper and paperboard grades, four pulp grades, three mechanical forest industry products, four forest industry by-products and two recycled products (S. M1). Bioenergy includes traditional fuelwood, woodchips and wood pellets. Forest industry and wood pellets production capacities are based on FAOSTAT production data for 2000–2020 (FAO, 2023). Final product demands are based on constant elasticity demand functions, which are parametrized by reference volumes, reference prices and elasticity coefficients. Exceptions are modern bioenergy demand, which is based on the socio-economic and climate pathways (SSP-RCP), sourced from the MESSAGE energy model (IIASA, 2020), and the traditional bioenergy demand, which is assumed to stay constant over time. After 2020, the demand volumes change over time based on gross domestic product (GDP) development and human population growth following the SSP-RCP scenario data (Riahi et al., 2017, IIASA, 2020). Trade is modelled by using bilateral trade flows. Bilateral trade calibration volumes are based on the BACI trade database for 2000-2020 (Gaulier and Zignago, 2010). After 2020, trade volumes evolve following trade dynamics, which depend on constant elasticity trade-cost functions parametrized by historical trade volumes and transport costs.

# 2.2. Protection scenarios

We simulated six scenarios on how to reach the EUBDS protection targets for 2030, contrasting them against a scenario of no change in protection, termed Current protection (Table 2). All scenarios assumed national timber demands for mitigating climate change aligned with the EU Green Deal policy package and the aim that EU should become climate neutral by 2050 (IIASA, 2020). In GLOBIOM, this is implemented as a bioenergy demand from the MESSAGE model (Riahi et al., 2017) following a 1.5 degrees warming scenario (RCP1.9) and a socio-economic development following a "middle of the road" (SSP2) until the year 2100. This leads to an increased EU woody biomass demand of 31 % during 2020–2100 (IIASA, 2020), see also *Sensitivity analyses*.

The model was first run for a calibration period of 20 years (2000–2020), where it was forced to reproduce the harvest volumes for this period according to FAOSTAT (FAO, 2023). Next, the model was run recursively until year 2100 in 10 years-time steps. The protection scenarios were implemented after year 2020, i.e. in year 2030.

# 2.3. Sensitivity analyses

We investigated effects of a stricter interpretation of the conservation policy objective compared to the main scenarios (Table 2) by applying '30 % strict protection', i.e. no harvest, on the total 30 % of forests under future protection. We also investigated effects of assuming lower increase in future EU wood demands (8 % increase during 2020–2100 instead of 31 %). This could reflect a broad combination of socioeconomic growth (SSPs) and climate development (RCPs), reducing the expected increasing rate of wood demand in the future, similarly to assuming 'RCP6.5' instead of RCP1.9 (Lauri et al. 2019). The sensitivity analyses were applied to three main scenarios: EU, Country, Green infrastructure.

#### Table 2

Assumptions o	f the	e six	protection	scenario	s simu	lated

Scenario name	Description
Current protection EU	Maintains current EU protection areas. An economy-driven scenario. Protection is distributed to land where it has the least impact on economy in terms of wood production and harvested wood products within the EU. Areas that are currently least profitable to harvest are progressively moved to strict protection and 'closer-to-nature' management until the 10 %/20 % target is reached. Forest areas mapped in each grid cell as potential primary forests (7.5 Mha in EU) in Sabatini et al. (2020) were excluded from future management.
Biogeographical region (BioGeo)	As EU, but objectives are achieved for each EU biogeographical region. This scenario aims to reach representativeness of the EU biogeographical regions in the allocation of protection.
Country	As EU, but objectives are achieved for each EU country. This scenario aims for equal distribution of protection areas in each EU country, and it is the intended implementation according to the EU Guidelines (EC 2022).
Country BioGeo	As EU, but objectives are achieved for both each EU country and EU biogeographical region. This scenario aims to reach representativeness of the EU biogeographical regions in the country-level allocation of protection.
Country Productivity	As Country BioGeo, but allocation of protection within countries so that protected and unprotected forests have the same country-level mean productivity (including both new and areas protected already in 2020). This scenario aims to distribute protection to not only the least profitable land, but considers also representativeness concerning productivity (affecting biodiversity) and EU Habitat types.
Green Infrastructure	As EU, but allocation of the 10 $\%/20$ % protection target within each grid cell (0.5°). This scenario aims to distribute protection to rebuild the EU green infrastructure and facilitate dispersal and migration of species.

#### 2.4. Leakage effects of biodiversity protection scenarios

Given the scenarios for protection within the EU, we finally also investigated how changes in future harvest levels within the EU may affect wood harvest in other biomes. Leakage means that a decrease of production or biophysical activity in one region is offset by an increase of production or biophysical activity in other regions (S.M3.3). We quantified leakage by absolute harvest volumes (increased harvest in other biomes) or by rates (overall change in other regions relative to the EU). Specifically, we investigated leakage of roundwood harvest and production of semi-finished wood products.

#### 3. Results

# 3.1. Protected forest area in Europe

The EU scenario, collectively reaching the policy objectives at EU level, would mean doubling the forest area under strict protection and increase the area under a 'closer-to-nature' management by 1 % compared to Current protection (Table 3). All other scenarios would mean protecting larger proportions of the area than 10 % and 20 % across all the EU (Table 3). The Green Infrastructure scenario means protecting the largest area, with an 8 % increase of the forest area under strict protection, compared to the EU scenario. This is explained by the current protection area of some simulation units already exceeding the unitary protection targets, while at the same time increasing the protection area in other units, summing up to protecting 13 %/25 % at EU scale (Table 3).

Under the Current protection scenario, there will be a 22 % expansion of high intensity forest management, increasing by 22 Mha between 2020 and 2100, with a parallel reduction of multifunctional management (Fig. 1). This intensification trend in unprotected forest areas is driven by the increasing wood demands under the climate neutrality assumed for all simulated scenarios. The six scenarios with expansion of protected areas reduce the areas available for other types of management (multifunctional and high intensity) (S.R2.1). Relative to the Current protection scenario, the increase in protected area of the Green Infrastructure scenario (Fig. 1), firstly reduces the areas under both multifunctional and high intensity management. Later, the development is similar to Current protection with an increase of high intensity management at the cost of decreasing multifunctional management. The other scenarios will fall between these two extremes.

# 3.2. Distribution of forest protected area

The EU scenario, collectively reaching the policy objectives means increasing strict protection in countries where wood production gives a relatively low profit (Fig. 2). Bulgaria, Ireland, Italy, Netherlands, and Romania would protect more than 15 % of their forest area, while countries like Finland and Sweden instead would contribute minimally (Fig. 2). This contrasts the Country scenario, where strictly protected areas are evenly distributed among all the EU countries. The same holds true under the Green Infrastructure scenario, although leading to a higher overall protection (Fig. 2, Table 3. S.R2). In some countries (e.g. Portugal, Spain), the Country scenario means decreased 'closer to nature' management compared to Current protection. This is due to areas changing into becoming strictly protected areas are preferably distributed to areas that are currently under close to nature management.

# 3.3. Harvested volume and wood prices

The EU biomass harvest volume (sum of roundwood and logging residues) projected under the Current protection scenario (and aiming for climate neutrality) is expected to increase by 31 % during 2020–2100, and by 10 % already by 2030 (Fig. 3). Compared to Current protection, the EU scenario will reduce harvest by 1 % in 2030 and 3 % by 2100. A harvest reduction of 5–7 % by 2100 is projected if additional protection is distributed equally among biogeographical regions or countries. However, if the additional protection is distributed equally

#### Table 3

Strictly protected forest and 'closer to nature' management given EU protection scenarios.

Scenario	Strictly protected	Closer to nature management (Mha)	Strictly protected (% of total forest area)	Closer to nature management (% of total forest area)	
	(Mha)		<b>(</b> • • • • • • • • • • • • • • • • • • •		
Current protection	7.8	30.5	5 %	19 %	
EU	16.2	32.5	10 %	20 %	
BioGeo	16.8	36.1	10 %	22 %	
Country	16.6	37.1	10 %	23 %	
Country BioGeo	17.5	37.2	11 %	23 %	
Country Productivity	16.6	38.5	10 %	23 %	
Green Infrastructure	20.7	40.6	13 %	25 %	

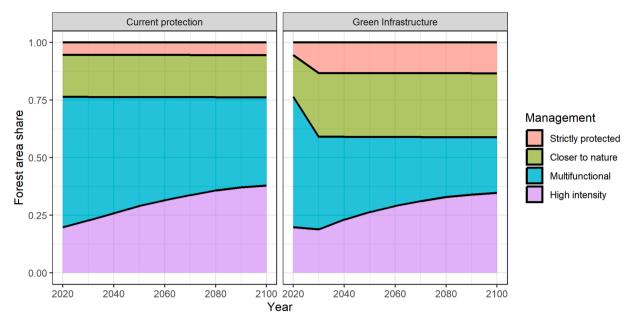


Fig. 1. Development of the proportion of EU forest area (of total forest area) managed by different regimes from year 2020 to 2100 assuming the Current protection and Green Infrastructure scenarios. (For interpretation of the references to colour in this figure legend, the reader is referred to the web version of this article.)

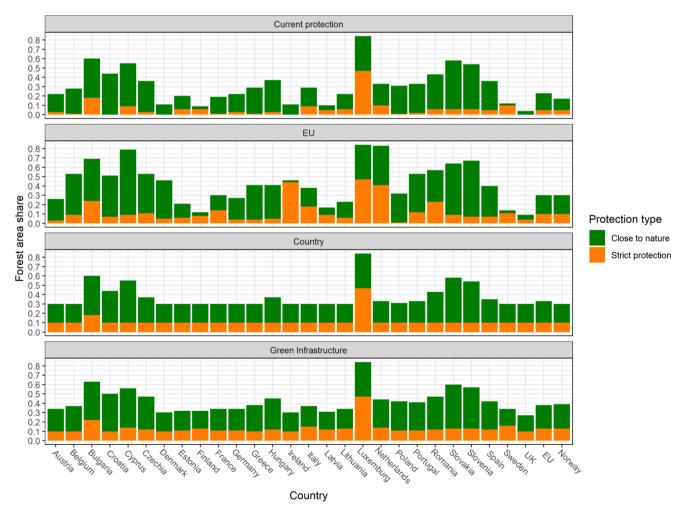


Fig. 2. Share of forest area protected in different EU countries (incl. UK and Norway) from year 2030 onwards assuming different scenarios of strict protection and close to nature management.

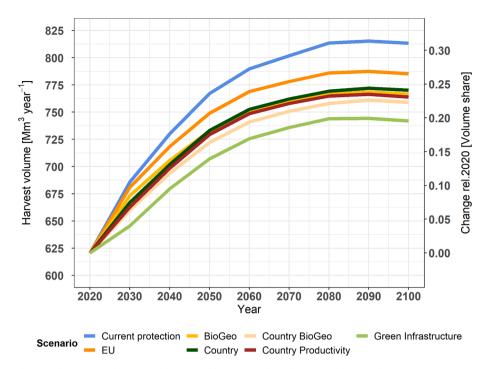


Fig. 3. Development of yearly forest biomass (including roundwood and logging residues) harvest volumes in the EU under SSP2-RCP1.9 and the protection scenarios simulated in absolute volumes (left) and as relative changes compared to the year 2020 (right).

among the  $0.5^{\circ}$  grid cells, Green Infrastructure, the reduction will be higher (6 % by 2030 and 9 % by 2100), simply because a larger area is protected (Table 3, Fig. 2). Nevertheless, all scenarios that increase protection will allow to increase EU harvest by at least 21 % (increasing from Current protection to Green Infrastructure; Fig. 3).

The impact of protection scenarios on wood assortment prices for the EU follows the same order as the one observed for harvest volumes (S. R5). Under the EU scenario, we observe increases in prices of 2-4 % until 2030 compared to the Current protection scenario. Under the Country scenario the increases are 6–15 % and under the Green Infrastructure prices increase between 10–30 %. Conifer wood assortments tend to be more impacted than broadleaves and the price increase is transient and declines over time.

# 3.4. Distribution of harvest volume among countries

Independently of protection scenario, individual countries will generally increase their future harvest by 2100 (4a-b), albeit less than with Current protection (Fig. 4c-d). The changes over time in harvest volumes will range between -28 % (Czech Republic) and +125 % (Luxembourg).

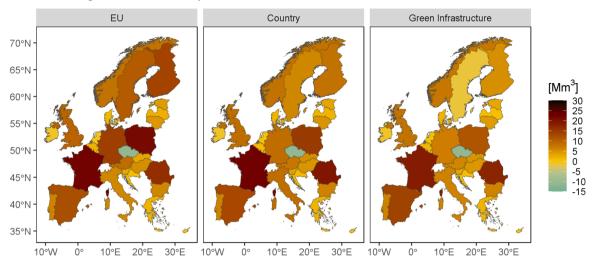
Five out of the six countries with the largest harvest volumes by 2100 (Germany, Finland, France, Poland and Romania) (Fig. 4, S.R3) will maintain or increase harvest levels in the long term with all protection scenarios, but in short term (2030), Germany and Sweden could experience a 6-12 % harvest reduction compared to the current levels under the Green Infrastructure scenario. The proportional impacts of the protection scenarios on harvests are generally higher on single countries than on the whole EU (cf. Fig. 3 and Fig. 4). For example, the EU level harvest reductions would reach 9 % under the Green Infrastructure scenario, while for individual countries this would range between 0-17 %. However, this conclusion concerns comparison with the Current protection scenario and year 2100 with assuming strong harvest increase overall (Fig. 3). If we would compare year 2100 to current harvest levels (2020) under the Green infrastructure scenario, we find harvest increase by 20 % at EU level, but between -28 % and +105 % at national level.

Compared to Current protection, the EU scenario means larger reductions (more than 5 Mm<sup>3</sup> in 2100) for only one country (France; 4c-d), while with the Green infrastructure scenario particularly the Northern countries and countries with large wood harvest volumes will be affected (4c-d). However, harvests in the Northern countries will still be similar to current level (4b). In the Country scenario, the country-level relative impact is lower, as more countries are involved in reaching the protection target. In this scenario, impacts over 5 Mm<sup>3</sup> are projected for Germany, Poland, Finland and Sweden. Whereas, with the Green Infrastructure scenario, impacts over 5 Mm<sup>3</sup> are projected for Sweden, Poland, Finland, Germany, France and Norway.

# 3.5. Net export and wood industry production

In year 2020, the EU was a net importer of roundwood biomass (-18)Mm<sup>3</sup> in 2020), and will remain being importer independently of scenario (Fig. 5a). However, because of increasing harvests under all scenarios (Fig. 3), we project reduced roundwood net import rate towards 2100. Regarding semi-finished wood products, the EU was instead a net exporter (57 Mm<sup>3</sup> in 2020) (Fig. 5b). When summing the historical roundwood and semifinished products net exports (Fig. 5c), the EU was instead a net exporter of 39 Mm<sup>3</sup> in 2020. The trade advantage in semifinished products is expected to continue to grow until 2040 and thereafter decline, due to increased competitiveness of other regions (South America, Asia). The impact of increased protection on the net export closely reflects the impacts on the EU biomass harvest volumes (Figs. 3-5). Net export of semi-finished products will decline under current levels in most scenarios after year 2040, whereas the EU scenario would allow to maintain current export levels. By 2100, the net export of semi-finished products would decline with 4 Mm<sup>3</sup> under the Country scenario and with 21 Mm<sup>3</sup> under the Green Infrastructure scenario, compared with the current level (Fig. 5b). Nevertheless, under all scenarios EU would remain a net exporter of semi-finished wood products.

Among the major wood production countries (Germany, Finland, France, Sweden), all protection scenarios will reduce the net export, given their expected future decreased economic competitiveness within the EU (S.R3). In parallel, other major wood-producing countries are



A. Change in harvest compared to 2020



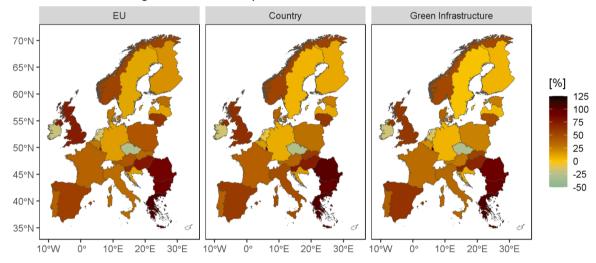


Fig. 4. Change in harvest volumes per country in 2100 for three protection scenarios when comparing to year 2020 in absolute (A) and relative values (B), and when comparing in year 2100 to the Current protection scenario in absolute (C) and relative values (D) under SSP2-RCP1.9.

expected to increase their competitiveness, as a result of their expected economic growth (Poland, Romania).

Net export of semifinished products is closely related to the industrial production levels. The changes in production of the main semifinished products categories (sawnwood and woodpulp) reflect the order of scenario impacts and magnitude for harvest volumes in EU (Fig. S.R3.7). However, among the main producing countries, Sweden is unable to maintain current sawnwood and wood pulp production under the Country and the Green Infrastructure scenarios. Similarly, Germany experiences a reduction of sawnwood production under these scenarios, while the other major wood producers are still able to maintain or increase their levels (Fig. S.R3.8-Fig S.R3.10). French woodpulp production is more strongly affected under the EU scenario compared to the Country scenario (Fig S.R3.10), and oppositely to the other major producers, the Country and Green Infrastructure scenarios are more impactful than the EU scenario.

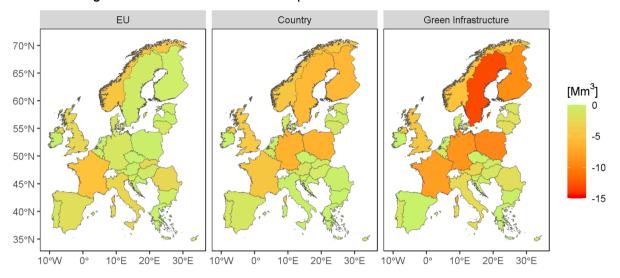
The EU was historically a net importer of wood pellets  $(-30 \text{ Mm}^3 \text{ in } 2020)$ . The net export of wood pellets is projected to become more negative over time under all scenarios (Fig. 5d), because of the expected bioenergy demand increase under RCP1.9 which cannot be satisfied by

increased harvest from EU forests. Hence, there is a larger increase of woodpellets import from other global regions. Compared to the Current protection scenario, the protection scenarios will lead to a 3-6~% decrease in wood pellets net export towards year 2100 (i.e. import increase by  $1-2~{\rm Mm}^3$ ) (Fig. 5d).

# 3.6. Sensitivity analyses

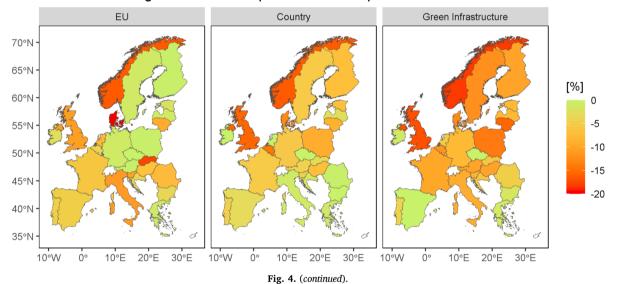
Assuming strict protection of 30 %, instead of 10 % of the EU forest area, the biomass harvest would decline by 10–15 % compared to the Current protection scenario in year 2100 (Fig. 6a). However, also with this high protection level, the harvest would increase compared to today. Nevertheless, the harvest reduction given by 30 % strict protection combined with the Green Infrastructure scenario is large enough for EU to become a net importer after 2040 (Fig. 6b). With all other scenarios simulated, the EU would remain a net exporter albeit at a lower level than with Current protection.

If instead assuming lower future demand for wood, i.e. following climate change mitigation leading to the RCP6.5, EU harvests may still somewhat increase in the coming decades under all scenarios. Moreover,



# C. Change in harvest relative to Current protection

D. Relative change in harvest compared to Current protection



harvest could possibly decrease (2–6 %) with protection scenarios compared to the RCP6.5 Current protection scenario (Fig. 6c). Netexport would follow a similar development as for the main scenarios until 2060, and thereafter continue to decrease sharply, though remain positive until 2100 (Fig. 6d).

# 3.7. Leakage outside EU and in other sectors

Up to 79 % of the decreased roundwood harvest within the EU resulting from the protection scenarios would result in increased harvest in the rest of the world (Table 4, Fig. 7). This relative change is hereafter referred as EU leakage rate. Specifically, the average leakage rate in EU roundwood harvest ranged between 22–79 % (Table 4, Fig. 7). However, this is compared to the Current protection scenario, which assumes a large harvest increase in the future (Fig. 6). The leakage is immediate, but also transient resulting from new equilibria reached after the middle of the century.

During the coming 40–50 years, the leakage only affects harvest in the boreal parts of the world, with negligible impact on the tropics or the temperate region (Fig. 7). After 2060, the contribution of the tropical region starts to increase, going from 10 % to 20 %, in parallel to a decreasing contribution from the boreal region toward the end of the

century (Fig. 7, S.R4).

The relatively lower leakages after the middle of the century are caused by a contraction of the global roundwood harvest rate (Fig. 8A). This is driven by the development in bioenergy feedstock in the rest of the world, where a substitution of roundwood by wood pellets from energy crops is projected to take place (Fig. 8B). However, the relative effect of the EUBDS protection on these future leakages is rather low, and the trade dynamics are mainly driven by historical trade developments and GDP growth.

Hence, we note two complementary effects, on one hand a global contraction of roundwood harvest takes places reaching 50  $Mm^3$  in the most ambitious protection scenario (30 % Green Infrastructure). On the other hand, harvest from energy crops increases and reaches a maximum of 45  $Mm^3$  under the same scenario (Fig. 8). Another leakage effect is endogenously considered in the model, with an intensification of forest management in unprotected forest areas within the EU.

# 4. Discussion

We present the impacts of alternative spatial allocations of protected areas according to the EU Biodiversity Strategy (EUBDS) on the EU forest sector, considering specifically wood harvest and wood-based

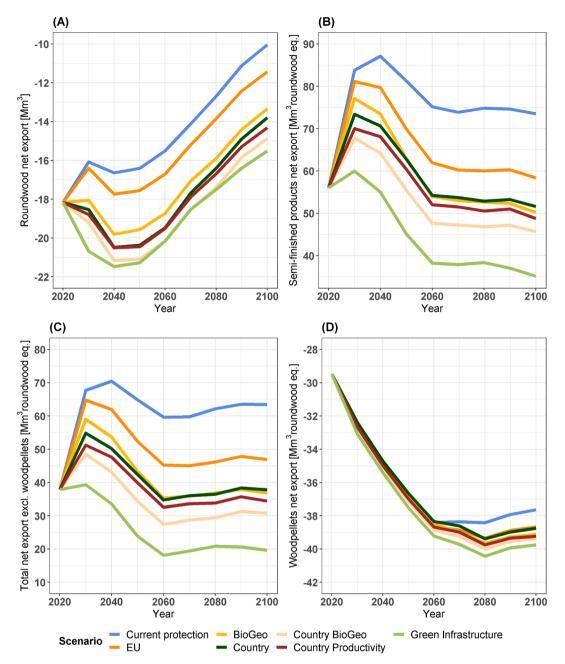
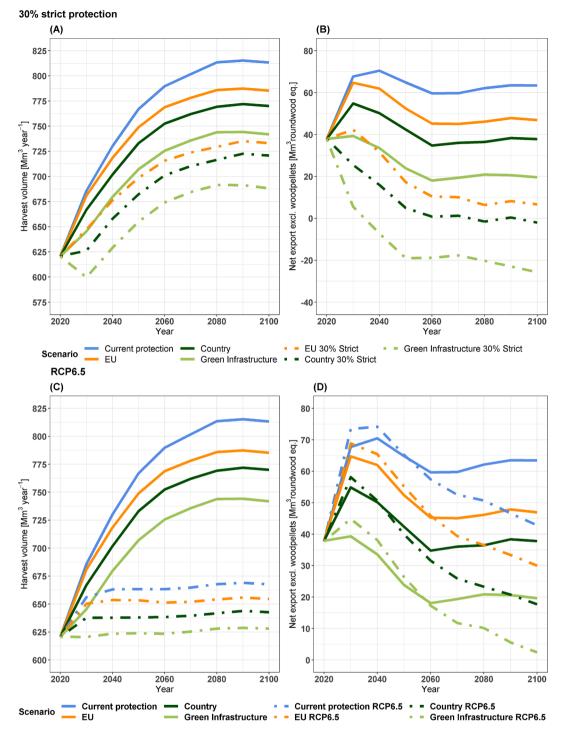


Fig. 5. Development of EU net export of different wood assortments under SSP2-RCP1.9 and different protection scenarios, specified by roundwood(e.g. sawlogs, pulpwood) (A), semi-finished products (eg. sawnwood, woodpulp) (B), total net export (sum of roundwood (A) and semi-finished products (B) (excluding wood-pellets), C) and woodpellets net export (D).

industry developments. We provide a seamless understanding of implications of different interpretations of targets, including spatial tradeoffs, EU external impacts and a global perspective. We show that the EUBDS will have a larger risk of impact on the forest industries than on harvest levels, and further that these risks concern the last half of the century. Implementation of the EUBDS in fact still allows for an increase of the EU forest harvest level by 21 % to 31 %, depending on the spatial distribution of the EUBDS targets. We expected the EUBDS protection to have larger impact on the EU forest industries net export of roundwood and semifinished products. However, the most likely scenario (10 %/20 % protection within each MS) means increased net exports than current ones in the coming decades, but slightly lower after 2050. With protection also aiming to be representative of site productivities or rebuilding the green infrastructures, net exports decreased also during the coming decades. The study also allowed to show that the leakage resulting from decreased EU roundwood harvest (ca. 80 % of the decrease) will correspond to higher harvest in boreal regions outside the EU, but a minor increase in tropical regions.

In terms of total EU woody biomass harvest levels, we show that an implementation of the EUBDS would have limited impact. At the EU level, we project that harvesting can increase by 24 % between 2020 and 2100 with implementation at the EU level, and by 21 % with the green infrastructure implementation. A very strict interpretation of 'closer-to-nature' management, i.e. strict protection and hence excluding 30 % of the land from harvesting, would restrict harvest more, but still allow for higher future harvests than today (10 % increase). Nevertheless, our scenarios did not account for potential impacts of increased harvests on non-timber ecosystem services and biodiversity outside protected areas. Increasing harvest to strive for climate mitigation targets might cause adverse effects on forest multifunctionality (Blattert et al., 2023), which



**Fig. 6.** Harvest (A) and net export (B) projections assuming considerably increased protection, specifically 30% strict protection (dashed line) under SSP2-RCP1.9. C and D show the corresponding trajectories under a considerably lower demand for wood following the SSP2-RCP6.5 scenario (dashed line). For comparison, the trajectories of the four main scenarios from Figs. 3 and 5 are also shown (solid lines).

is a central objective of the EU Forest Strategy (EC, 2021b). Future studies should take this multifunctional view into account while exploring the interplay of protection and mitigation policies.

Our findings of 2 % to 6 % harvest reductions by 2050 contrast the recent study of Schier et al. (2022) who claimed an overall harvest reduction of 11-58 % by 2050 for the EU27 compared to a baseline development that aligns to our RCP6.5. The main reason for the difference is that they applied a stricter definition of "old-growth" forest adopting variable age thresholds (i.e. more than 120 years for Norway spruce and 160 years for oaks) for identifying old growth forests and

further excluding them from wood harvesting in the future. Such an agebased identification of "old growth" forests implies that any forests that over time reaches the threshold automatically becomes protected. We believe that this overestimates impact of the EUBDS, as forest age per se cannot be considered sufficient information for identifying "old-growth" forests according to EC (2023b), recommending that a series of complementary structural indicators would be needed.

Our study is the first one to provide long-term developments beyond 2050 and also effects of alternative distributions of increased protection in accordance with the EUBDS. Concerning the near-coming decades,

# Table 4

Change in roundwood harvest (Mm<sup>3</sup> over bark) in EU and Rest of the World compared to the Current protection scenario and associated harvest leakage rate\* (%) for six scenarios (see also Fig. 7 for scenario EU, Country, Green Infrastructure).

		Year							
	Scenario	2030	2040	2050	2060	2070	2080	2090	2100
EU	EU	-4.3	-11.2	-16.8	-18.8	-19.6	-21.8	-21.7	-21.2
Harvest decrease	Country	-17.1	-24.9	-29.0	-30.0	-30.2	-32.5	-31.7	-31.0
$(Mm^3 year^{-1})$	Green Infrastructure	-35.6	-43.3	-49.5	-49.2	-48.0	-49.4	-50.3	-50.1
	EU 30 % Strict	-36.8	-52.5	-65.6	-68.9	-70.3	-74.2	-70.3	-70.2
	Country 30 % Strict	-54.8	-65.7	-75.8	-76.2	-76.2	-79.6	-76.0	-76.8
	Green infrastructure 30 %	-77.4	-90.4	-97.9	-96.0	-95.0	-97.5	-99.3	-100.1
Rest of World	EU	2.8	6.5	11.6	12.0	4.6	6.0	4.9	6.4
Harvest increase	Country	12.5	18.2	20.9	20.1	11.4	13.8	11.7	13.8
(Mm <sup>3</sup> year <sup>-1</sup> )	Green Infrastructure	26.6	33.5	38.4	29.8	20.2	24.9	25.0	29.2
	EU 30 % Strict	24.2	36.3	43.6	32.5	22.5	29.0	30.8	35.3
	Country 30 % Strict	38.8	50.2	55.0	39.0	31.9	39.3	43.2	48.0
	Green infrastructure 30 %	56.9	71.2	76.9	52.4	44.3	53.1	57.5	64.7
Leakage rate (%)*	EU	66	58	69	64	24	28	22	30
	Country	73	73	72	67	38	42	37	44
	Green Infrastructure	75	77	78	60	42	50	50	58
	EU 30 % Strict	66	69	66	47	32	39	44	50
	Country 30 % Strict	71	77	73	51	42	49	57	62
	Green infrastructure 30 %	73	79	79	55	47	55	58	65

\*Leakage rate (%) = -100 x (Rest of the Word harvest increase/EU harvest decrease).

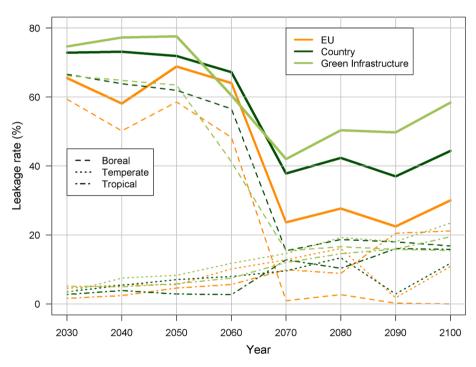


Fig. 7. Roundwood leakage rates from the EU to three forest biomes outside the EU given three protection scenarios compared to the Current protection scenario (Table 4).

our results are generally in line with Lotze-Campen et al. (2018) that estimated a reduction in stemwood harvest of 45–73 Mm<sup>3</sup> by 2040 when expanding protection areas (compared to their baseline), which aligns to the 43 Mm<sup>3</sup> reduction in stemwood harvest estimated with the Green Infrastructure scenario in the present study. The main reason for the higher reductions in Lotze-Campen et al. (2018) is that they assumed a higher 26 % increase in strictly protected area by 2040 instead of our 13 % in the Green Infrastructure scenario.

We find that applying any scenario to protect forests according to the EUBDS will dampen the future increase in EU harvest level and will be compensated by increasing roundwood harvests in other regions. For every cubic meter reduction, the rest of the world will increase its harvest by up to 0.79 cubic meters. The rest of EU harvest reduction will translate into lower global roundwood harvest and consumption, which

could potentially be replaced by the consumption of other commodities (including fossil based) not included in our modelling. However, the global roundwood harvest reduction (max. 50  $\text{Mm}^3 \text{ year}^{-1}$ ) is relatively insignificant compared to the current global roundwood harvest (3.9 Billion m<sup>3</sup> in 2020). This level of leakage is in agreement with estimated the roundwood harvest leakage by Kallio et al. (2018) for the year 2030. However, we further show that the leakage will decrease over time. This decreasing leakage trend occurs due to the structure of GLOBIOM-forest that accounts for the interaction of roundwood from natural forests with other sources of woody material from energy crops within and outside the EU. These additional sources of wood will compensate for the reduced roundwood harvest in the EU and thus decrease the leakage outside the EU in the long run. The boreal region absorbs most of the leakage, as also concluded by Kallio et al. (2018) and Schier et al.

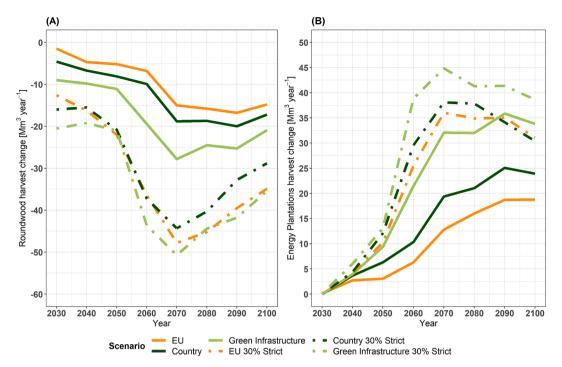


Fig. 8. Development of roundwood harvest (A) and energy crops harvest (B) compared to the Current protection scenario under SSP2-RCP1.9.

(2022). However, although the impact on the tropical region is fairly low in terms of wood volumes, the tropical region has much higher endemic biodiversity that might be under risk (Hill et al., 2019, Jung et al., 2021). Rosa et al. (2023) have accounted for global biodiversity impacts of combined increasing wood harvests and expansion of protection/'closer-to-nature' management of the EU forest. They conclude that increasing strict protection to more than 25 % of the EU currently managed forestland by 2100 increased the global extinction risk compared to continuing current forest management. However, a major limitation of their work is that their biodiversity measure concerned plants, mammals and birds by applying the countryside species-area relationship. A focus on just these species groups will underestimate the biodiversity benefits of protection within the EU, as these are not the main groups negatively affected by forestry in the EU. These are instead species associated with deadwood, old forests (rich in deciduous trees) (EEA, 2020b) and concern species-rich groups of invertebrates, fungi, cryptogams etc. Thus, further work is needed to understand leakage effects of adopting protection according to the EUBDS.

Our study also presents solutions for how decreased harvest due to increased protection can be mitigated for within the EU, by intensified management in non-protected forest. An intensification of harvest in forest area managed for production is identified as a solution to mitigate for the implementation of the EUBDS in Finland (Räty et al., 2023). Intensified harvesting on production land also comes out as a solution when optimizing management to reach also biodiversity targets under increasing wood demands (Blattert et al., 2022, Eggers et al., 2022), and constitutes a strategy under climate smart forestry for facing future European wood demands (Yousefpour et al., 2018).

Additionally, non-strictly protected areas under 'closer to nature' forest management (20%), can provide up to an extra 6–7% of EU wood harvest, compared to strictly protecting entirely 30% of the area. Therefore, wood harvest from non-strictly protected areas could also substantially contribute to mitigate impacts of the protection policy.

Our work can provide quantitative basis for calculating the compensation for economic loss of harvest income through a payment for ecosystem services (Wunder, 2015), assuming the generation of at least biodiversity and climate benefits, alongside wood production as a co-produced provisioning service (Kangas and Ollikainen, 2022). This

analysis could also help targeting the marketing of voluntary payments (Forsius et al., 2021). Yet, it is important to keep in mind that the national policies and institutional framework in each MS will vary (Primmer et al., 2021).

Overall, this study provides a negotiation basis for demonstrating that it would be possible to achieve common EU policy conservation targets without disproportionate impacts on the forest sector demands inside and outside EU by applying different approaches to allocate protected areas.

# CRediT authorship contribution statement

Fulvio Di Fulvio: Writing - review & editing, Writing - original draft, Methodology, Investigation, Formal analysis, Data curation, Conceptualization. Tord Snäll: Writing - review & editing, Writing original draft, Conceptualization, Pekka Lauri: Writing - review & editing, Writing - original draft, Methodology, Formal analysis, Data curation, Conceptualization. Nicklas Forsell: Writing - review & editing, Writing - original draft, Conceptualization. Mikko Mönkkönen: Writing - review & editing, Funding acquisition, Conceptualization. Daniel Burgas: Writing - review & editing, Conceptualization. Clemens Blattert: Writing - review & editing, Conceptualization. Kyle Eyvindson: Writing - review & editing, Conceptualization. Astor Toraño Caicoya: Writing - review & editing. Marta Vergarechea: Writing - review & editing. Clara Antón-Fernández: Writing - review & editing. Julian Klein: Writing - review & editing. Rasmus Astrup: Writing - review & editing. Jani Lukkarinen: Writing - review & editing. Samuli Pitzén: Writing - review & editing. Eeva Primmer: Writing - review & editing.

#### Declaration of competing interest

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests: Fulvio Di Fulvio reports financial support was provided by ERA-Net Cofund scheme. Fulvio Di Fulvio reports a relationship with International Institute for Applied Systems Analysis that includes: employment. If there are other authors, they declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

# Acknowledgments

This work was conducted within the project MultiForest, which is supported under the umbrella of ERA-NET Cofund ForestValue by: Academy of Finland (326321), Business Finland, Federal Ministry of Agriculture, Forestry, Environment & Water Management (Austria), Agency for Renewable Resources (Germany), Research Council of Norway, Vinnova (2018-04982; Sweden). ForestValue has received funding from the European Union Horizon 2020 research and innovation program under grant agreement N° 773324. Additionally, we acknowledge from the EU Horizon Europe support project Forest-Navigator-Navigating European forests and forest bioeconomy sustainably to EU climate neutrality (GA No. 101056875). We are thankful to Dr. Nicolas Robert for the critical reading of this paper.

# Appendix A. Supplementary data

Supplementary data to this article can be found online at https://doi.org/10.1016/j.gloenvcha.2025.102986.

# Data availability

Data will be made available on request.

#### References

- Aggestam, F., Giurca, A., 2021. The art of the "green" deal: Policy pathways for the EU Forest Strategy. Forest Policy Econ. 128, 102456.
- Besnard, S., Koirala, S., Santoro, M., Weber, U., Nelson, J., Gütter, J., Herault, B., Kassi, J., N'Guessan, A., Neigh, C., Poulter, B., Zhang, T., Carvalhais, N., 2021. Mapping global forest age from forest inventories, biomass and climate data. Earth Syst. Sci. Data 13 (10), 4881–4896.
- Betts, M., Phalan, B., Wolf, C., et al., 2021. Producing wood at least cost to biodiversity: integrated triad and sharing-sparing approaches to inform forest landscape management. Biol. Rev. 96, 1301–1317.
- Blattert, C., Eyvindson, K., Hartikainen, M., Burgas, D., Potterf, M., Lukkarinen, J., Mönkkönen, M., 2022. Sectoral policies cause incoherence in forest management and ecosystem service provisioning. Forest Policy and Economics, 136, art. no. 102689, doi: 10.1016/j.forpol.2022.102689.
- Blattert, C., Mönkkönen, M., Burgas, D., Di Fulvio, F., Toraño Caicoya, A., Vergarechea, M., Klein, J., Hartikainen, M., et al., 2023. Climate targets in European timber-producing countries conflict with goals on forest ecosystem services and biodiversity. Commun. Earth Environ. 4 (1), e119. https://doi.org/10.1038/s43247-023-00771-z.
- Cerullo G, Barlow J, Betts M, Edwards D, Eyres A, França F, Garrett R, Swinfield T, Tew E, White T, Balmford A., 2023. The global impact of EU forest protection policies. Science. 2023 Aug 18;381(6659): 740. doi: 10.1126/science.adj0728.
- Di Fulvio, F., Forsell, N., Lindroos, O., 2016. Spatially explicit assessment of roundwood and logging residues availability and costs for the EU28. Scand. J. for. Res. 31 (7), 691–707.
- EEA 2020a. Annual European Union greenhouse gas inventory 1990–2018 and inventory report 2020. Available online at: https://www.eea.europa.eu/publications/european-union-greenhouse-gas-inventory-2020.
- EEA 2020 b. State of nature in the EU. Results from reporting under the nature directives 2013-2018. EEA Report No 10/2020. Luxembourg: Publications Office of the European Union, 2020.

Eggers, J., Lundström, J., Snäll, T., Öhman, K., 2022. Balancing wood production and biodiversity in intensively managed boreal forest. Scand. J. For. Res. 37 (3), 213–225.

- European Commission 2013. Communication from the Commission to the European Parliament, the Council, the Economic and Social Committee and the Committee of the Regions. Green Infrastructure (GI)—Enhancing Europe's Natural Capital; COM (2013) 249 final; European Commission: Brussels, Belgium, 2013. Available online: https://eur-lex.europa.eu/resource.html?uri=cellar:d41348f2-01d5-4abeb817-4c73e6f1b2df.0014.03/DOC\_1&format=PDF.
- European Commission 2019a. Guidance on a strategic framework for further supporting the deployment of EU-level green and blue infrastructure. COMMISSION STAFF WORKING DOCUMENT. Brussels, 24.5.2019 SWD (2019) 193 final. https://ec. europa.eu/environment/nature/ecosystems/pdf/SWD\_2019\_193\_F1\_STAFF\_ WORKING\_PAPER\_EN\_V4\_P1\_1024680.PDF.
- European Commission 2019b. Communication from the Commission to the European Parliament, the European Council, the Council, the European Economic and Social Committee and the Committee of the Regions- The European Green Deal. European

Commission. https://eur-lex.europa.eu/resource.html?uri=cellar:b828d165-1c22-11ea8c1f-01aa75ed71a1.0002.02/DOC 1&format=PDF.

- European Commission 2020a. EU Biodiversity Strategy for 2030. Bringing nature back into our lives. Communication from the Commission to the European parliament, the council, the European economic and social committee and the committee of the regions. Brussels, 20.5.2020 COM(2020) 380 final. https://eur-lex.europa.eu/ resource.html?uri=cellar:a3c806a6-9ab3-11ea-9d2d-01aa75ed71a1.0001.02/DOC\_ 1&format=PDF.
- European Commission 2020b, Note on criteria and guidance for protected areas design. Commission Staff Working Document. N 23. Brussels END. D.3/JC. https:// environment.ec.europa.eu/system/files/2022-01/SWD\_guidance\_protected\_areas. pdf.
- European Commission 2021a. 'Fit for 55': delivering the EU's 2030 Climate Target on the way to climate neutrality. Communication from the commission to the European parliament, the council, the European economic and social committee and the committee of the regions. Brussels, 14.7.2021, COM(2021) 550 final. https://eur-lex. europa.eu/legal-content/EN/TXT/PDF/?uri=CELEX:52021DC0550&from=EN.
- European Commission 2021b. Communication from the Commission to the European Parliament, the Council, the European Economic and Social Committee and the Committee for the Regions. New EU Forest Strategy for 2030. COM(2021) 572 final., Brussels. https://eurlex.europa.eu/legal-content/EN/TXT/PDF/?uri=CELEX: 52021DC0572.
- European Commission 2022. Criteria and guidance for protected areas designations. Commission staff working document. Brussels, 28.1.2022. SWD (2022) 23 final. https://environment.ec.europa.eu/system/files/2022-01/SWD\_guidance\_protected\_ areas.pdf.
- European Commission 2023a, Guidelines on Closer-to-Nature Forest Management. Commission staff working document. Brussels, 27.7.2023 SWD (2023) 284 final https://environment.ec.europa.eu/system/files/2023-07/SWD\_2023\_284\_F1\_ STAFF\_WORKING\_PAPER\_EN\_V2\_P1\_2864149.PDF.
- European Commission 2023b. Commission Guidelines for Defining, Mapping, Monitoring and Strictly Protecting EU Primary and Old-Growth Forests. Commission staff working document. Brussels, 20.3.2023 SWD (2023) 62 final. https://ec. europa.eu/transparency/documents-register/detail?ref=SWD(2023)62&lang=en.
- FAO, 2023. FAOSTAT Database. Available at: https://www.fao.org/faostat.
- Forsius, M., Kujala, H., Minunno, F., Holmberg, M., Leikola, N., Mikkonen, N.,Autio, I., Paunu, V.-V., Tanhuanpää, T., Hurskainen, P., Mäyrä, J., Kivinen, S., Keski-Saari, S., Kosenius, A.-K., Kuusela, S., Virkkala, R., Viinikka, A., Vihervaara, P., Akujärvi, A., Bäck, J., Karvosenoja, N., Kumpula, T., Kuzmin, A., Mäkelä, A., Moilanen, A., Ollikainen, M., Pekkonen, M., Peltoniemi, M., Poikolainen, L., Rankinen, K., Rasilo, T., Tuominen, S., Valkama, J., Vanhala, P., Heikkinen, R.K., 2021. Developing a spatially explicit modelling and evaluation framework for integrated carbon sequestration and biodiversity conservation: Application in southern Finland. Science of the Total Environment, 775, art. no. 145847.
- FRA, 2020. Food and Agriculture Organization of the UN (FAO), "Global Forest Resources Assessment 2020", FAO, Rome. Available at: http://www.fao.org/ documents/card/en/c/ca9825en.
- Gaulier, G., Zignago, S., 2010. BACI: International Trade Database at the Product Level, CEPII Working Paper 2010-23.
- GBF, 2022. Kunming-Montreal Global Biodiversity Framework. CBD/COP/DEC/15/4 19 December 2022.
- Griscom, B.W., Adams, J., Ellis, P.W., Houghton, R.A., Lomax, G., Miteva, D.A., Schlesinger, W.H., Shoch, D., Siikamäki, J.V., Smith, P., Woodbury, P., Zganjar, C., Blackman, A., Campari, J., Conant, R.T., Delgado, C., Elias, P., Gopalakrishna, T., Hamsik, M.R., Herrero, M., Kiesecker, J., Landis, E., Laestadius, L., Leavitt, S.M., Minnemeyer, S., Polasky, S., Potapov, P., Putz, F.E., Sanderman, J., Silvius, M., Wollenberg, E., Fargione, J., 2017. Natural climate solutions. PNAS 114 (44), 11645–11650. https://doi.org/10.1073/pnas.1710465114.
- Gusti, M., Kindermann, G., 2011. An approach to modeling land-use change and forest management on a global scale. In: Kacprzyk, J., Pina, N., Filipe, J. (Eds.), Proceedings of 1st International Conference on Simulation and Modeling Methodologies, Technologies and Applications.
- Havlík, P., Schneider, U.A., Schmid, E., Böttcher, H., Fritz, S., Skalský, R., Aoki, K., Cara, S.D., Kindermann, G., Kraxner, F., Leduc, S., McCallum, I., Mosnier, A., Sauer, T., Obersteiner, M., 2011. Global land-use implications of first and second generation biofuel targets. Energy Policy 39 (10), 5690–5702.
- Havlík, P., Valin, H., Herrero, M., Obersteiner, M., Schmid, E., Rufino, M.C., Mosnier, A., Thornton, P.K., Böttcher, H., Conant, R.T., Frank, S., Fritz, S., Fuss, S., Kraxner, F., Notenbaert, A., 2014. Climate change mitigation through livestock system transitions. PNAS 111 (10), 3709–3714.
- Hermoso, V., Morán-Ordóñez, A., Lanzas, M., Brotons, L., 2020. Designing a network of green infrastructure for the EU. Landscape and Urban Planning, 196 art. no. 103732. doi:10.1016/j.landurbplan.2019.103732.
- Hill, S.L.L., Arnell,A., Maney, C., Butchart, S.H.M., Hilton-Taylor, C., Ciciarelli, C., Davis, C., Dinerstein, E., Purvis,A., Burgess, N.D., 2019. Measuring Forest Biodiversity Status and Changes Globally. Frontiers in Forests and Global Change, 2, art. no. 70,.
- Hurmekoski, E., Kunttu, J., Heinonen, T., Pukkala, T., & Peltola, H. 2023. Does expanding wood use in construction and textile markets contribute to climate change mitigation? Renewable and Sustainable Energy Reviews, 174, art. no. 113152. doi: 10.1016/j.rser.2023.113152.
- IIASA, 2020. SSP Database. https://tntcat.iiasa.ac.at/SspDb.
- IPBES, 2019. Global assessment report on biodiversity and ecosystem services of the Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services. E. S. Brondizio, J. Settele, S. Díaz, and H. T. Ngo (editors). IPBES secretariat, Bonn, Germany. 1148 pages. https://doi.org/10.5281/zenodo.3831673.

- Jung, M., Arnell, A., de Lamo, X., García-Rangel, S., Lewis, M., Mark, J., Merow, C., Miles, L., Ondo, I., Pironon, S., Ravilious, C., Rivers, M., Schepaschenko, D., Tallowin, O., van Soesbergen, A., Govaerts, R., Boyle, B.L., Enquist, B.J., Feng, X., Gallagher, R., Maitner, B., Meiri, S., Mulligan, M., Ofer, G., Roll, U., Hanson, J.O., Jetz, W., Di Marco, M., McGowan, J., Rinnan, D.S., Sachs, J.D., Lesiv, M., Adams, V. M., Andrew, S.C., Burger, J.R., Hannah, L., Marquet, P.A., McCarthy, J.K., Morueta-Holme, N., Newman, E.A., Park, D.S., Roehrdanz, P.R., Svenning, J.-C., Violle, C., Wieringa, J.J., Wynne, G., Fritz, S., Strassburg, B.B.N., Obersteiner, M., Kapos, V., Burgess, N., Schmidt-Traub, G., Visconti, P., 2021. Areas of global importance for conserving terrestrial biodiversity, carbon and water. Nat. Ecol. Evol. 5 (11), 1499–1509.
- Kallio, A.M.I., Solberg, B., Käär, L., Päivinen, R., 2018. Economic impacts of setting reference levels for the forest carbon sinks in the EU on the European forest sector. Forest Policy Econ. 92, 193–201. https://doi.org/10.1016/j.forpol.2018.04.010.
- Kangas, J., Ollikainen, M., 2022. A PES scheme promoting forest biodiversity and carbon sequestration Forest Policy and Economics, 136, art. no. 102692.
- Kindermann, G., McCallum, I., Fritz, S., Obersteiner, M., 2008. A global forest growing stock, biomass and carbon map based on FAO statistics. Silva Fenn. 42, 387–396.
  Kindermann, G., Obersteiner, M., Rametsteiner, E., McCallum, I., 2006. Predicting the deforestation-trend under different carbon-prices. Carbon Bal. Manage. 1, 1–17.
- Larsen, J.B., Angelstam, P., Bauhus, J., Carvalho, J.F., Diaci, J., Dobrowolska, D., Gazda, A., Gustafsson, L., Krumm, F., Knoke, T., Konczal, A., Kuuluvainen, T., Mason, B., Motta, R., Pötzelsberger, E., Rigling, A., Schuck, A., 2022. Closer-to Nature Forest Management. From Science to Policy 12. European Forest Institute. https://doi.org/ 10.36333/fs12.
- Latta, G., Sjolie, H., Solberg, B., 2013. A review of recent developments and applications of partial equilibrium models of the forest sector. J. For. Econ. 19, 350–360.
- Lauri, P., Forsell, N., Di Fulvio, F., Snäll, T., Havlik, P., 2021. Material substitution between coniferous, non-coniferous and recycled biomass – Impacts on forest industry raw material use and regional competitiveness. Forest Policy and Economics, 132, art. no. 102588.
- Lauri, P., Forsell, N., Gusti, M., Korosuo, A., Havlík, P., Obersteiner, M., 2019. Global woody biomass harvest volumes and forest area use under different SSP-RCP scenarios. J. For. Econ. 34 (3–4), 285–309.
- Leclère, D., Obersteiner, M., Barrett, M., Butchart, S.H.M., Chaudhary, A., De Palma, A., DeClerck, F.A.J., Di Marco, M., Doelman, J.C., Dürauer, M., Freeman, R., Harfoot, M., Hasegawa, T., Hellweg, S., Hilbers, J.P., Hill, S.L.L., Humpenöder, F., Jennings, N., Krisztin, T., Mace, G.M., Ohashi, H., Popp, A., Purvis, A., Schipper, A. M., Tabeau, A., Valin, H., van Meijl, H., van Zeist, W.-J., Visconti, P., Alkemade, R., Almond, R., Bunting, G., Burgess, N.D., Cornell, S.E., Di Fulvio, F., Ferrier, S., Fritz, S., Fujimori, S., Grooten, M., Harwood, T., Havlík, P., Herrero, M., Hoskins, A. J., Jung, M., Kram, T., Lotze-Campen, H., Matsui, T., Meyer, C., Nel, D., Newbold, T., Schmidt-Traub, G., Stehfest, E., Strassburg, B.B.N., van Vuuren, D.P., Ware, C., Watson, J.E.M., Wu, W., Young, L., 2020. Bending the curve of terrestrial biodiversity needs an integrated strategy. Nature 585 (7826), 551–556.
- Lesiv, M., Schepaschenko, D., Buchhorn, M., See, L., Dürauer, M., Georgieva, I., Jung, M., Hofhansl, F., Schulze, K., Bilous, A., Blyshchyk, V., Mukhortova, L., Brenes, C.L.M., Krivobokov, L., Ntie, S., Tsogt, K., Pietsch, S.A., Tikhonova, E., Kim, M., Di Fulvio, F., Su, Y.-F., Zadorozhniuk, R., Sirbu, F.S., Panging, K., Bilous, S., Kovalevskii, S.B., Kraxner, F., Rabia, A.H., Vasylyshyn, R., Ahmed, R., Diachuk, P., Kovalevskii, S.S., Bungnamei, K., Bordoloi, K., Churilov, A., Vasylyshyn, O., Sahariah, D., Tertyshnyi, A.P., Saikia, A., Malek, Ž., Singha, K., Feshchenko, R., Prestele, R., Akhtar, I.H., Sharma, K., Domashovets, G., Spawn-Lee, S.A., Blyshchyk, O., Slyva, O., Ilkiv, M., Melnyk, O., Sliusarchuk, V., Karpuk, A., Terentiev, A., Bilous, V., Blyshchyk, K., Bilous, M., Bogovyk, N., Blyshchyk, I., Bartalev, S., Yatskov, M., Smets, B., Visconti, P., Mccallum, I., Obersteiner, M., Fritz, S. 2022. Global forest management data for 2015 at a 100 m resolution. Scientific Data, 9 (1), art. no. 199.
- Lier, M., Köhl, M., Korhonen, K.T., Linser, S., Prins, K., 2021. Forest relevant targets in EU policy instruments - can progress be measured by the pan-European criteria and indicators for sustainable forest management? Forest Policy and Economics, 128, art. no. 102481.
- Lotze-Campen, H., Verburg, P.H., Popp, A., Lindner, M., Verkerk, P.J., Moiseyev, A., Schrammeijer, E., Helming, J., Tabeau, A., Schulp, C.J.E., van der Zanden, E.H., Lavalle, C., e Silva, F.B., Walz, A., Bodirsky, B., 2018. A cross-scale impact assessment of European nature protection policies under contrasting future socioeconomic pathways. Regional Environmental Change, 18 (3) pp. 751-762.

- Mazziotta, A., Lundström, J., Forsell, N., Moor, H., Eggers, J., Subramanian, N., Aquilué, N., Morán-Ordóñez, A., Brotons, L., Snäll, T., 2022. More future synergies and less trade-offs between forest ecosystem services with natural climate solutions instead of bioeconomy solutions. Glob. Chang. Biol. 28 (21), 6333–6348.
- Mishra, A., Humpenöder, F., Churkina, G., Reyer, C.P.O., Beier, F., Bodirsky, B.L., Schellnhuber, H.J., Lotze-Campen, H., Popp, A., 2022. Land use change and carbon emissions of a transformation to timber cities. Nature Communications, 13 (1), art. no. 4889,.
- Pilli, R., Alkama, R., Cescatti, A., Kurz, W.A., Grassi, G., 2022. The european forest carbon budget under future climate conditions and current management practices. Biogeosciences 19 (13), 3263–3284. https://doi.org/10.5194/bg-19-3263-2022.
- Pilli, R., Grassi, G., Kurz, W.A., Fiorese, G., Cescatti, A., 2017. The european forest sector: past and future carbon budget and fluxes under different management scenarios. Biogeosciences 14 (9), 2387–2405. https://doi.org/10.5194/bg-14-2387-2017.
- Pitzén, S., Lukkarinen, J., Primmer, E., 2023. Coherent at face value: Integration of forest carbon targets in Finnish policy strategies. Ambio 52 (11), 1861–1877. https://doi. org/10.1038/s43247-023-00771-.
- Primmer, E., Varumo, L., Krause, T., Orsi, F., Geneletti, D., Brogaard, S., Aukes, E., Ciolli, M., Grossmann, C., Hernández Morcillo, M., Kister, J., Kluvánková, T., Loft, L., Maier, C., Meyer, C., Schleyer, C., Spacek, M., Mann, C., 2021. Mapping Europe's institutional landscape for forest ecosystem service provision, innovations and governance. Ecosystem Services, 47, art. no. 101225.
- Räty, M., Juutinen, A., Korhonen, K.T., Syrjänen, K., Kärkkäinen, L., 2023. EU wood production vs. biodiversity goals–possible reconciliation in Finland? Scand. J. For. Res. 38 (5), 287–299.
- Riahi, K., van Vuuren, D.P., Kriegler, E., Edmonds, J., O'Neill, B.C., Fujimori, S., Bauer, N., Calvin, K., Dellink, R., Fricko, O., Lutz, W., Popp, A., Cuaresma, J.C., KC, S., Leimbach, M., Jiang, L., Kram, T., Rao, S., Emmerling, J., Ebi, K., Hasegawa, T., Havlik, P., Humpenöder, F., Da Silva, L.A., Smith, S., Stehfest, E., Bosetti, V., Eom, J., Gernaat, D., Masui, T., Rogelj, J., Streffer, J., Drouet, L., Krey, V., Luderer, G., Harmsen, M., Takahashi, K., Baumstark, L., Doelman, J.C., Kainuma, M., Klimont, Z., Marangoni, G., Lotze-Campen, H., Obersteiner, M., Tabeau, A., Tavoni, M., 2017. The Shared Socioeconomic Pathways and their energy, land use, and greenhouse gas emissions implications: An overview. Global Environmental Change, 42, pp. 153-168.
- Riviere, M., Caurla, S., 2020. Representations of the forest sector in economic models. Oeconomia 10–3. https://doi.org/10.4000/oeconomia.9418.
- Rosa, F., Di Fulvio, F., Lauri, P., Felton, A., Forsell, N., Pfister, S., Hellweg, S., 2023. Can forest management practices counteract species loss arising from increasing european demand for forest biomass under climate mitigation scenarios? Environ. Sci. Tech. 57 (5), 2149–2161. https://doi.org/10.1021/acs.est.2c07867.
- Sabatini, F.M., Keeton, W.S., Lindner, M., Svoboda, M., Verkerk, P.J., Bauhus, J., Bruelheide, H., Burrascano, S., Debaive, N., Duarte, I., Garbarino, M., Grigoriadis, N., Lombardi, F., Mikoláš, M., Meyer, P., Motta, R., Mozgeris, G., Nunes, L., Ódor, P., Panayotov, M., Ruete, A., Simovski, B., Stillhard, J., Svensson, J., Szwagrzyk, J., Tikkanen, O.-P., Vandekerkhove, K., Volosyanchuk, R., Vrska, T., Zlatanov, T., Kuemmerle, T., 2020. Protection gaps and restoration opportunities for primary forests in Europe. Divers. Distrib. 26 (12), 1646–1662.
- Schier, F., Iost, S., Seintsch, B., Weimar, H., Dieter, M., 2022. Assessment of possible production leakage from implementing the EU biodiversity strategy on forest product markets. Forests 2022 (13), 1225. https://doi.org/10.3390/f13081225.
- Toppinen, A., Kuuluvainen, J., 2010. Forest sector modelling in the Europe—the state of the art and future research directions. Forest Policy Econ. 12, 2–8.
- WDPA, 2020. World Database on Protected Areas. https://www.iucn.org/theme/ protected-areas/our-work/quality-and-effectiveness/world-database-protectedareaswdpa.
- Wolfslehner B., Linser S., Pülzl H., Bastrup Birk A., Camia A., Marchetti M., 2016. Forest bioeconomy - a new scope for sustainability indicators. European Forest Institute, 2016. JRC105224. http://www.efi.int/files/attachments/publications/efi\_fstp\_4\_ 2016.pdf.
- Wunder, S., 2015. Revisiting the concept of payments for environmental services. Ecol. Econ. 117, 234–243.
- Yousefpour, R., Augustynczik, A.L.D., Reyer, C.P., Lasch-Born, P., Suckow, F., Hanewinkel, M., 2018. Realizing mitigation efficiency of European commercial forests by climate smart forestry. Sci. Rep. 8 (1), 1–11.